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Air Pollutant Emissions 1990-2018



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This report documents the methodologies used in the Norwegian emission inventory of acidifying pollutants, particulate matters, heavy metals and persistent organic pollutants submitted under the UNECE Convention on Long-range Transboundary Air Pollution.

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Executive Summary

This report documents the methodologies used in the Norwegian emission inventory of acidifying pollutants, particulate matters, heavy metals and persistent organic pollutants.

The Norwegian emission inventory is a joint undertaking between the Norwegian Environment Agency¹ and Statistics Norway. This report has been prepared by the Norwegian Environment Agency in close collaboration with Statistics Norway's Division for Energy, Environment and Transport.

The most important changes since the 2019 submission are:

- Emission of NH₃ and NO_x from animal manure has been revised as a result of updated activity data from Statistics Norway 2018 survey on manure management systems.
- There are recalculations for most components in the year 1997 due to an error in the energy balance in the previous submission. Activity data from the statistics "Energy use in the manufacturing sector" for the year 1998 was wrongly used in both 1997 and 1998.
- The emission factors for particles for 2A1 Cement production, 2A2 Lime production, 2A5a Quarrying and mining of minerals other than coal, 2C3 Primary aluminium production, 2C3 Secondary aluminium production and 2H1 Pulp and paper industry has been updated according to EEA (2019) guidebook.

Chapter 8.1 Recalculations gives a more thorough description of changes in the most recent emission calculations.

¹ Former names are "Climate and Pollution Agency" and "Norwegian Pollution Control Authority".

1 Introduction

1.1 National Inventory Background

The Norwegian national inventory for long-range transboundary air pollutants includes emission data for the years 1990-2018. The emissions covered in this report are those embraced by the convention on long-range transboundary air pollution, i.e. they are defined with a territorial delimitation. The calculation methods used and the documentation of these, are, as far as possible, in accordance with the strict demands formulated in the emission convention.

1.2 Institutional arrangements

The Norwegian emissions inventories have been produced for about three decades as collaboration between Statistics Norway (SSB) and the Norwegian Environment Agency (NEA).

Statistics Norway is responsible for the official statistics on emissions to air, and contributes to the reporting to the UNECE. Their tasks include:

- collection of activity data
- operation and further development of models for emission estimation
- emission calculations
- filling in most of the tables for international reporting to UNECE
- publishing national official statistics on emissions to air

The Norwegian Environment Agency is responsible for:

- international reporting to the UNECE
- emission factors for all emission sources
- measured emission data from large industrial plants based on individual reports submitted to the Norwegian Environment Agency on a regular basis
- considering the quality and assuring necessary updating of emission models like, e.g. the road traffic model

Activity data² are collected either internally at Statistics Norway (e.g. data on energy use, industrial production, number of animals, etc.) or reported to Statistics Norway, and in some cases to the Norwegian Environment Agency, from external sources such as the Norwegian Petroleum Directorate (OD) and the Norwegian Public Roads Administration (VD). Emission figures are derived from models operated by Statistics Norway. In the modelling activities Statistics Norway makes use of the data collected by the Norwegian Environment Agency on emission factors and emissions from industrial plants.

² Data on the magnitude of human activity resulting in emissions or removals taking place during a given period of time.

The Norwegian Environment Agency is responsible for quality control of the data they deliver to the emission model operated by Statistics Norway, but Statistics Norway makes an additional consistency check (see chapter 1.6). Statistics Norway is responsible for quality control of the activity data and the emission figures from the model, and the Norwegian Environment Agency also participates in this quality control before reporting to the UNECE.

1.3 Inventory preparation process

The Norwegian emission inventory is based on a general emission model and a series of more detailed supplementary models, which cover specific emission sources and pollutants (e.g. road traffic, air traffic, solvents). These smaller models feed results into the general model. All models are operated by Statistics Norway.

Data and information on point sources are recorded at the Norwegian Environment Agency in the database *Forurensning* and published in *Norske utslipp* (<http://www.norskeutslipp.no>). This is the Norwegian Pollutant Release and Transfer Register (PRTR). *Forurensning* is a further development of the old register Inkosys, which was introduced in 1978 as an internal tool for the authorities. The database was upgraded in 1992, and has later been under continuous development in order to harmonise with the PRTR adopted by the OECD in 1996. Each polluting industrial installation or plant is subjected to licensing and is obliged to produce an annual report to the pollution control authorities. The report should provide activity data, emission figures and information about the particular source, and it should address compliance with current environmental standards. The Norwegian Environment Agency supplies Statistics Norway with data from the Norwegian PRTR which are relevant for the preparation of the national emission inventory.

1.3.1 Pollutants included, data collection, processing and archiving

Statistics Norway collects the majority of data necessary to run the Norwegian emission model. These are as follows: activity levels, emission factors, aggregated results from the smaller, supplementary models and emission figures for point sources. Table 1.1 gives an overview of pollutants included in the emission inventory which are restricted by CLRTAP.

Table 1.1. Definition of pollutants in the Norwegian emission inventory which are restricted by CLRTAP

| Class | Pollutant | Symbol | Definition |
|--------------------------------------|--|-------------------|---|
| Acidifying gases | Sulphur dioxide | SO ₂ | NO + NO ₂ |
| | Nitrogen oxides | NO _x | |
| | Ammonia | NH ₃ | |
| Heavy metals (HM) | Lead | Pb | |
| | Cadmium | Cd | |
| | Mercury | Hg | |
| | Arsenic | As | |
| | Chromium | Cr | |
| | Copper | Cu | |
| Persistent organic pollutants (POPs) | Polycyclic Aromatic Hydrocarbons | PAH-4 | Emissions are calculated for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene. |
| | Dioxins | - | Dioxin emissions are given in the unit I-TEQ, which is required for reporting to CLRTAP. I-TEQ is based on the international model ("Nato-modell") and is the sum of PCDD/PCDF multiplied by the components toxicity equivalency factor (I-TEF). TEQ = sum (PCDD _i * TEF _i) + sum (PCDD _j * TEF _j). |
| | Hexachlorobenzene | HCB | |
| | Polychlorinated biphenyl | PCB | |
| | | | |
| Particulates | Total suspended particulates | TSP | Particulate matter with diameter less than 10µm Particulate matter with diameter less than 2.5µm |
| | - | PM ₁₀ | |
| | - | PM _{2.5} | |
| | Black carbon | BC | |
| Other pollutants | Carbon monoxide | CO | |
| | Non-methane volatile organic compounds | NMVOC | |

Source: Statistics Norway/Norwegian Environment Agency

The collected data are subjected to the Quality Assurance and Quality Control (QA/QC) routines described in section 1.6 as well as source specific routines as described under each source chapter. They are subsequently processed by Statistics Norway into a format appropriate to enter the emission models. The models are designed in a manner that accommodates both the estimation methodologies reflecting Norwegian conditions and those recommended internationally.

Input data used and the model output are all stored at Statistics Norway. Relevant information including dates and procedures followed are also recorded.

1.3.2 Definitions and structure

The structure of this documentation follows the nomenclature used for reporting to the Convention on Long-Range Transboundary Air Pollution (CLRTAP) as Nomenclature For Reporting (NFR14).

The main sectors here are:

- NFR 1. Energy
- NFR 2. Industrial processes and product use
- NFR 3. Agriculture
- NFR 6. Waste

The description of the pollutants included is given in Table 1.1.

Emissions of heavy metals, POPs and particulates are further described in the reports Finstad et al. (2001), Finstad et al. (2002b), Finstad and Rypdal (2003) and Finstad et al. (2003). Emissions of black carbon is described in Aasestad (2013), emissions of HCB in Aasestad (2014) and emissions of PCB in Aasestad (2016).

1.3.3 Archiving

The national emission inventory is a part of Statistics Norway's data archiving system. All input data to, and results from, the general Norwegian emission model from every publication cycle are stored and documented in this system.

Several input data are used in preliminary calculations before entering into the general Norwegian emission model. This includes supplementary models such as road traffic and air traffic, as well as a number of simpler calculations that do not fit into the framework of the general model. The preliminary calculations are not included in the central archiving system, which is not suited for such a diverse collection of data. For some supplementary models there is an established archiving routine where all input data and results from every calculation cycle are stored.

1.4 Methods and data sources

This chapter describes the general structure of the Norwegian emission model. The model was developed by Statistics Norway (Daasvatn et al. 1992), (Daasvatn et al. 1994). It was redesigned in 2003 in order to improve reporting to the UNFCCC and UNECE, and to improve QA/QC procedures.

The Norwegian emission model is organised around a general emission model called “Kuben” (“the Cube”). Several emission sources, e.g. road traffic, air traffic and solvents are covered by more detailed supplementary models. Aggregated results from the supplementary models are used as input to the general model. The supplementary models are presented in the appropriate sections of chapters 3-8. This chapter describes the general emission model.

1.4.1 Structure of the general emission model

The general emission model is based on equation (1.1).

$$(1.1) \quad \text{Emissions (E)} = \text{Activity level (A)} \cdot \text{Emission Factor (EF)}$$

For emissions from *combustion*, the activity data concern energy use. In the Norwegian energy balance, the use of different forms of energy is allocated to industries (economic sectors). In order to calculate emissions to air, energy use must also be allocated to technical sources (e.g. equipment).

The energy use data are combined with a corresponding matrix of emission factors. In principle, there should be one emission factor for each combination of fuel, industry, source and pollutant. However, in a matrix with a cell for each combination, most of the cells would be empty (no consumption). In addition, the same emission factor would apply to many cells.

Emissions of some pollutants from major manufacturing plants (point sources) are available from measurements or other plant-specific calculations. When such measured data are available it is possible to replace the estimated values by the measured ones:

$$(1.2) \quad \text{Emissions (E)} = [(A - A_{PS}) \cdot EF] + E_{PS}$$

where A_{PS} and E_{PS} are the activity and the measured emissions at the point sources, respectively. Emissions from activity for which no point source estimate is available ($A - A_{PS}$) are still estimated with the regular emission factor.

Non-combustion emissions are generally calculated in the same way, by combining appropriate activity data with emission factors. Some emissions may be obtained from current reports and investigations, and some are measured directly as described in chapters 3-8. The emissions are fitted into the general model using the parameters industry, source, and pollutant. The fuel parameter is not relevant here. The source sector categories are based on EMEP/NFR categories, with further subdivisions where more detailed methods are available.

1.4.2 Pollutants, industries, fuels, and sources

The *pollutants* currently included in the model are listed in Table 1.1. The model uses approximately 130 *industries* (economic sectors). The classification is common with the basis data in the energy balance/accounts, and is almost identical to that used in the national accounts, which is aggregated from the European NACE (rev. 2) classification. The large number of sectors is an advantage in dealing with important emissions from manufacturing industries. The disadvantage is an unnecessary disaggregation of sectors with very small emissions. To make the standard sectors more appropriate for calculation of emissions, a few changes have been made, e.g. "Private households" is defined as a sector.

The *fuels* and technical *sources* used for combustion with energy use (NFR source sector 1A) are shown in Table 1.2, Table 1.3 and Table 1.4.

Table 1.2. Energy commodities in the Norwegian emission inventory

| Energy commodity | Aggregate fuel category in NFR |
|---------------------------|--------------------------------|
| Coal | Solid Fuels |
| Coke | Solid Fuels |
| Petrol coke | Liquid Fuels |
| Wood | Biomass |
| Wood waste | Biomass |
| Black liquor | Biomass |
| Wood pellets | Biomass |
| Wood briquettes | Biomass |
| Charcoal | Biomass |
| Natural gas | Gaseous Fuels |
| Refinery gas | Liquid Fuels |
| CO gas | Solid Fuels |
| Landfill gas | Biomass |
| Biogas | Biomass |
| Fuel gas | Liquid Fuels |
| LPG | Liquid Fuels |
| Gasoline (road transport) | Liquid Fuels |
| Aviation gasoline | Liquid Fuels |
| Kerosene (heating) | Liquid Fuels |
| Jet kerosene | Liquid Fuels |
| Autodiesel | Liquid Fuels |
| Marine gas oil | Liquid Fuels |
| Light fuel oils | Liquid Fuels |
| Heavy distillate | Liquid Fuels |
| Heavy fuel oil | Liquid Fuels |
| Municipal waste | Other Fuels |
| Special waste* | Other Fuels |

* Special waste was moved from *Liquid* to *Other* fuels in 2014.

Source: Statistics Norway/Norwegian Environment Agency

Table 1.3. Sources for energy combustion in the Norwegian emission inventory

| Source | NFR |
|----------------------------------|-----------------------|
| Stationary combustion | |
| Direct fired furnaces | 1A1, 1A2 |
| Gas turbines | 1A1c, 1A4a |
| Boilers | 1A1, 1A2, 1A4, 1A5 |
| Small stoves | 1A2, 1A4, 1A5 |
| Flaring | 1B2C, 6C |
| Mobile combustion* | |
| Passenger car | 1A3b i, 1A5b |
| Light duty vehicles | 1A3b ii, 1A5b |
| Heavy duty vehicles | 1A3b iii, 1A5b |
| Motorcycle | 1A3b iv |
| Moped | 1A3b iv |
| Snowscooter | 1A4b, c |
| Railway | 1A3c |
| Aviation jet/turboprop (0-1000m) | 1A3a ii (i), 1A5b |
| Aviation jet/turboprop (cruise) | 1A3a ii (ii), 1A5b |
| Aviation helicopter (0-1000m) | 1A3a ii (i) |
| Aviation helicopter (cruise) | 1A3a ii (ii) |
| Aviation small craft (0-1000m) | 1A3a ii (i) |
| Aviation small craft (cruise) | 1A3a ii (ii) |
| Ships | 1A3d, 1A4c, 1A5b |
| Small boats 2 stroke | 1A4b |
| Small boats 4 stroke | 1A4b, c |
| Equipment 2 stroke | 1A4b, c |
| Equipment 4 stroke, tractor | 1A2g-vii, 1A4a, b, c, |

* For road transport the source split is more detailed in the sub-model. See section 3.2.4.2.

Source: Statistics Norway/Norwegian Environment Agency

Table 1.4. Combinations of fuels and sources in use

| | Direct fired furnaces | Gas turbines | Boilers | Small stoves | Flaring | Passenger car | Light duty vehicles | Heavy duty vehicles | Motorcycle | Moped | Snowscooter | Railway | Aviation jet/turboprop | Aviation helicopter | Aviation small craft | Ships | Small boats 2 stroke | Small boats 4 stroke | Equipment 2 stroke | Equipment 4 stroke, tractor |
|-----------------------|-----------------------|--------------|---------|--------------|---------|---------------|---------------------|---------------------|------------|-------|-------------|---------|------------------------|---------------------|----------------------|-------|----------------------|----------------------|--------------------|-----------------------------|
| Coal | X | .. | X | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Coke | X | .. | X | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Petrol coke | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Fuel wood | .. | .. | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Wood waste | .. | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Black liquor | .. | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Wood pellets | .. | .. | X | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Wood briquettes | .. | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Charcoal | .. | .. | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Natural gas | X | X | X | .. | X | X | .. | X | .. | .. | .. | .. | .. | .. | .. | X | .. | .. | .. | .. |
| Refinery gas | X | .. | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| CO gas | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Landfill gas | .. | .. | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Biogas | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Fuel gas | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| LPG | .. | .. | X | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Motor gasoline | .. | .. | .. | .. | .. | X | X | X | X | X | X | .. | .. | .. | .. | .. | X | X | X | X |
| Aviation gasoline | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | X | .. | .. | .. | .. | .. |
| Kerosene (heating) | .. | .. | X | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Jet kerosene | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | X | X | .. | .. | .. | .. | .. | .. |
| Auto diesel | .. | .. | X | .. | .. | X | X | X | .. | .. | .. | X | .. | .. | .. | .. | .. | X | .. | X |
| Marine gas oil/diesel | X | X | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | X | .. | .. | .. | .. |
| Light fuel oils | .. | .. | X | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | X | .. | .. | .. | X |
| Heavy distillate | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | X | .. | .. | .. | .. |
| Heavy fuel oil | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | X | .. | .. | .. | .. |
| Municipal waste | .. | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |
| Special waste | X | .. | X | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. | .. |

Source: Statistics Norway

The sources for non-combustion emissions and for combustion without energy use are based on EMEP/NFR categories, with further subdivisions where more detailed methods are available.

1.5 Key Categories

Information about key categories is given in Appendix A.

1.6 QA/QC and Verification methods

This chapter describes general QA/QC procedures. For source specific QA/QC, see each source sector for detailed descriptions.

The QA/QC work has several dimensions. In addition to accuracy, also timeliness is essential. As these two aspects may be in conflict, the QA/QC improvements in recent years have been focused on how to implement an effective QA/QC procedure and how to obtain a more efficient dataflow in the inventory system.

During the past years several quality assurance and quality control procedures for the preparation of the national emission inventory have been established in Norway. Statistics Norway made its first emission inventory for some gases in 1983 for the calculation year 1973. The emission estimation methodologies and the QA/QC procedures have been developed continuously since then. Norway has implemented a formal quality assurance/quality control plan, which covers the reporting of long-range transboundary air pollution as well as greenhouse gases. A detailed description of this is presented in Annex V in the National Inventory Report for Norway.

The established QA/QC procedures include the following:

- The Norwegian Environment Agency is the national entity designated to be responsible for the reporting of the national inventory of greenhouse gases to the UNFCCC and the reporting of long-range transboundary air pollution to the UNECE. This includes coordination of the QA/QC procedures;
- Statistics Norway is responsible for the quality control system with regard to technical activities of the inventory preparation;
- General inventory level QC procedures, as listed in table 6.1 in chapter 6 of the 2006 IPCC Guidelines (IPCC 2006), is performed every year;
- Source category-specific QC procedures are performed for all key categories and some non-key categories; with regard to emission factors, activity data and uncertainty estimates.

1.6.1 QA Procedures

According to the IPCC Good practice guidance, good practice for QA procedures requires an objective review to assess the quality of the inventory and to identify areas where improvements could be made. Furthermore, it is good practice to use QA reviewers that have

not been involved in preparing the inventory. In Norway, the Norwegian Environment Agency is responsible for reviewing the inventory with regard to quality and areas for improvement. For most source categories it is a person within the Norwegian Environment Agency who has not been involved in the calculations and the quality controls who performs the QA for the particular source.

Norway has performed several studies comparing inventories from different countries (Haakonsen et al. 2000). Verification of emission data is another element to be assessed during the elaboration of a QA/QC and verification plan.

Both Statistics Norway and the Norwegian Environment Agency are responsible for archiving the data they collect and the estimates they calculate with associated methodology documentation and internal documentation on QA/QC. Due to the differences in the character of data collected, Norway has chosen to keep archiving systems in the separate institutions, which means that not all information is archived at a single location. These archiving systems are, however, consistent, and operate under the same rules. Although the data are archived separately, all can be accessed efficiently during a review.

1.6.2 General QC procedures

The Norwegian emission inventory is produced in several steps. Preliminary estimates are first produced by Statistics Norway 4-5 months after the end of the inventory year. These data are based on preliminary statistics and indicators and data that have been subjected to a less thorough quality control. The "final" update takes place about one year after the inventory year. At this stage, final statistics are available for all sources. Recalculations of the inventory are performed annually, as methodological changes and refinements are implemented. In itself, this stepwise procedure is a part of the QA/QC-procedure since all differences in data are recorded and verified.

For each of the steps described above, general quality control procedures are performed, but with different levels of detail and thoroughness as mentioned. The national emission model was revised in 2002 in order to facilitate the QC of the input data rather than the emission data only. Input data include emissions reported from large plants, activity data, emission factors and other estimation parameters.

In the following, the procedures listed in table 6.1 in chapter 6 of the 2006 IPCC Guidelines (IPCC 2006), are described, as well as how these checks are performed for the Norwegian emission inventory.

Check that assumptions and criteria for the selection of activity data and emissions factors are documented

Thorough checks of emission factors and activity data and their documentation have been performed for existing emission sources. When new sources appear (for example a new industrial plant) or existing sources for the first time are recognised as a source, the Norwegian Environment Agency delivers all relevant information to Statistics Norway. This information is

then thoroughly checked by the inventory team at Statistics Norway. All changes in methodologies or data are documented and kept up to date.

Check for transcription errors in data input and references

Activity data are often statistical data. Official statistical data undergo a systematic revision process, which may be manual or, increasingly frequently, computerised. The revision significantly reduces the number of errors in the statistics used as input to the inventory. Furthermore, all input data (reported emissions, emission factors and activity data) for the latest inventory year are routinely compared to those of the previous inventory year, using automated procedures. Large changes are automatically flagged for further, manual QC. In addition, implied emission factors are calculated for emissions from stationary combustion at point sources. The IEFs are subjected to the same comparison between the years t and $t-1$. The most thorough checks are made for the gases and categories with the largest contribution to total emissions.

Check that emissions are calculated correctly

When possible, estimates based on different methodologies are compared. The Norwegian Environment Agency and Statistics Norway control and verify emission data reported to the Norwegian Environment Agency by industrial enterprises, registered in the database *Forurensning*. First, the Norwegian Environment Agency checks the data received from these plants, and if errors are discovered, they may then ask the plants' responsible to submit new data. Subsequently, Statistics Norway makes, where possible, occasional comparable emission calculations based on activity data sampled in official statistics, and deviations are explained through contact with the plants.

Check that parameter and emission units are correctly recorded and that appropriate conversion factors are used

All parameter values are compared with values used in previous years and with any preliminary figures available. Whenever large deviations are detected, the value of the parameter in question is first checked for typing errors or unit errors. Changes in emissions from large plants are compared with changes in activity level. If necessary, the primary data suppliers (e.g. The Norwegian Petroleum Directorate, Norwegian Public Roads Administration, various plants etc.) are contacted for explanations and possible corrections.

Check the integrity of database files

Control checks of whether appropriate data processing steps and data relationships are correctly represented are made for each step of the process. Furthermore, it is verified that data fields are properly labelled and have correct design specifications and that adequate documentation of database and model structure and operation are archived.

Check that the movement for inventory data among processing steps is correct

Statistics Norway has established automated procedures to check that inventory data fed into the model does not deviate too much from the figures for earlier years, and that the calculations within the model are correctly made. Checks are also made that emissions data are correctly transcribed between different intermediate products. The model is constructed so that it gives error messages if factors are lacking, which makes it quite robust to miscalculations.

Check that uncertainties in emissions and removals are estimated correctly

For long-range transboundary air pollutants the last uncertainty analysis was undertaken in 2001. See further information about the uncertainty analysis in section 1.7 and Appendix C.

Undertake review of internal documentation

For some sources, expert judgements dating some years back are employed with regard to activity data/emission factors. In most of the cases these judgements have not been reviewed since then, and may not be properly documented, which may be a weakness of the inventory. The procedures have improved the last few years, and the requirements for internal documentation to support estimates are now quite strict; all expert judgements and assumptions made by the Statistics Norway staff should be documented. This should increase reproducibility of emissions and uncertainty estimates. The model at Statistics Norway has improved the process of archiving inventory data, supporting data and inventory records, which does facilitate review. The model runs are stored and may be reconstructed, and all input data from the Norwegian Environment Agency as well as notes with explanations on changes in emissions are stored. This is a continuous process of improvement at Statistics Norway.

Check of changes due to recalculations

Emission time series are recalculated every year to ensure time series consistency. The recalculated emission data for a year is compared with the corresponding figures estimated the year before. The intention is to explain all major differences as far as possible. Changes may be due to revisions in energy data, new plants, correction of former errors and new emission methodologies.

Undertake completeness checks

Estimates are reported for all source categories and for all years as far as we know, apart from a few known data gaps, which are listed in section 1.8. There may, of course, exist sources which are not covered. However, we are quite certain that emissions from potentially additional sources are very small or negligible. During comparisons with previous emission estimates, any emission calculations that have been erroneously omitted during the most recent production cycle will be identified and included.

Compare estimates to previous estimates

Internal checks of time series for all emission sources are performed every year when an emission calculation for a new year is done. It is then examined whether any detected inconsistencies are due to data or/and methodology changes. For example, in 2016 Statistics Norway/the Norwegian Environment Agency calculated emission data for 2015 for the first time. These data were compared with the 2014 figures for detection of any considerable deviations. Some large deviations may be correct, caused for instance by the shutdown of large industrial plants or the launch of new ones.

1.6.3 Source category-specific QC procedures

Statistics Norway and the Norwegian Environment Agency have carried out several studies on specific emission sources, e.g. emissions from road, sea, and air transport. These projects are

repeated in regular intervals when new information is available. During the studies, emission factors have been assessed and amended in order to represent the best estimates for national circumstances, and a rationale for the choice of emission factor is provided. The emission factors are often compared with default factors from the most recent EMEP/EEA air pollutant inventory guidebook and emission factors from other literature. Furthermore, activity data have been closely examined and quality controlled.

The QC procedures with regard to emission data, activity data and uncertainty estimates for the different emission sources are described in the QA/QC-chapters of the relevant source-categories. The source category-specific analyses have primarily been performed for key categories on a case-by-case basis, which is described as being good practice. The QA/QC process for many of the sources could be improved.

In the following is a more detailed description of QC of emission data reported from plants:

Plant emission data that are used in the European Emission trading system (EU-ETS) will undergo annual QC checks. Activity data and emission estimates from plants that are included in the EU-ETS undergoes annual third-party verification. The source-specific QC checks for other plants are performed as part of the controls of the reporting under the emission permits.

The plant specific data undergo further QC from the emission inventory team at the Norwegian Environment Agency before figures are sent to Statistics Norway for inclusion in the emission inventory. Statistics Norway is responsible for reporting the results of the key category analysis to the Norwegian Environment Agency, and the agency places special emphasis on plants that belong to key categories.

At some point since the inclusion in the inventory, each plant has been QC checked more thoroughly, including:

- An assessment and documentation of measurements and sampling
 - Measurement frequency
 - Sampling
 - Use of standards (e.g. ISO)
- An assessment and explanation of changes in emissions over time (e.g. changes in technology, production level or fuels) (annual check)
- An assessment of time-series consistency back to 1990 in cooperation with Statistics Norway (if plant emission data are missing for some years and estimates are made using aggregate activity data and emission factors)
- A comparison of plant emissions to production ratios with those of other plants, including explanations of differences
- A comparison of the production level and/or fuel consumption with independent statistics (in collaboration with Statistics Norway)
- An assessment of reported uncertainties (including statistical and non-statistical errors) to the extent this has been included in the reporting

The QC checks are made in close cooperation with the emission reporting plants. For more details of QA/QC of specific source categories, see “source specific QA/QC” in relevant chapters.

1.6.4 Verification studies

In general, the final inventory data provided by Statistics Norway are checked and verified by the Norwegian Environment Agency.

In the following, some verification studies which have been performed are briefly described. Emission estimates for a source are often compared with estimates performed with a different methodology.

In 2004, the Nordic Council of Ministers initiated a new project that was finalised in 2006. This project focused on NMVOC, heavy metals and POPs. An unpublished, final report has been worked out, containing the following elements:

- comparisons of the emission estimation methodologies and emission factors used in each country (review)
- identification of gaps in knowledge
- identification of possible "burden sharings" with respect to research areas (research taking place in one country, but used in all countries)
- discussions of the particular Nordic aspects influencing the emissions
- discussions of the possible contributions from research in the Nordic countries
- proposals for research areas

In 2006, the Nordic Council of Ministers initiated a new project that was finalised in 2010. This project focused on emission of particulate matter. The final report contains the following elements:

- comparisons of the emission estimation methodologies and emission factors used in each country (review)
- identification of gaps in knowledge
- discussions of the particular Nordic aspects influencing the emissions
- discussions of the possible contributions from research in the Nordic countries
- proposals for research areas
- recommendations for further work

In 2015, a Nordic project started, financed by the Nordic Council of Ministers, with the aim to improve the Nordic emission inventories on heavy metals and POP compounds. In the proposed programme the experts working with air pollutant inventories in Denmark, Finland, Iceland, Norway and Sweden will have a yearly meeting for knowledge exchange. The main focus is to compare emission factors and methodologies used, inform about national studies performed, and study and set up further plans to develop national air pollution inventories, especially for emission sources where studies and other Nordic information sources can be used in developing methodologies suitable for Nordic countries.

In 2017, a Nordic project, financed by the Nordic Council of Ministers, went through the

emission factors for SLCP emissions from residential wood combustion in the Nordic countries. The overall objective of this project is to improve the Nordic emission inventories of SLCPs (Kindbom et al. 2017). This project included comparisons of emission factors for elemental carbon (EC), organic carbon (OC), particulate matter (PM_{2.5}), methane (CH₄) and non-methane volatile organic compounds (NMVOC).

1.7 General uncertainty evaluation

1.7.1 Acidifying substances and NMVOC

The emission estimates for long-range air pollutants in the Norwegian emission model may be ranked roughly in order of increasing uncertainty as follows:

$$\text{SO}_2 < \text{NO}_x < \text{NH}_3 \approx \text{NMVOC}$$

The sources of uncertainty in the emission estimates include sampling errors, poor relevance of emission factors or activity data, and gross errors.

Evaluation of the uncertainty in the long-range air pollutants is given in the report Rypdal and Zhang (2001). Summary tables with the results are given in Appendix C.

1.7.2 Heavy metals and POPs

The uncertainty is generally higher for HM and POPs than for other components in the Norwegian emission model except for N₂O. There are various reasons for this high uncertainty. The most important reason is that there is limited information about emission factors, and it is not clear how usable the emission factors found in international literature are for Norwegian conditions. Emission factors for some HM and POPs components are insufficient for some sources, so emission factors for similar sources have then been used. In addition it is not certain that all emission sources are known or sufficiently mapped. The industrial reporting to the Norwegian Environment Agency has improved in recent years. The reported figures can, however, vary a great deal from one year to another. For earlier years they can be insufficient, and since HM and POPs are to be calculated from 1990, recalculations are necessary. These recalculations are based on a combination of assumptions and knowledge of the plants. Emission figures from the early 1990s are therefore more uncertain than figures produced today.

1.8 General Assessment of Completeness

Norway is requested to report emissions to UNECE for the pollutants restricted by CLRTAP (Convention on Long-Range Transboundary Air Pollution). Minimum reporting request each year includes the acidifying pollutants (NO_x, SO₂, NH₃) and NMVOC, the heavy metals Pb, Cd and Hg, particulate matter (TSP, PM₁₀ and PM_{2.5}), CO and the POPs dioxins, Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4), HCB and PCB.

Norway also voluntary reports the heavy metals As, Cr and Cu and BC. Emissions of Ni, Se and Zn are not estimated.

In terms of spatial coverage, the calculated air emissions cover all activities within Norway's jurisdiction.

In the case of temporal coverage, emission figures for all pollutants are reported for all years from 1990.

Regarding sectoral coverage, sources not covered in the inventory even if emissions can be expected, have been reported as Not Estimated (NE) in the NFR tables. The table below provides explanations and an overview of the sources and pollutants that are not estimated, even if emissions might be expected. In each sector chapter more details about completeness is given.

Table 1.5 Explanation to the Notation key NE

| NFR category | Substance(s) | Further details |
|--------------|--|---|
| 1.A | NH ₃ | According to the EMEP/EEA Guidebook, emission of ammonia (NH ₃) are not caused by combustion process, and hence in general not occurring in NFR category 1.A "Energy combustion". Emissions resulting from incomplete reaction of NH ₃ additive in denitrification processes will however occur in some combustion categories, and is estimated where relevant in the Norwegian inventory. 'NA' ('not applicable') might be a more accurate notation key in NFR categories where no NH ₃ additive is applied, and this will be investigated |
| 1.A.3.b.iv | TSP, PM ₁₀ , PM _{2,5} , BC | Emission factors for particulate matter (TSP, PM ₁₀ , PM _{2,5}) and black carbon (BC) are not provided in the EMEP/EEA Guidebook for the NFR category 1.A.3.b.iv "Road transport: Mopeds & motorcycles" |
| 1.A.3.b.vii | Pb, Hg, Cr, Cu | No emission factors for heavy metals are provided in the 2019 EMEP/EEA Guidebook for 1.A.3.b.vii "Road transport: Automobile road abrasion ". The use of 'NA' ('not applicable') in stead of 'NE' will be investigated |
| 1.B.1.b | NMVOC | Provision and use of emission factor for NMVOC from 1.B.1.b "Fugitive emission from solid fuels: Solid fuel transformation" in 2019 EMEP/EEA Guidebook will be investigated |
| 1.B.1.b | TSP, PM ₁₀ , PM _{2,5} , BC | Provision and use of emission factors for particulate matter and black carbon from 1.B.1.b "Fugitive emission from solid fuels: Solid fuel transformation" in 2019 EMEP/EEA Guidebook will be investigated |
| 1.B.1.b | Pb, Cd, Hg, As | Provision and use of emissions factor for heavy metals from 1.B.1.b "Fugitive emission from solid fuels: Solid fuel transformation" in 2019 EMEP/EEA Guidebook will be investigated |
| 1.B.1.b | PAH | Provision and use of emission factors for PAHs (Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) from 1.B.1.b "Fugitive emission from solid fuels: Solid fuel transformation" in 2019 EMEP/EEA Guidebook will be investigated |
| 1.B.1.b | PCB, HCB | No emission factors are provided for PCB and HCB in 2019 EMEP/EEA Guidebook for 1.B.1.b "Fugitive emission from solid fuels: Solid fuel transformation" and use of the notation key 'NE' (not estimated) is recommended by the EMEP/EEA Guidebook |
| 1.B.2.a.i | SOx | No emission factors provided in 2019 EMEP/EEA Guidebook for 1.B.2.a.i "Fugitive emissions oil: Exploration, production, transport" and use of the notation key 'NE' (not estimated) is recommended by the EMEP/EEA Guidebook |
| 1.B.2.b | SOx | No emission factors provided in 2019 EMEP/EEA Guidebook for 1.B.2.b "Fugitive emissions from natural gas (exploration, production, processing, transmission, storage, distribution and other)" and use of the notation key 'NE' (not estimated) is recommended by the EMEP/EEA Guidebook |

| | | |
|------------|------------------------|--|
| 1.B.2.d | Hg, As | Provision and use of emission factor for heavy metals from 1.B.2.d " Other fugitive emissions from energy production " in 2019 EMEP/EEA Guidebook will be investigated |
| 2.D.3.c | NMVOC, PM | Emissions are not estimated |
| 3.D.a.2 | NO _x | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |
| 3.D.a.4 | NH ₃ | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |
| 3.D.e | NH ₃ | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |
| 5.B.1 | NMVOC | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |
| 5.B.2 | NMVOC | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |
| 5.C.1 | NMVOC, NH ₃ | Emissions are not estimated because of lack of activity data |
| 5.C.1.b.iv | All pollutants | Emissions are not estimated because of lack of activity data |
| 5.C.2 | All pollutants | Assumed to not occur in Norway – the notation key should be changed |
| 5.D | NH ₃ | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |
| 5.E | Hg | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |
| 5.E | Dioxins | Emissions are not estimated because there is no emission factor or methodology description in the EMEP/EEA Guidebook |

2 Explanation of key trends

2.1 Acidifying substances and NMVOC

2.1.1 Total acidifying emission

Emissions of gases that transform into acid can be expressed in terms of acid equivalents. Total emissions of the three gases NO_x , SO_2 and NH_3 measured as acid equivalents have been reduced by 24 per cent since 1990, from 7 890 tonnes acid equivalents to 6 030 tonnes acid equivalents in 2018. SO_2 and NO_x emissions have been reduced by 67 and 20 per cent since 1990, respectively, while NH_3 emissions have remained more or less constant. In 1990, NO_x constituted 55 per cent of the acidifying emissions, NH_3 25 per cent and SO_2 20 per cent, while, in 2018, NO_x , NH_3 and SO_2 were responsible for 58, 34 and 8 per cent of these emissions, respectively. Norway has met the 2010 targets defined by the Gothenburg Protocol for SO_2 and NMVOC, while NO_x and NH_3 emissions are above the Gothenburg 2010 target.

The 2020 targets defined by the revised Gothenburg Protocol have already been met for SO_2 and NO_x .

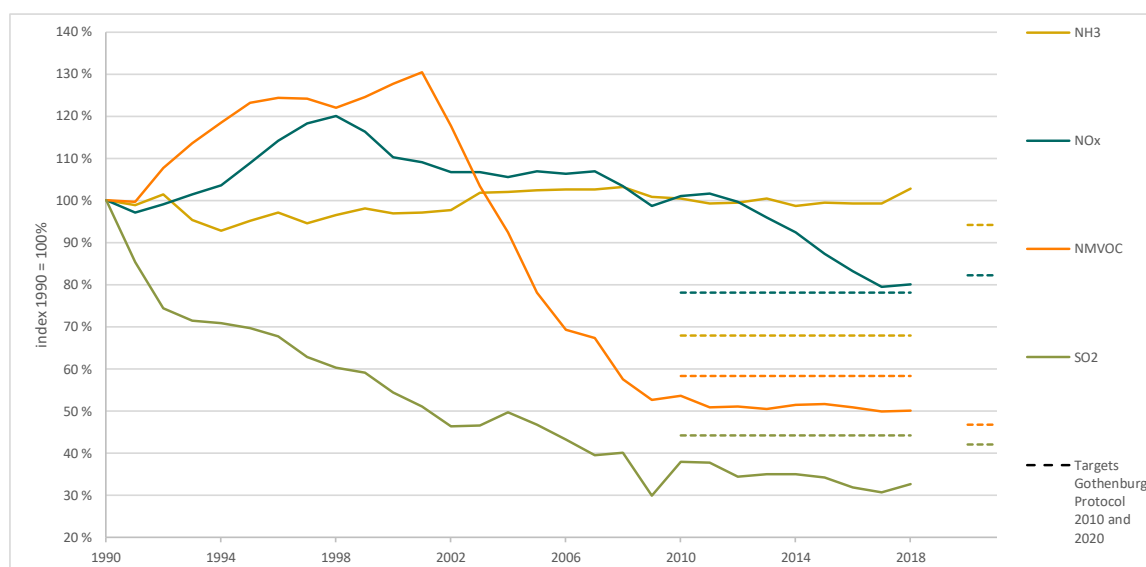


Figure 2.1. Trends in emissions for NO_x , SO_2 , NH_3 and NMVOC. 1990-2018. Index 1990 = 100%

Source: Statistics Norway/Norwegian Environment Agency

2.1.2 NO_x

- 2010 commitment of the Gothenburg Protocol: 156 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: a 23 per cent reduction compared to emissions in 2005, 167 570 tonnes.

Norway's NO_x emissions totaled 160 230 tonnes in 2018. NO_x emissions have been reduced by 20 per cent since 1990 and by 25 per cent since 2005. The biggest sources of NO_x emissions in

2018 were transport and combustion in energy industries, accounting for 42 and 29 per cent of total emissions, respectively. Emissions from combustion in the energy industries sector overall have increased by 69 per cent since 1990. Within the energy industries sector, manufacturing of solid fuels and other energy industries accounts for 94 per cent, and have increased by 80 per cent since 1990.

Emissions from the transport sector overall have been reduced by 35 per cent since 1990. These reductions nevertheless hide significant differences within the sector: emissions from domestic and international aviation have increased by 69 per cent and 83 per cent respectively since 1990, whereas emissions from passenger cars have been reduced by 60 per cent and emissions from heavy duty vehicles and buses have been reduced by 64 per cent.

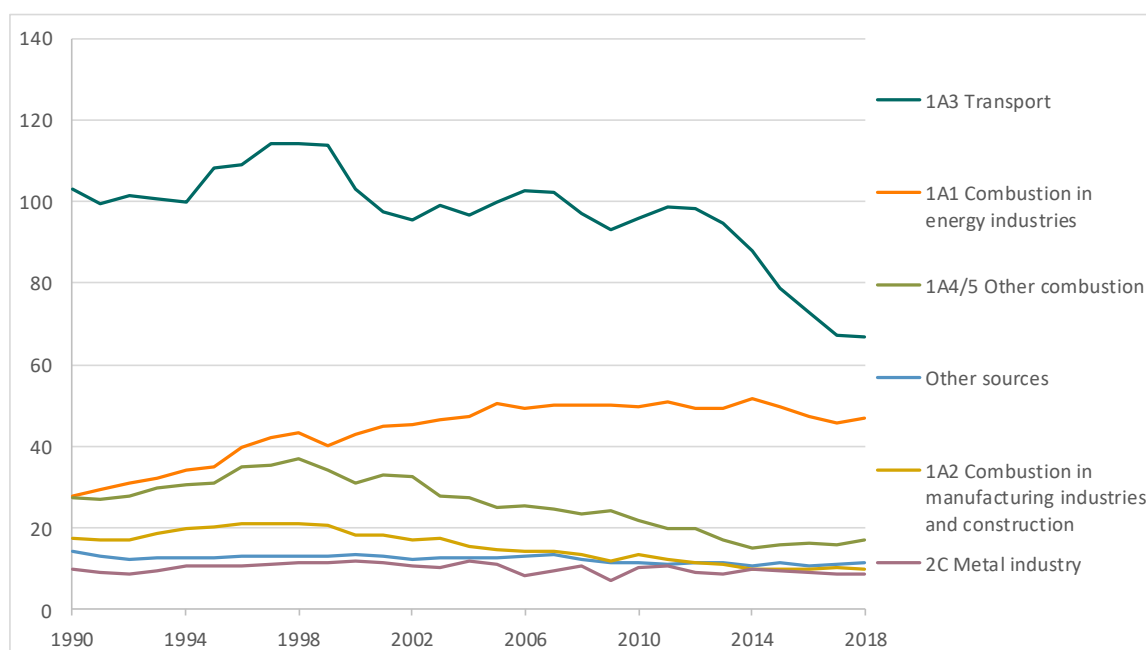


Figure 2.2. Trends in NO_x emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

Emissions from aviation have increased primarily due to increased traffic. Within road transport, traffic has also increased significantly from 1990 to 2018. NO_x emissions have however been reduced due to policy measures. Stricter emission requirements for new passenger cars, heavy duty vehicles and buses, is the main cause of the reduction.

Emissions from national navigation increased up to the year 2012, and has mostly decreased since then, except for a 4 per cent increase from 2017 to 2018. Emissions are now back to the level of emissions in 1990. This is partly due to measures implemented by the NO_x Fund, which is financed by industry and businesses. The measures are not directly linked to fuel efficiency, and energy consumption in national navigation has not had the same decrease as the emissions.

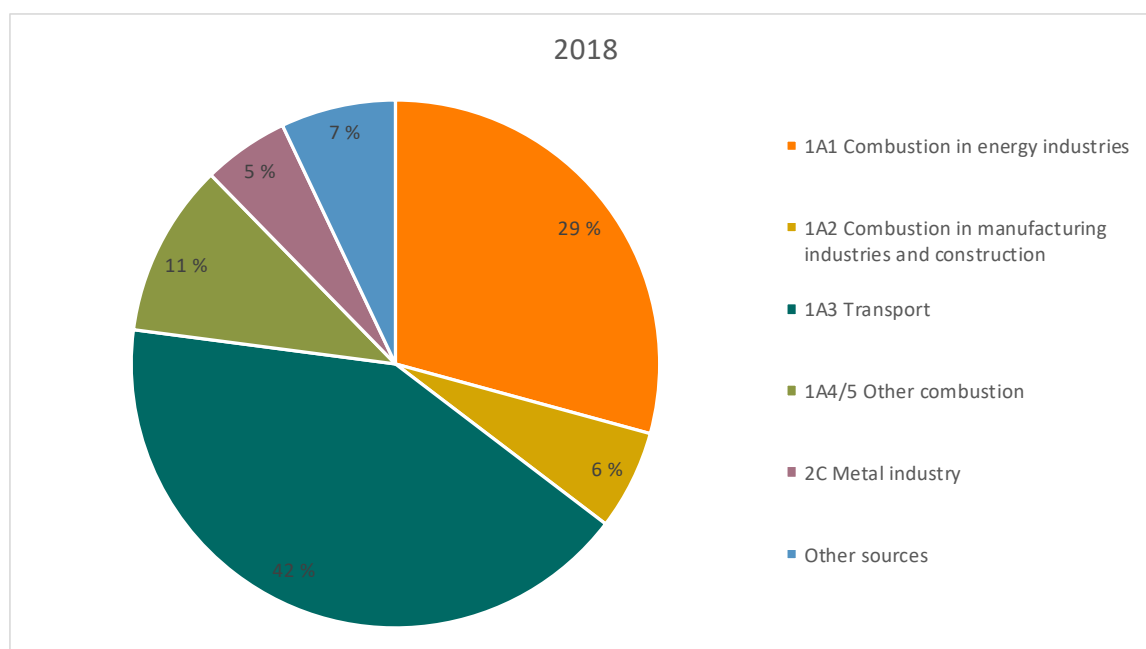


Figure 2.3. Distribution of NO_x emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.1.3 SO₂

- 2010 commitment of the Gothenburg Protocol: 22 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: a 10 per cent reduction compared to emissions in 2005, 20 920 tonnes.

Norway's SO₂ emissions totaled 16 220 tonnes in 2018. Both the 2010 and 2020 commitments of the Gothenburg Protocol for SO₂ emissions have been fulfilled for some years.

The SO₂ emissions in Norway in 2018 have been reduced by 67 per cent since 1990. This has been achieved by pollution control, the closure of pollution intensive businesses, a reduction of sulfur content in petroleum products and a reduced consumption of petroleum products. Emissions have increased by 6 per cent between 2017 and 2018, due to an increase in emissions both from metal industry and from combustion in energy industries.

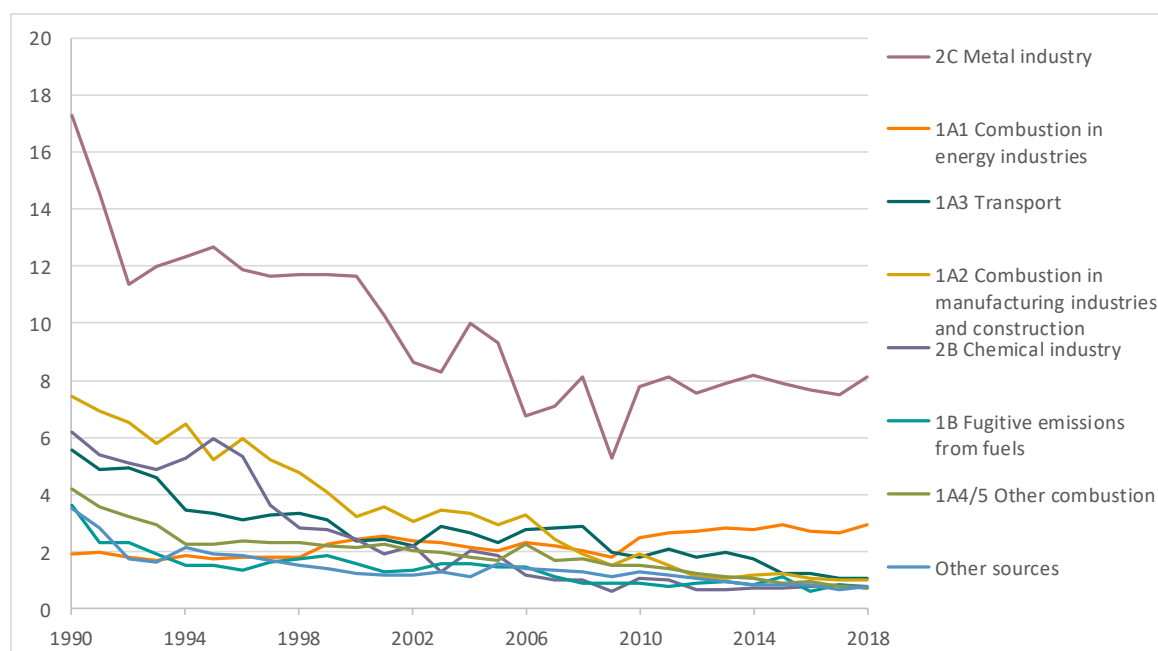


Figure 2.4. Trends in SO₂ emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

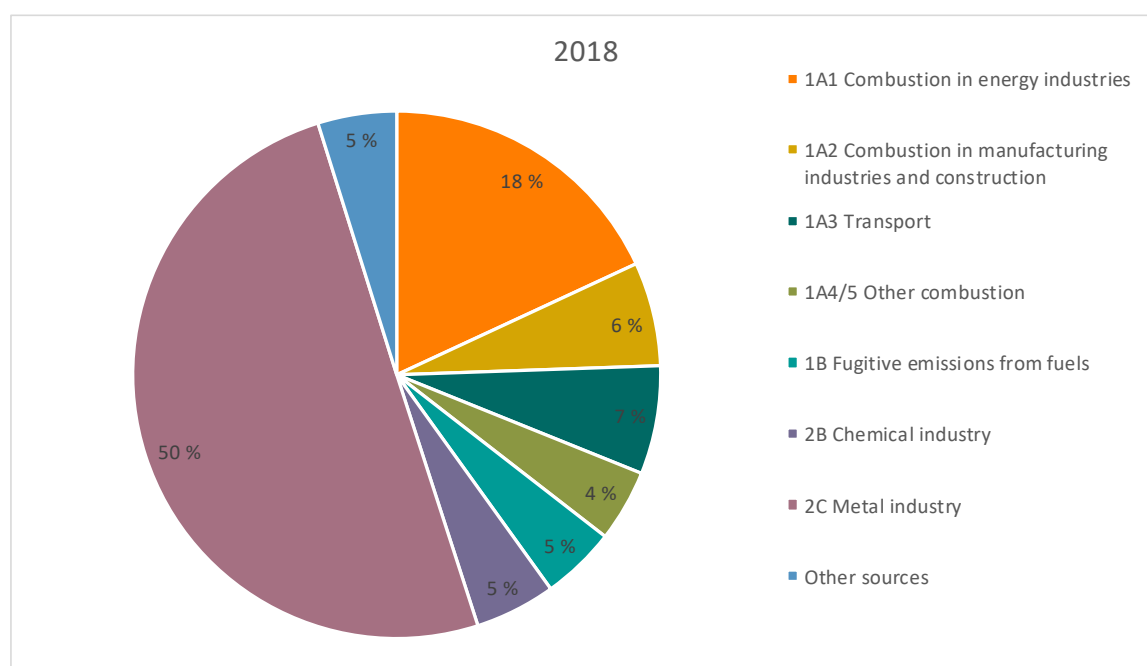


Figure 2.5. Distribution of SO₂ emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

Metal industry was the largest source of SO₂ emissions in 2018, representing 50 per cent of total emissions. Emissions from this sector have been reduced by 53 per cent since 1990, primarily due to reductions in the ferroalloys production. Nonetheless, the production of ferroalloys remains the most significant source of emissions within metal industry, being responsible for 69 per cent of emissions from the metal industry in 2018. Emissions from the metal industry have increased by 9 per cent between 2017 and 2018.

Combustion in energy industries is the second largest source of emissions in 2018, accounting for 18 per cent of the total emission. Emissions from combustion in energy industries have increased by 52 per cent since 1990. Within this sector, public electricity and heat production has been the largest source of SO₂ emissions since 1990. It accounted for 72 per cent of emissions from the sector combustion in energy industries in 2018.

Transport constituted 7 per cent of total emissions of SO₂ in 2018. Emissions from transport have been reduced by 81 per cent since 1990, mainly due to less sulphur content in fuels. Most of the reduction took place at the beginning of the period. Indeed, in 1994, emissions were reduced by almost 40 per cent compared to 1990. Emissions from all transport subcategories have been reduced. Notably, there are currently only significant emissions from national navigation in this category, representing, in 2018, 92 per cent of the SO₂ emissions in the transport category.

Emissions from combustions in manufacturing industries and construction have decreased by 86 per cent since 1990, contributing to only 6 per cent of the total SO₂ emissions in 2018. Similarly, process emissions from chemical industries, including carbide production, have decreased by 87 per cent since 1990, contributing to 5 per cent of the total SO₂ emissions in 2018. The reduction is a result of lower production and closure of two plants.

2.1.4 NH₃

- 2010 commitment of the Gothenburg Protocol: 23 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: an 8 per cent reduction compared to emissions in 2005, 31 840 tonnes.

Norway's NH₃ emissions totaled 34 750 tonnes in 2018. The 2010 commitment of the Gothenburg Protocol for NH₃ emissions has not yet been fulfilled. Despite the fact that the revised commitment for 2020 is higher than the 2010 commitment, it will require further reductions to reach this level.

The Norwegian emissions of NH₃ are at the same level in 2018 as in 1990. Agriculture is the dominant source, and was responsible for 95 per cent of the NH₃ emissions in Norway in 2018. Animal manure is the most predominant source of emissions. 55 per cent of the total Norwegian emissions of NH₃ originated from agricultural soils, of which application of animal manure to soils accounted for 69 per cent in 2018. In addition, 35 per cent of total NH₃ emissions originated from manure management in 2018. Within this sector, cattle and swine are the most important sources of emissions, dairy cattle, non-dairy cattle and swine representing respectively 33 per cent, 31 per cent and 14 per cent of the emissions from this sector in 2018.

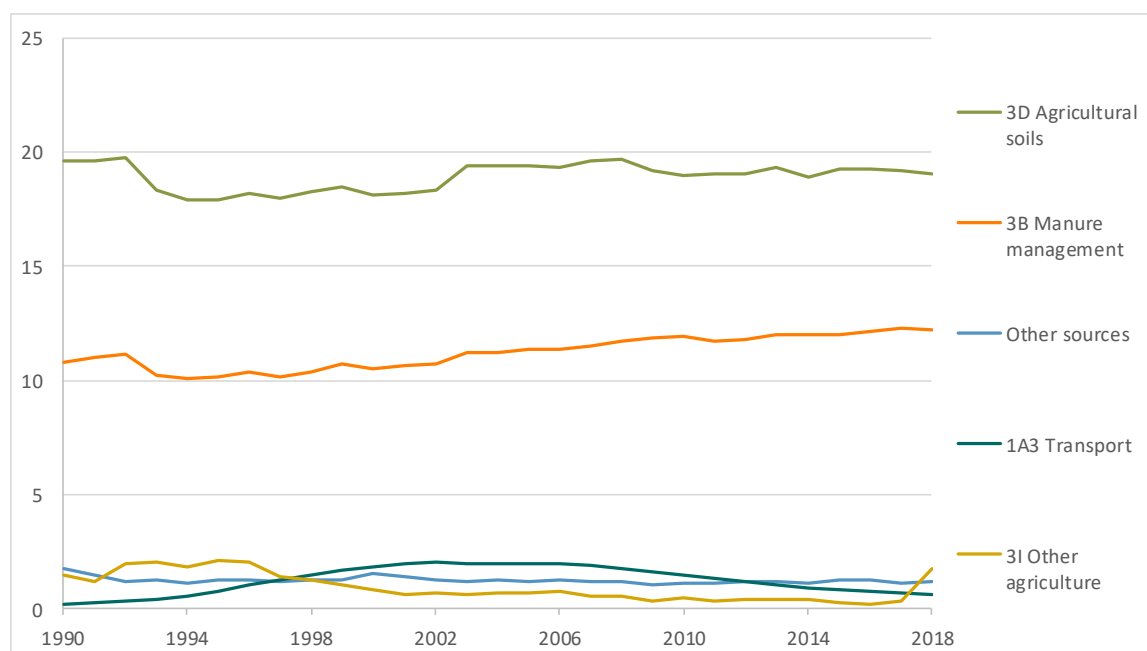


Figure 2.6. Trends in NH₃ emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

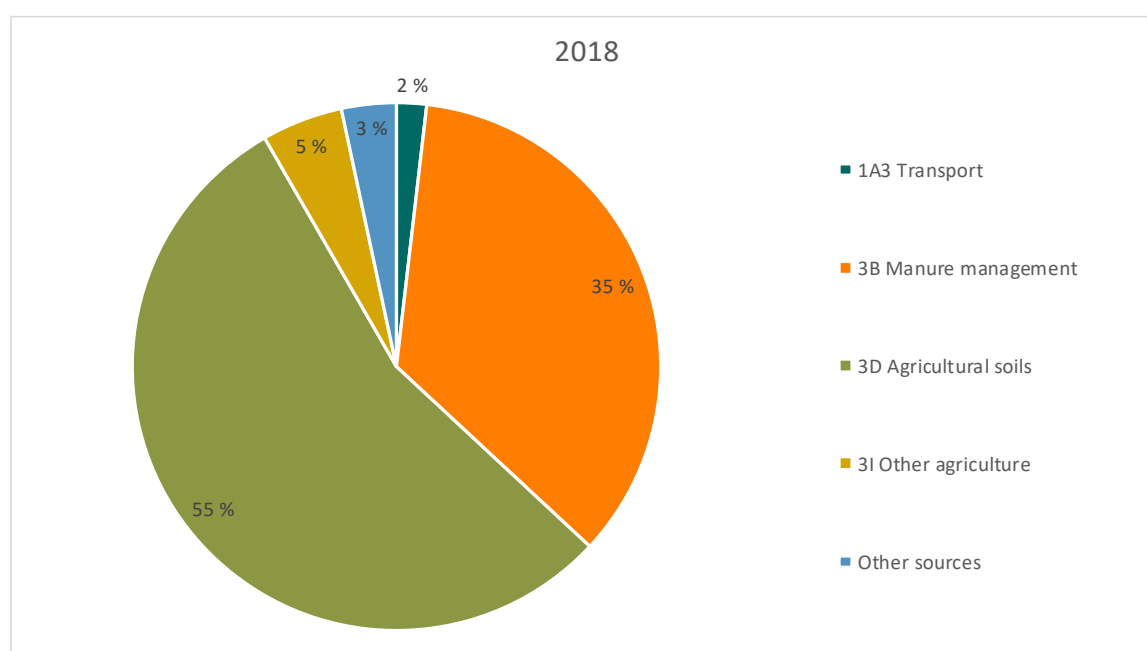


Figure 2.7. Distribution of NH₃ emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

Emissions of NH₃ from passenger cars increased up to a level of around 2000 tonnes due to the introduction of three-way catalytic converters. However, emissions from cars have since decreased considerably, partly due to improved catalyst technologies and partly due to a shift from petrol to diesel cars.

2.1.5 NMVOC

- 2010 commitment of the Gothenburg Protocol: 195 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: a 40 per cent reduction compared to emissions in 2005, 156 750 tonnes.

Norway's NMVOC emissions totaled 167 300 tonnes in 2018. The 2010 commitment of the Gothenburg Protocol for NMVOC emissions has been fulfilled since 2008. The revised 2020 commitment will demand further reductions.

NMVOC emissions have been reduced by 50 per cent since 1990, and by 62 per cent since the peak in 2001. Loading of crude oil offshore was the main reason for the increase in emissions from 1990 to 2001. Measures to prevent these emissions resulted in a 84 per cent decrease in fugitive emissions from fuels from 2001 to 2011. Since 2011, fugitive emissions from fuels have increased by 7 percent because new fields have come into operation. Fugitive emissions from fuels represent 29 per cent of total emissions of NMVOC in 2018.

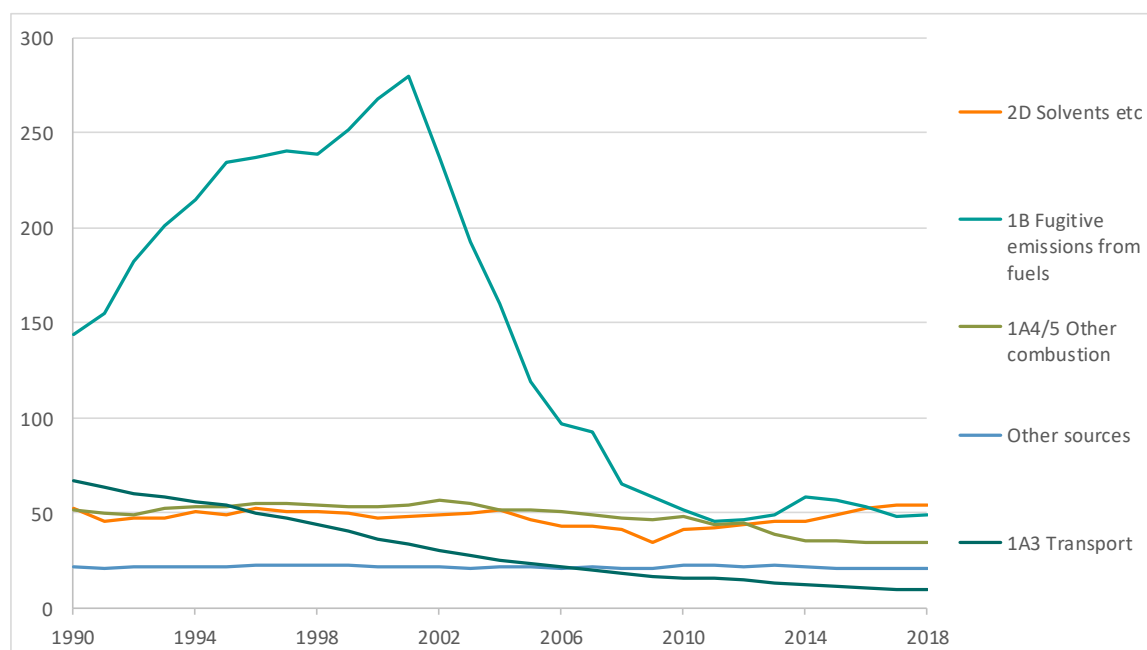


Figure 2.8. Trends in NMVOC emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

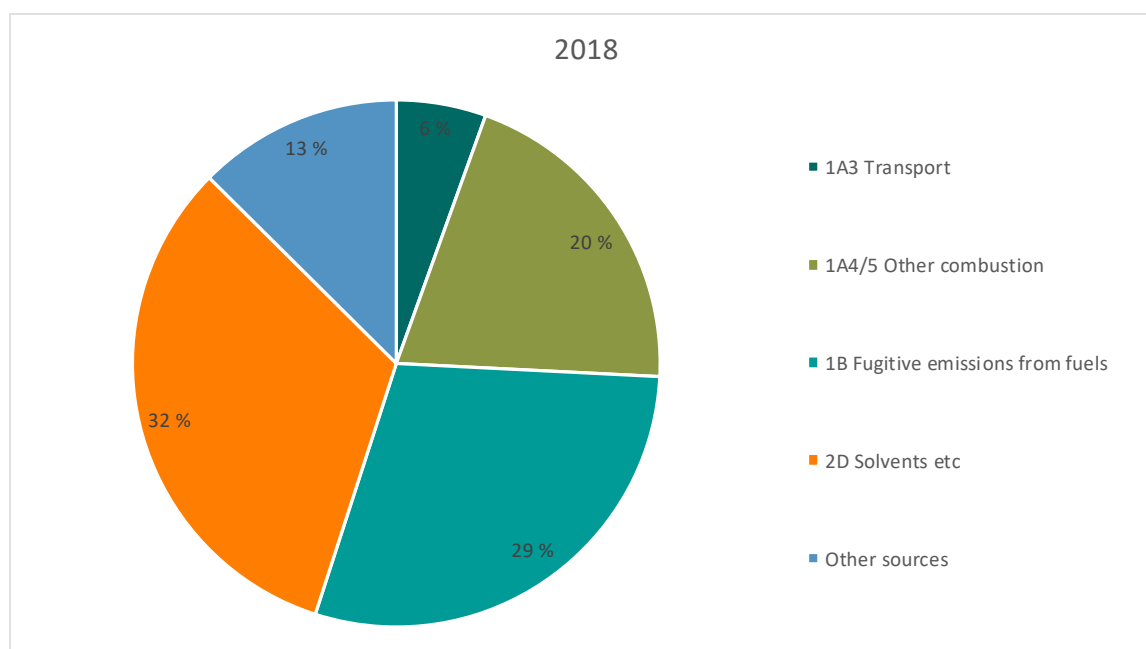


Figure 2.9. Distribution of NMVOC emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

32 per cent of the NMVOC emissions in 2018 originated from product use (solvents etc.). The most important subcategories within product use is "other solvent use" (2D3i), accounting for 51 per cent, and domestic solvent use including fungicides, accounting for 34 per cent of emissions within this sector.

The category "other combustion" (NFR 1A4 and 1A5) accounted for 20 per cent of total emissions of NMVOC in 2018. The two most significant sources of emissions within the sector are household and gardening (mobile) and stationary plants in the residential sector. These two subsectors were responsible for 98 per cent of emissions within this sector.

NMVOC emissions from transport have decreased by 86 per cent since 1990, mainly due to reductions in emissions from passenger cars and gasoline evaporation. Stricter emission standards for petrol passenger cars were implemented in 1989, and led to reduced emissions. In addition, the increased share of diesel cars within the vehicle fleet has strengthened the downwards trend.

2.2 CO

Emissions of carbon monoxide, CO, amounted to 434 320 tonnes in 2018. They have been reduced by 48 per cent since 1990 and by 1 per cent since 2017. This is mainly due to reductions in emissions from transport, which have been reduced by 89 per cent since 1990. Emissions from passenger cars represented 85 per cent of the transport sector in 1990. They have been reduced by 93 per cent since 1990, primarily due to stricter emission standards. Emissions from light duty vehicles have also been significantly reduced (95 per cent since 1990), albeit from a lower absolute level.

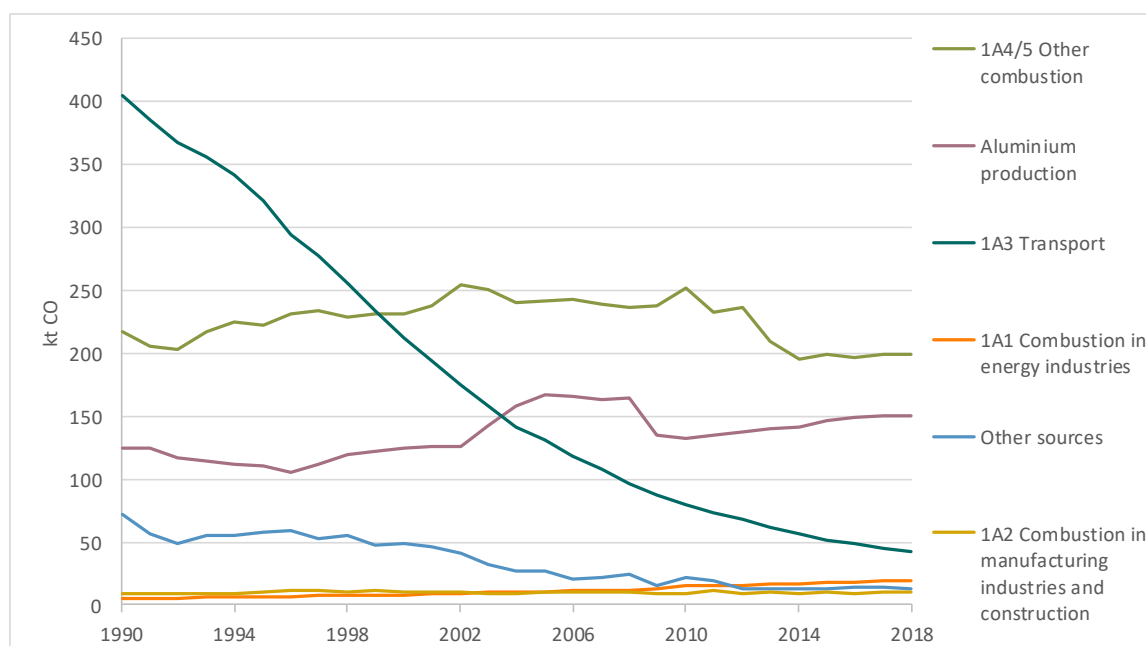


Figure 2.10. Trends in CO emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

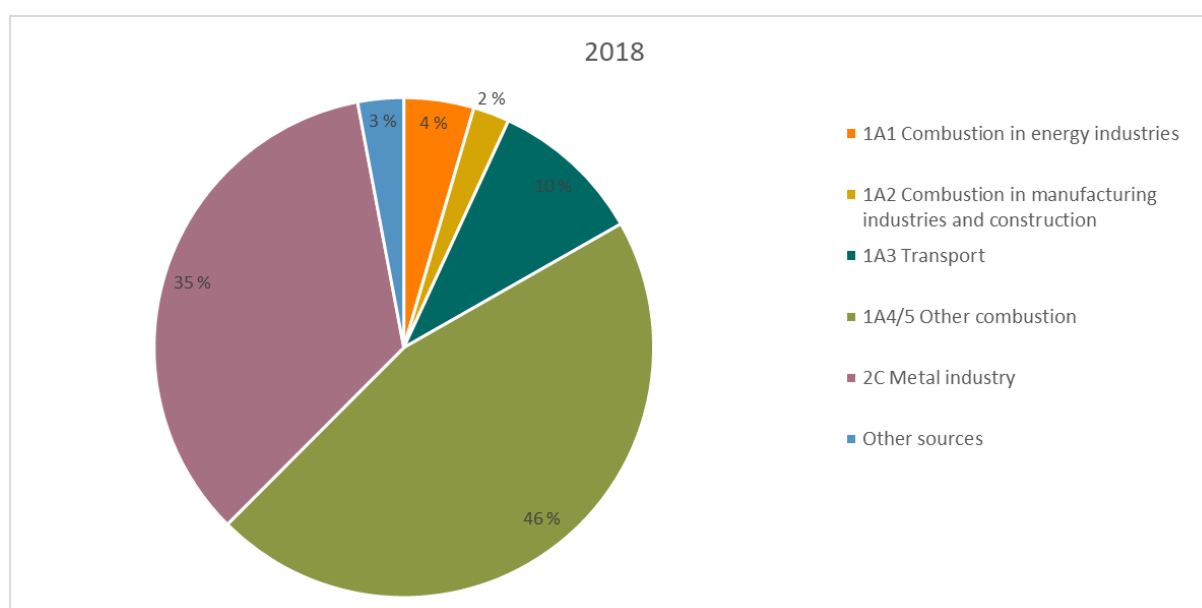


Figure 2.11. Distribution of CO emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

The category "other combustion" (NFR 1A4 and 1A5) and the metal industry were the largest sources of emissions in 2018, representing 46 and 35 per cent of the CO emissions, respectively.

The two most significant sources of emissions within "other combustion" are household and gardening (mobile) and stationary plants in the residential sector, which accounted for 51 and 47 per cent of the total emissions of the sector, respectively, in 2018. These emissions are mainly due to wood combustion for heating purposes. Emissions from household and gardening (mobile) in the residential sector have remained relatively stable since 1990. Emissions from

stationary plants in the residential sector were 12 per cent lower in 2018 than in 1990.

Within the metal industry the only source of CO emissions in 2018 is aluminium production. Emissions from aluminium production have increased by 44 per cent since 1990.

2.3 PM, POPs and heavy metals

Emissions of BC, PAH-4, dioxins, HCB, PCB, and all heavy metals except copper, although fluctuating throughout the period, shows a general downward trend since 1990.

Since 2017, BC and HCB emissions have decreased, and dioxines and PAH-4 emissions have increased, while PCB emissions have remained stable.

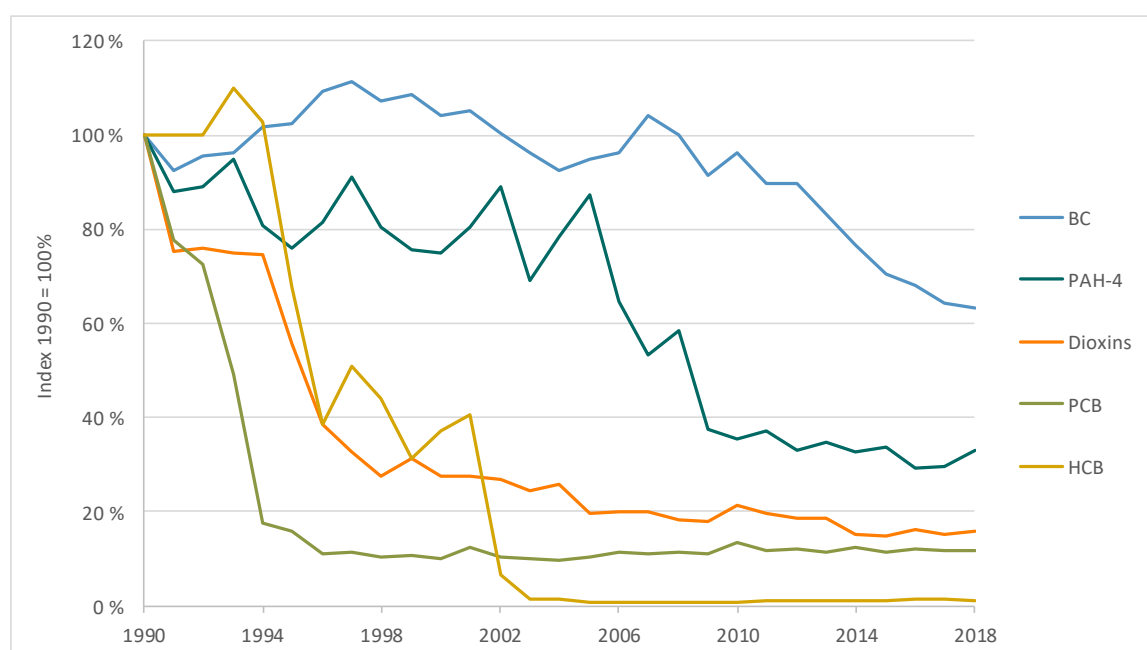


Figure 2.12. Trends in emissions for PAH-4, dioxins, PCB, HCB and BC. 1990-2018. Index 1990=100%

Source: Statistics Norway/Norwegian Environment Agency

The emissions of several hazardous substances including PAHs (polycyclic aromatic hydrocarbons) have been considerably reduced since 1990. The cause of the decrease is primarily reduced emissions within manufacturing and mining. Cleaning measures, changes in production processes and lower activity in some industries have all resulted in reduced emissions. Norway has set a national target to reduce emissions compared with the emission levels in 1995. Internationally, Norway has an obligation to reduce emissions of selected hazardous substances compared to the level of emissions in 1990 through the Aarhus Protocol.

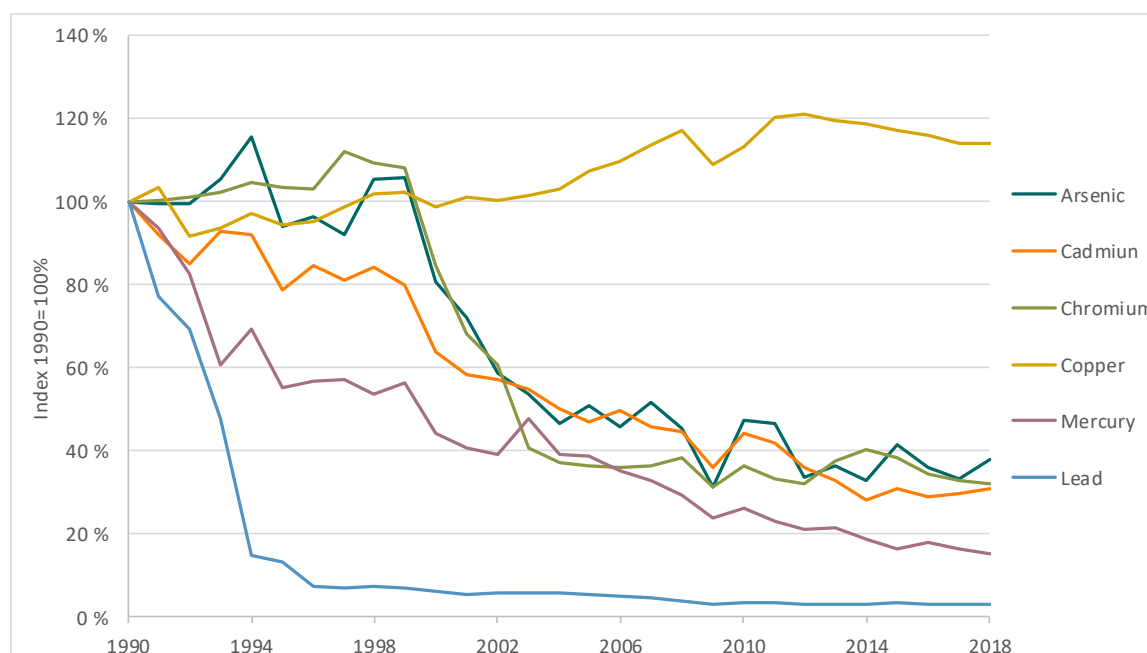


Figure 2.13 Trends in emissions for Heavy metals. 1990-2018. Index 1990=100%

Source: Statistics Norway/Norwegian Environment Agency

2.3.1 PM₁₀

Emissions of particulate matter (PM₁₀) totaled 33 160 tonnes in 2018. PM₁₀ emissions have decreased by 36 per cent since 1990.

The most important source of PM₁₀ emissions is stationary plants in the residential sector, included in "other combustion" (NFR 1A4 and 1A5). Emissions from "other combustion" amounted to 15 330 tonnes of PM₁₀ in 2018, of which 92 per cent stems from stationary plants, primarily due to wood burning in private households. Emissions from stationary plants in the residential sector have been reduced by 39 per cent since 1990. Since 2000, the reduction in particle emissions has been higher than reductions in wood consumption, due to an increased share of new technology in wood burning appliances.

Process emissions from manufacturing and mining amounted to 8 410 tonnes in 2018. Within the process sector, the largest sources were mineral industry and metal industry, which accounted for 49 and 30 per cent of the process category, respectively, in 2018. Emissions from metal industry have decreased by 61 per cent since 1990, mainly due to reduced production, whereas emissions from mineral industry have increased by 31 per cent since 1990. Emissions from this category are mainly from sandpit and rock-crushing plants. It should be noted that the emissions from this source varies highly from year to year, and that the data are uncertain.

Transport was responsible for 11 per cent of PM₁₀ emissions in 2018, of which 47 per cent originated from road abrasion and tyre and break wear.

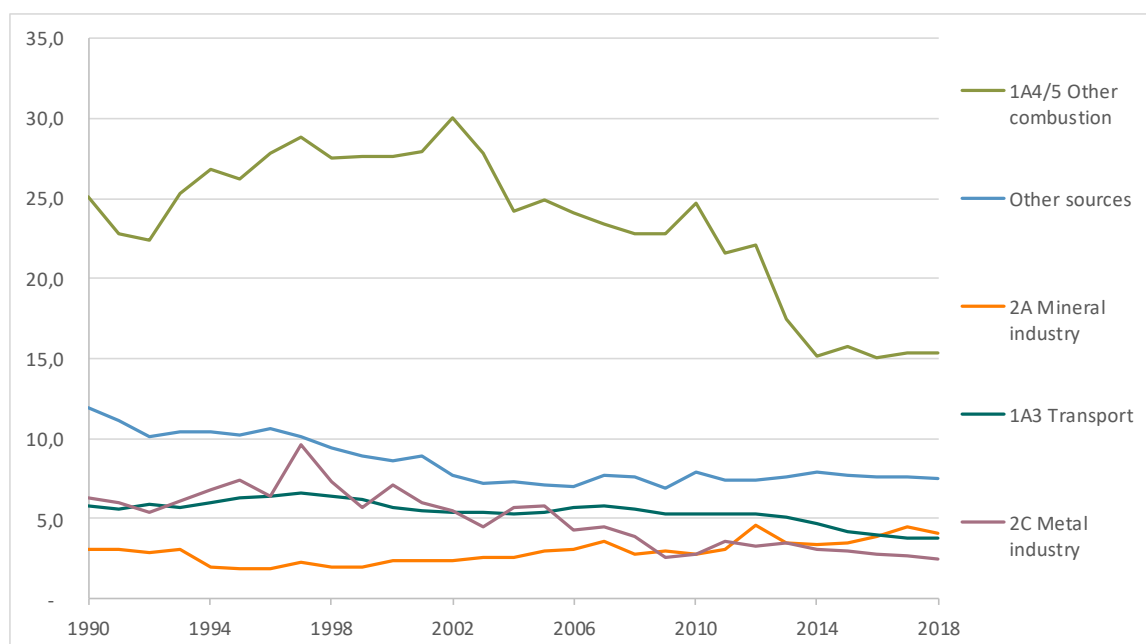


Figure 2.14. Trends in PM₁₀ emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

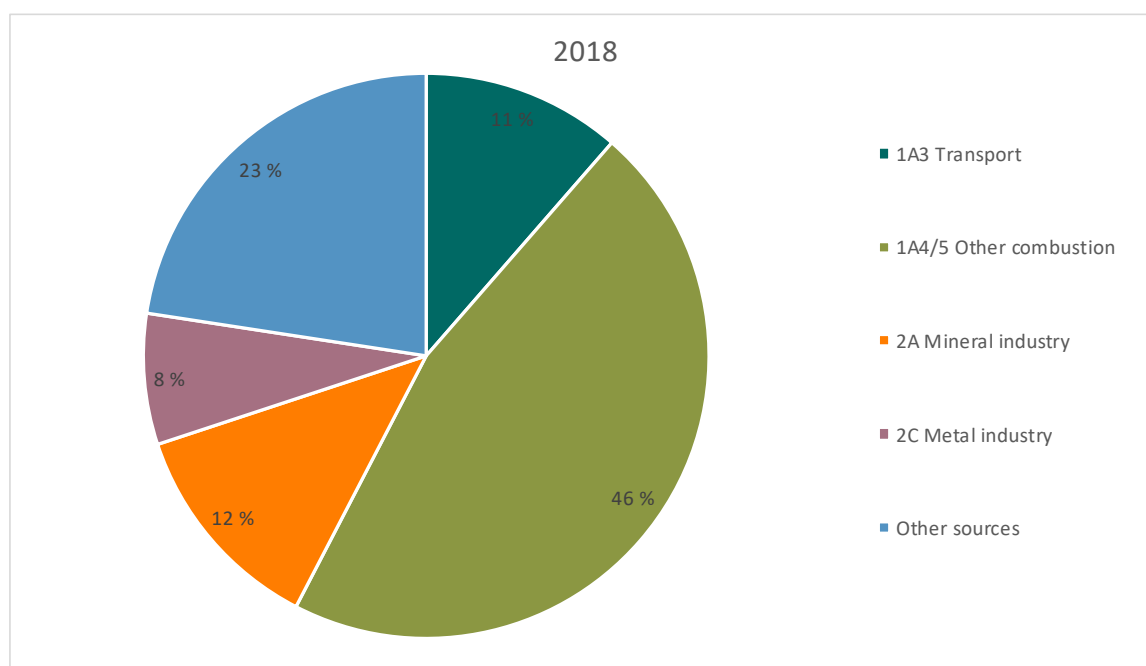


Figure 2.15. Distribution of PM₁₀ emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.2 PM_{2.5}

- 2020 commitment of the revised Gothenburg Protocol: a 30 per cent reduction compared to emissions in 2005, 26 140 tonnes.

Emissions of PM_{2.5} follows the same trend as PM₁₀ emissions. Norway's emissions totaled 24 470 tonnes in 2018. The 2020 commitment of the revised Gothenburg Protocol has been fulfilled since 2014.

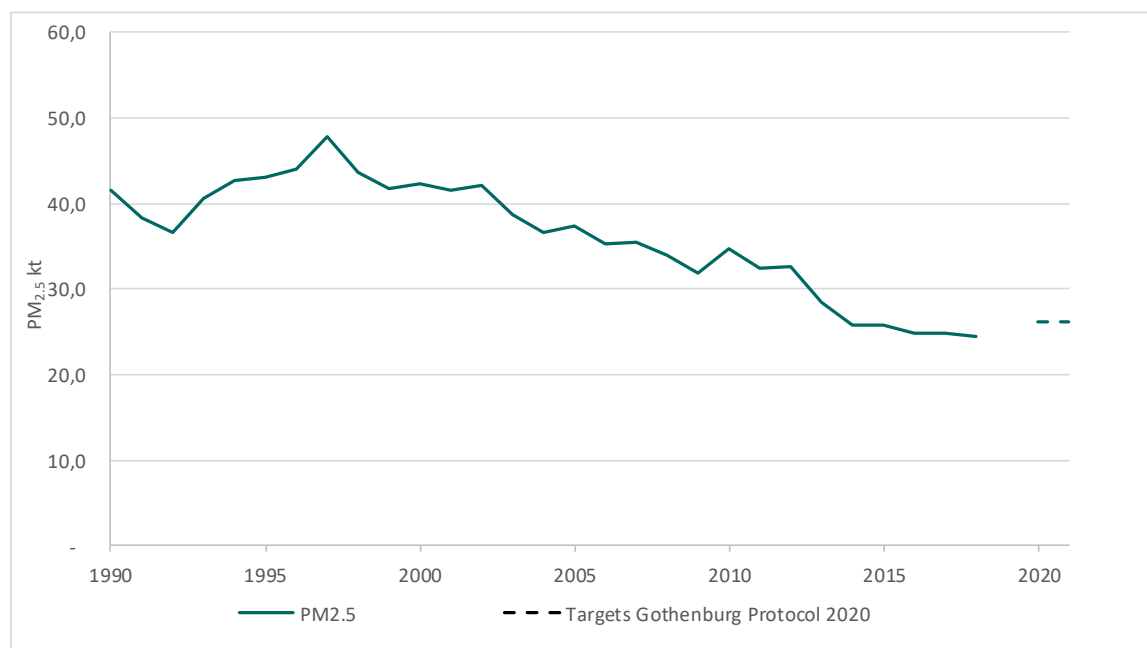


Figure 2.16. Trends in PM_{2.5} emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

2.3.3 Black carbon

The emissions of BC amounted to 2 970 tonnes in 2018, a total reduction of 37 per cent since 1990 and of 2 per cent since 2017.

In 2018, the most important source of emissions was "other combustion" (NFR 1A4 and 1A5), contributing to 40 per cent of the total emissions. From this category, 70 per cent of emissions originated in 2018 from residential stationary plants, primarily due to wood combustion in private households. From 1990 to 2018, emissions from residential stationary plants have been reduced by 24 per cent.

In 2018, the second most important source of emissions was transport. It contributed to 33 per cent of the total BC emissions. The greatest share of emissions within the transport sector stem from navigation, passenger cars and light duty vehicles, contributing to 55, 17 and 12 per cent of the emissions in 2018, respectively. From 1990 to 2018, emissions from navigation have increased by 15 per cent, while emissions from passenger cars have increased by 83 per cent. Emissions from light and heavy duty vehicles have been reduced by 53 and 83 per cent respectively since 1990.

Combustion in energy industries, which in 2018 accounted for 11 per cent of the total BC emissions, have increased by 235 per cent since 1990 due to increased production. The most prominent source of emissions within this category is manufacture of solid fuels and other energy industries. It represented 59 per cent of the sector emissions in 2018. During the period from 1990 to 2018, BC emissions from this subsector increased by 124 per cent.

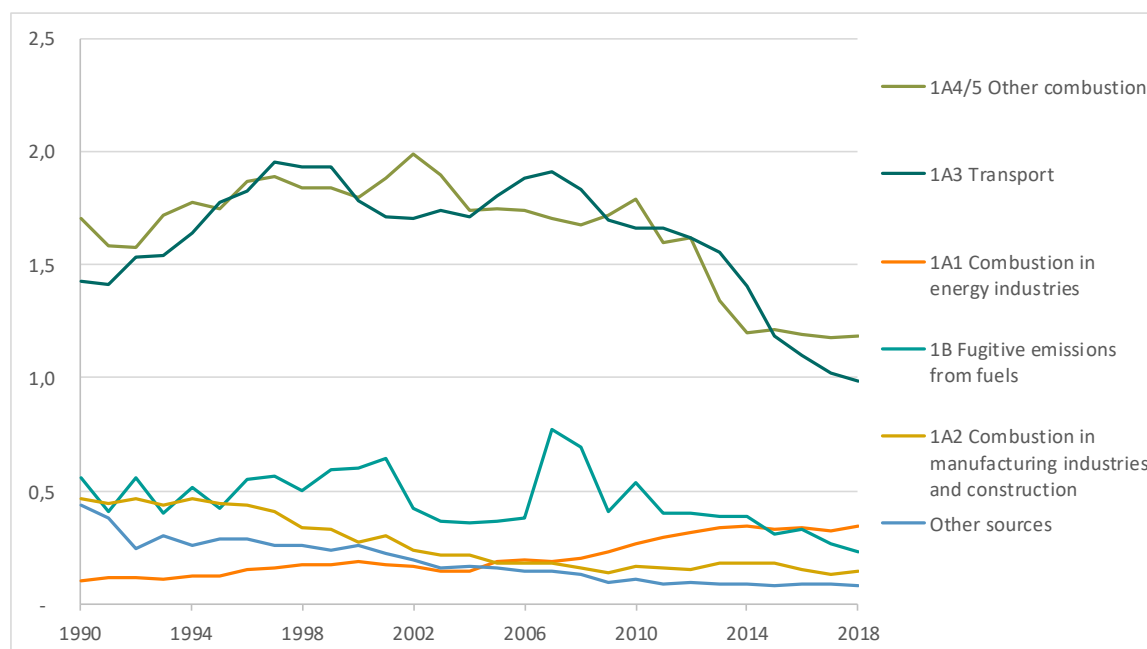


Figure 2.17. Trends in BC emissions, 1990-2018. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

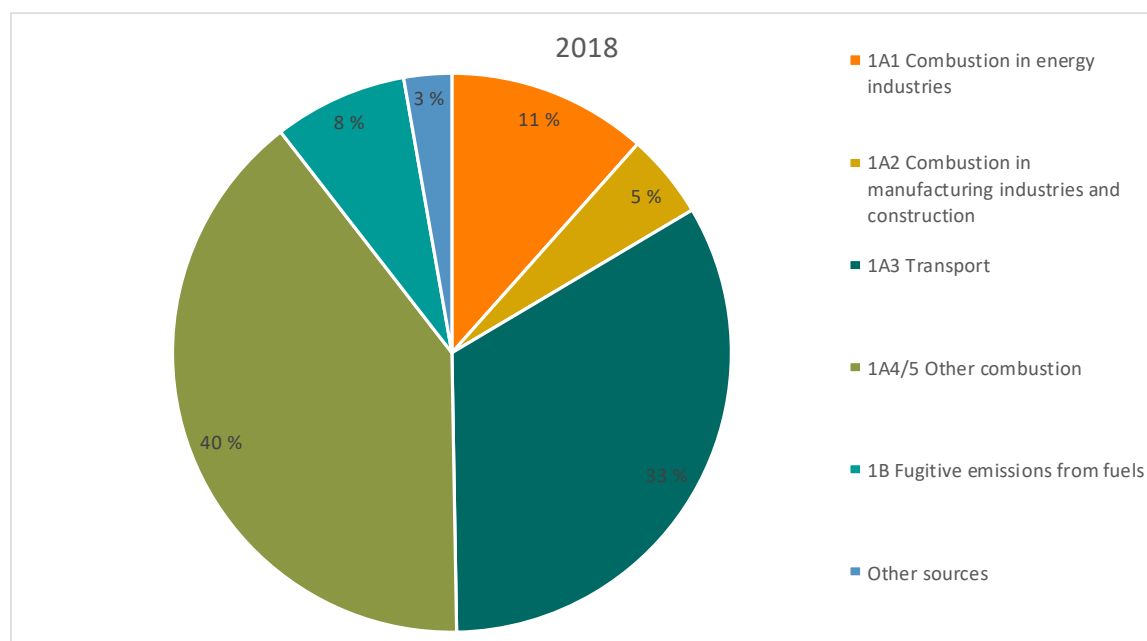


Figure 2.18. Distribution of BC emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.4 Dioxins

In 2018, 19.7 grams of dioxins were emitted. Since 1990, emissions of dioxins have decreased by 84 per cent. A large proportion of the reduction from 1990 is due to the closure of industrial plants and mines. In addition, emissions from energy industries were reduced by 86 per cent from 1990 to 2005 due to the introduction of cleaning measures at waste incineration plants. Since 2005, the emissions in the energy industries have increased, mainly due to increased activity in the oil and gas production.

From 1990 to 1996, the largest source of dioxins emissions was the category other industrial processes (2H) due to an ore mine with high dioxin emissions which was closed down in 1996. Since the closure, dioxins emissions from the source category 2H have been reduced to zero.

In 2018, the largest source of dioxins emissions was the category "other combustion" (NFR 1A4 and 1A5) contributing to 35 per cent. Within this category, stationary plants in the residential sector contributed to 90 per cent to the total dioxins emissions in 2018, primarily due to wood burning. National fishing, which is also included in this category, contributed to 7 per cent to the category.

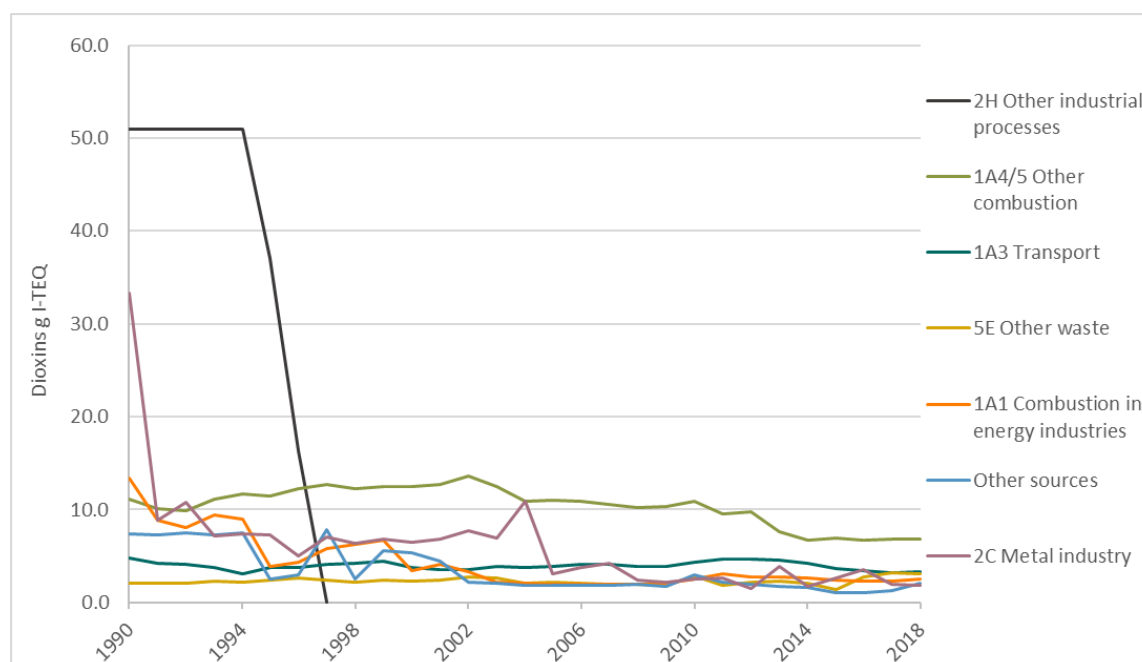


Figure 2.19. Trends in dioxins emissions, 1990-2018. Gram I-TEQ

Source: Statistics Norway/Norwegian Environment Agency

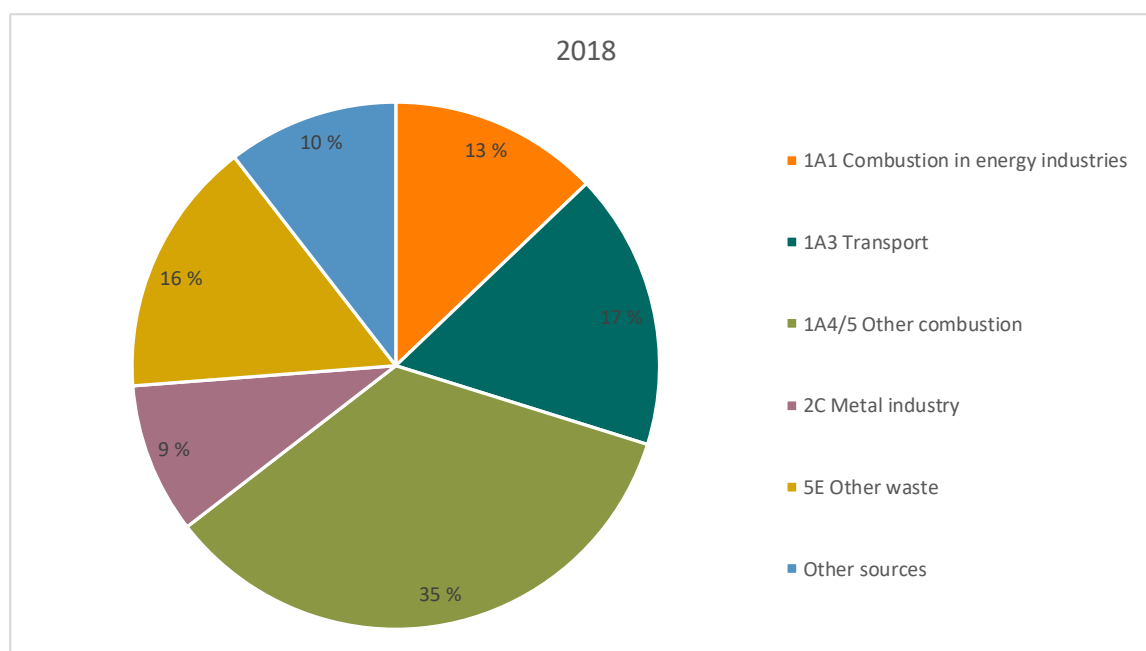


Figure 2.20. Distribution of dioxins emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

Transport was responsible for 17 per cent of total dioxin emissions in 2018. National navigation (shipping) is by far the most significant source of emissions within this category, representing 90 per cent of the transport emissions. Emissions from passenger cars, which contributed to almost half of emissions within this category in 1990, declined rapidly from 1990 to 1994 and further to 1996. Since then, they have been on approximately the same absolute level. In 2018, emissions from passenger cars were 92 per cent lower than in 1990.

Dioxins emissions from combustion in energy industries were responsible for 13 per cent of total emissions of dioxins in 2018. There has been a significant decrease in emissions from public electricity and heat production; emissions in 2018 were 92 per cent lower than in 1990. Emissions from manufacture of solid fuels and other energy industries became the largest source of emissions within this category in 2004 and has remained so since. In 2018, it was responsible for 61 per cent of dioxins emissions within the energy industries category. Emissions from manufacture of solid fuels and other energy industries have increased by 136 per cent since 1990.

Emissions from other waste (5E), which accounted for 16 per cent of the total dioxins emissions in 2018, have increased by 47 percent since 1990. Process emissions from metal production accounted for 9 per cent of the total emissions of dioxins in 2018, and have decreased by 95 per cent since 1990.

2.3.5 PAH-4

The present emission inventory for polycyclic aromatic hydrocarbons (PAH) includes four PAHs: benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene. The total emissions of PAH-4 in 2018 amounted to approximately 6.6 tonnes, which is a reduction of 67 per cent since 1990. Emissions increased by 12 per cent compared with 2017.

In 2018, benzo(b)fluoranthene contributed to 50 per cent of PAH-4 emissions while benzo(a)pyrene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene contributed to 19, 16 and 15 per cent, respectively.

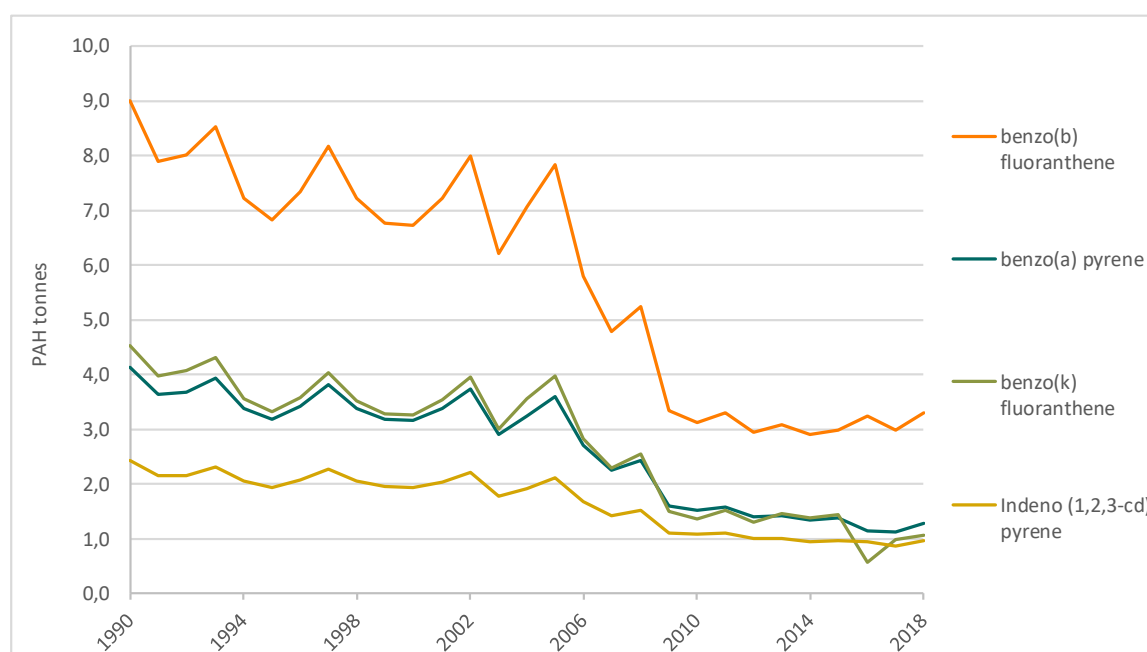


Figure 2.21. Trends in PAH emissions, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene, 1990-2018. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

Process emissions in aluminium production (in 2C Metal industry) is the most dominant source of PAH-4 emissions. It contributed to 71 per cent of the total PAH-4 emissions in 1990 and to 51 per cent in 2018. The PAH-4 emissions decreased primarily because of the discontinuation of Soederberg technology in the aluminium production. Emissions from aluminium production have been reduced by 76 per cent since 1990.

Road transport contributed to 22 per cent of the PAH-4 emissions in 2018. This includes both exhaust and tyre and brake wear. There is a general trend with increasing emissions from transport since 1990, emissions from passenger cars (without tyre and brake wear) have increased by 49 per cent, emissions from light duty vehicles have increased by 252 per cent whilst emissions from heavy duty vehicles have increased by 47 per cent.

Wood burning is by far the most significant source of emissions within the category "other combustion" (NFR 1A4 and 1A5). Emissions from stationary plants in the residential sector

accounted for 18 per cent of PAH-4 emissions in 2018 and have been reduced by 58 per cent since 1990, due to the increasing share of new technology in wood burning appliances.

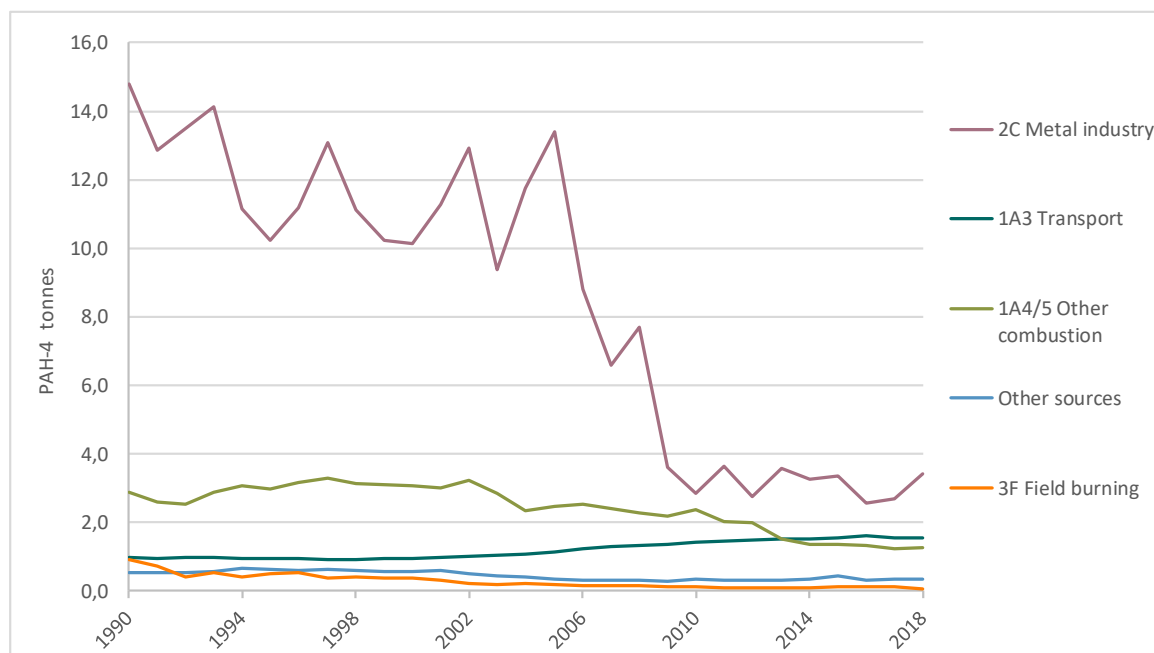


Figure 2.22. Trends in total PAH-4 emissions, 1990-2018. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

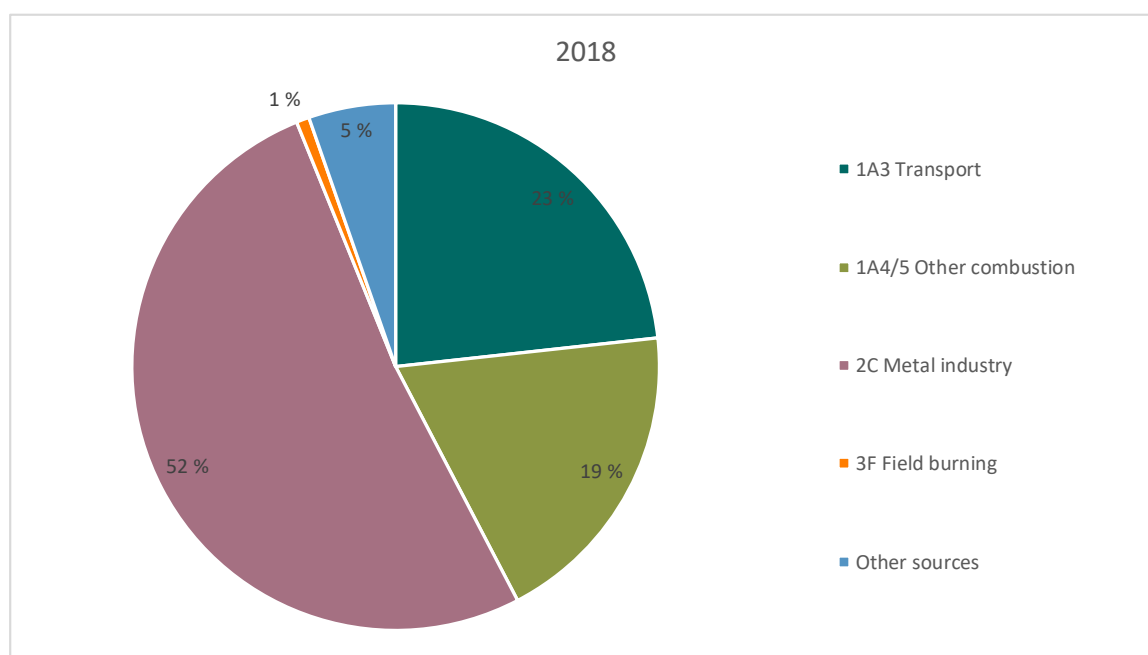


Figure 2.23. Distribution of total PAH-4 emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.6 HCB

Estimated HCB emissions in Norway amounted to 1.3 kilograms in 2018, and has decreased by 99 per cent since 1990. Emissions decreased mainly due to the closure of magnesium production which contributed to almost 99 per cent of total HCB emissions in 1990.

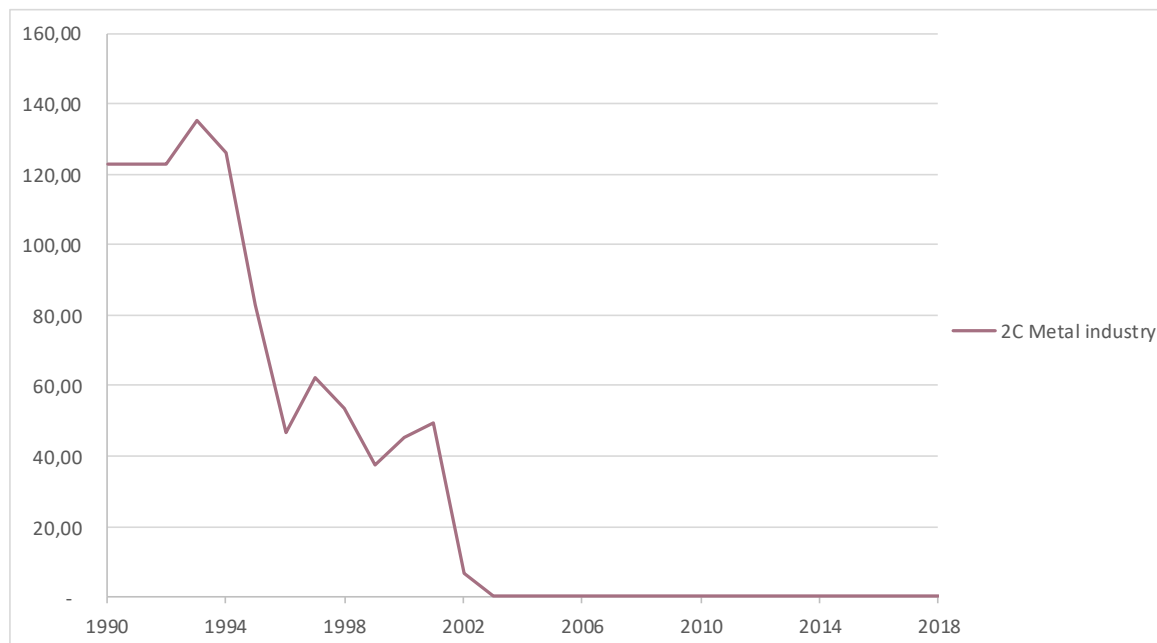


Figure 2.24. Metal industry, trends in HCB emissions, 1990-2018. Kilogram

Source: Statistics Norway/Norwegian Environment Agency

The most important source of emissions of HCB in 2018 was road transport, which contributed to 53 per cent of total emissions. Emissions from road transport have increased significantly since 1990, mainly due to increased traffic activity. For instance, HCB emissions from passenger cars were more than fifteen times higher in 2018 than in 1990.

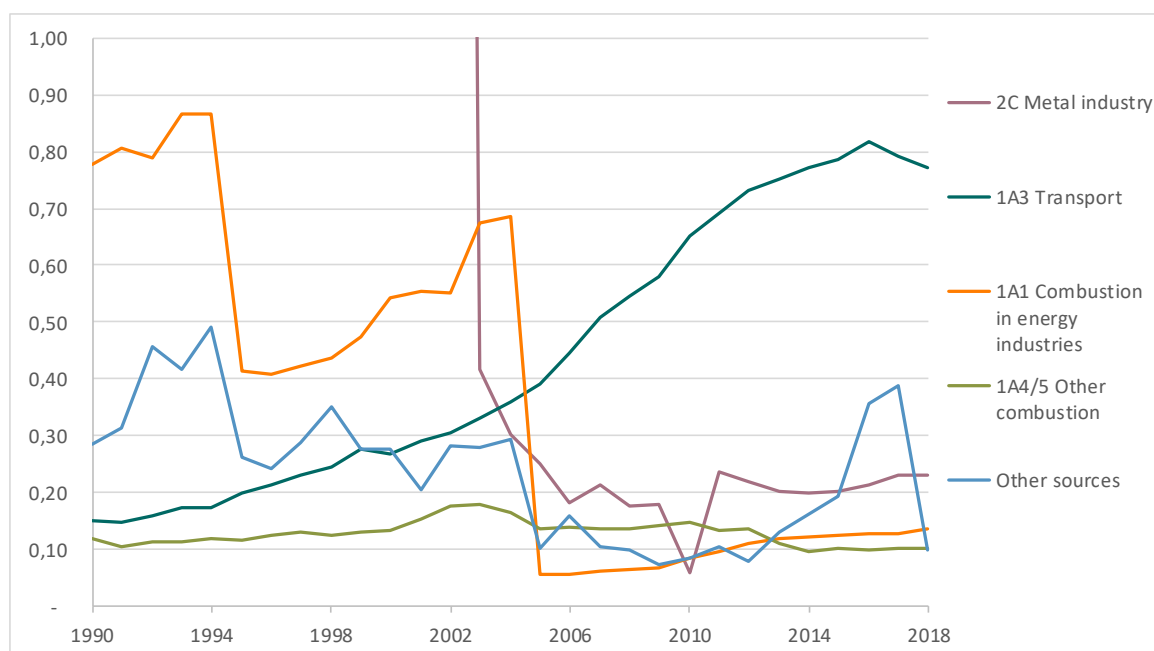


Figure 2.25. Trends in total HCB emissions, 1990-2018. Kilogram

Source: Statistics Norway/Norwegian Environment Agency

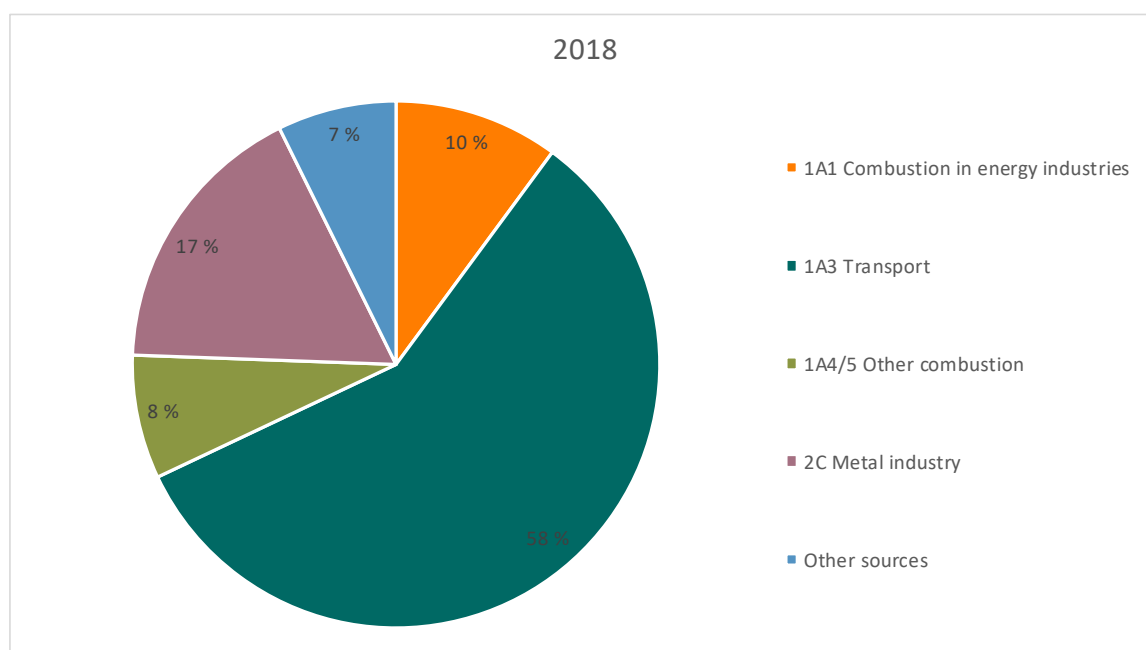


Figure 2.26. Distribution of HCB emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.7 PCB

Estimated PCB emissions in Norway amounted to 24.8 kilograms in 2018. Emissions have decreased by 88 per cent since 1990. Emissions from passenger cars, which accounted for 75 per cent of the total PCB emissions in 1990, decreased from 158 to 1 kilogram from 1990 to 2018.

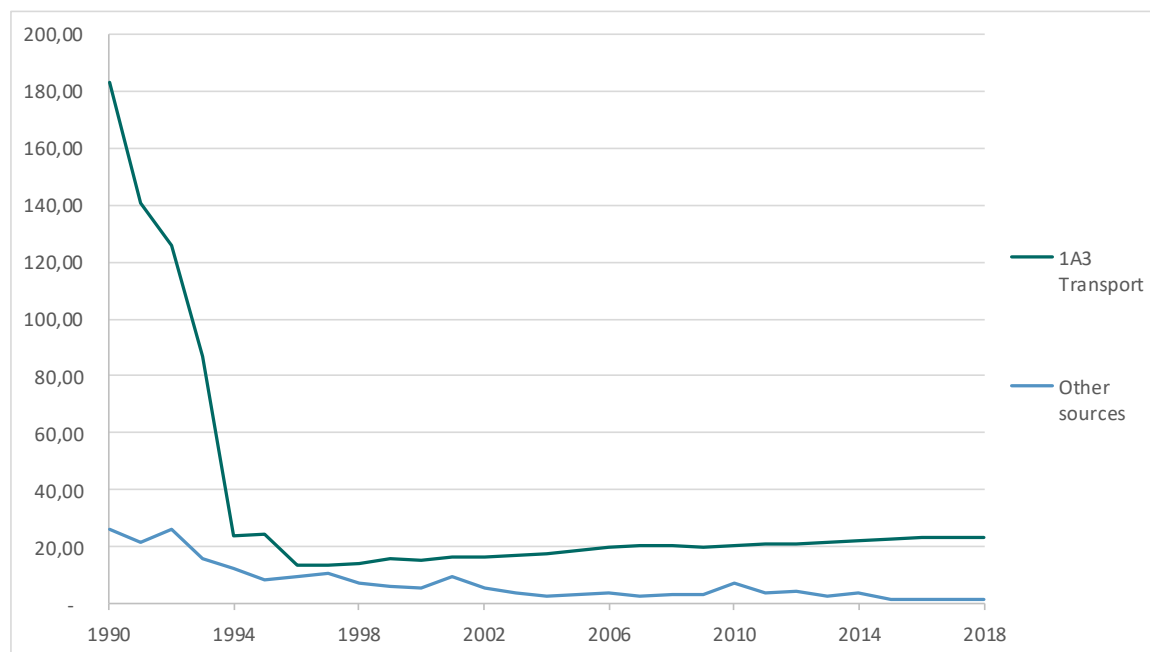


Figure 2.27. Trends in total PCB emissions, 1990-2018. Kilogram

Source: Statistics Norway/Norwegian Environment Agency

Despite large reductions, road transport remained the most important source of emissions of PCB in 2018. It contributed to 92 per cent of total emissions.

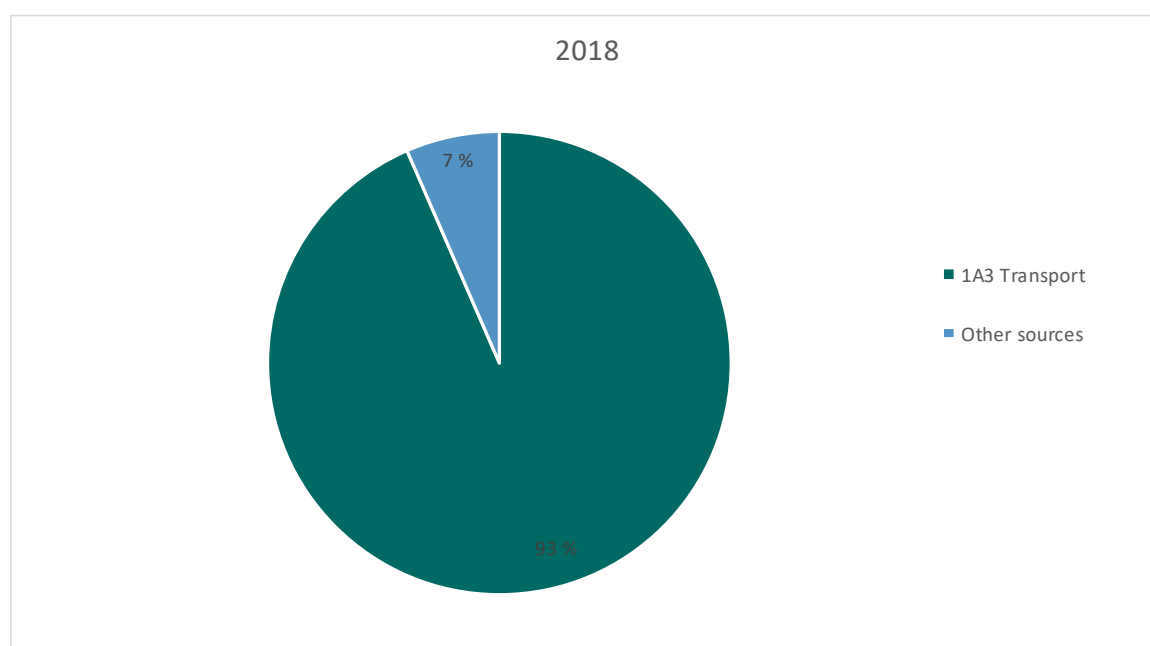


Figure 2.28. Distribution of PCB emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.8 Lead

Lead emissions totaled 5.8 tonnes in 2018 and have been reduced by 97 per cent since 1990. Regulations on lead content in fuels are the main reason for this reduction. Indeed, emissions from passenger cars constituted 77 per cent of the total in 1990, and only 2 per cent in 2018. Emissions of lead have been relatively constant in recent years.

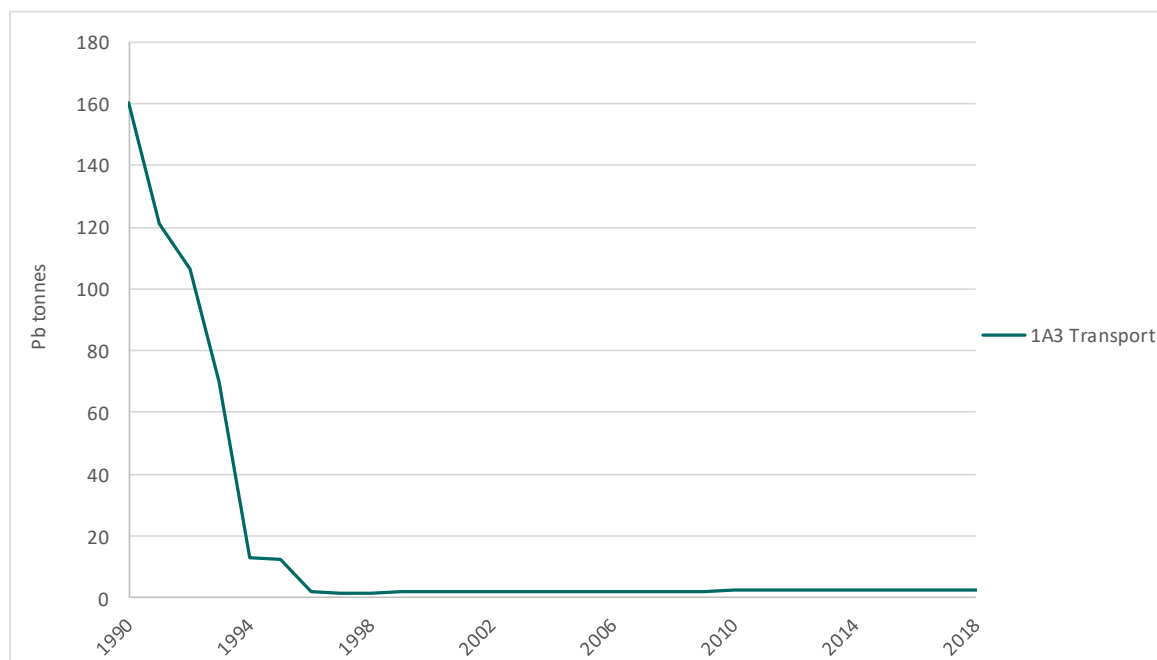


Figure 2.29. Transport, trends in lead emissions, 1990-2018. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

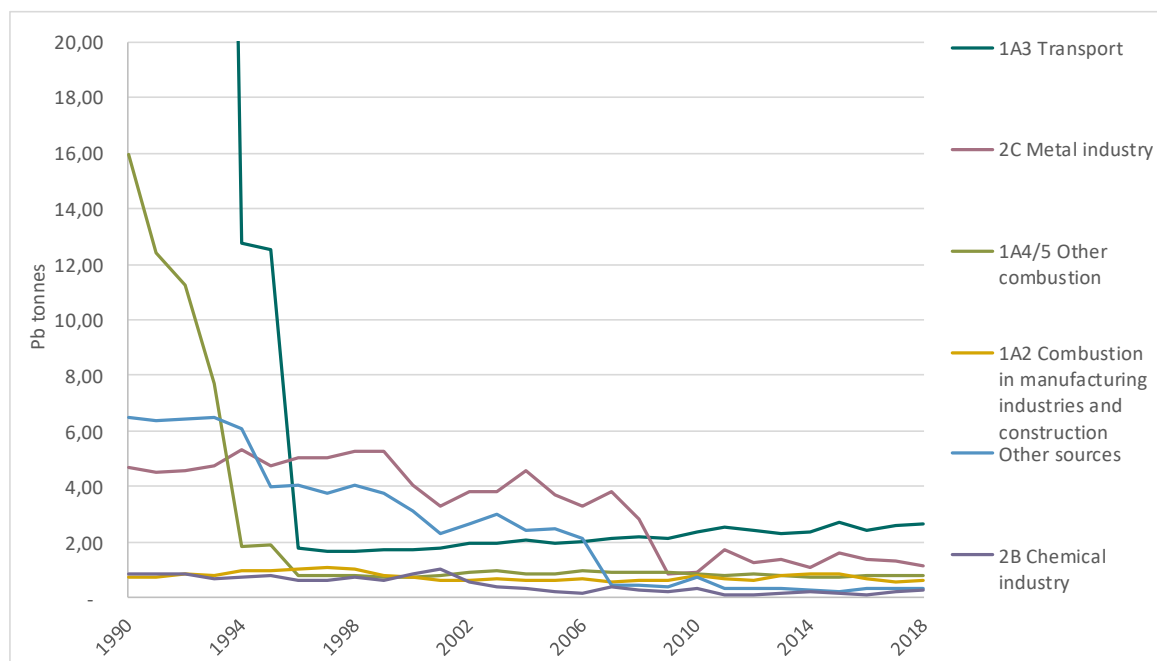


Figure 2.30. Trends in lead emissions, 1990-2018. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

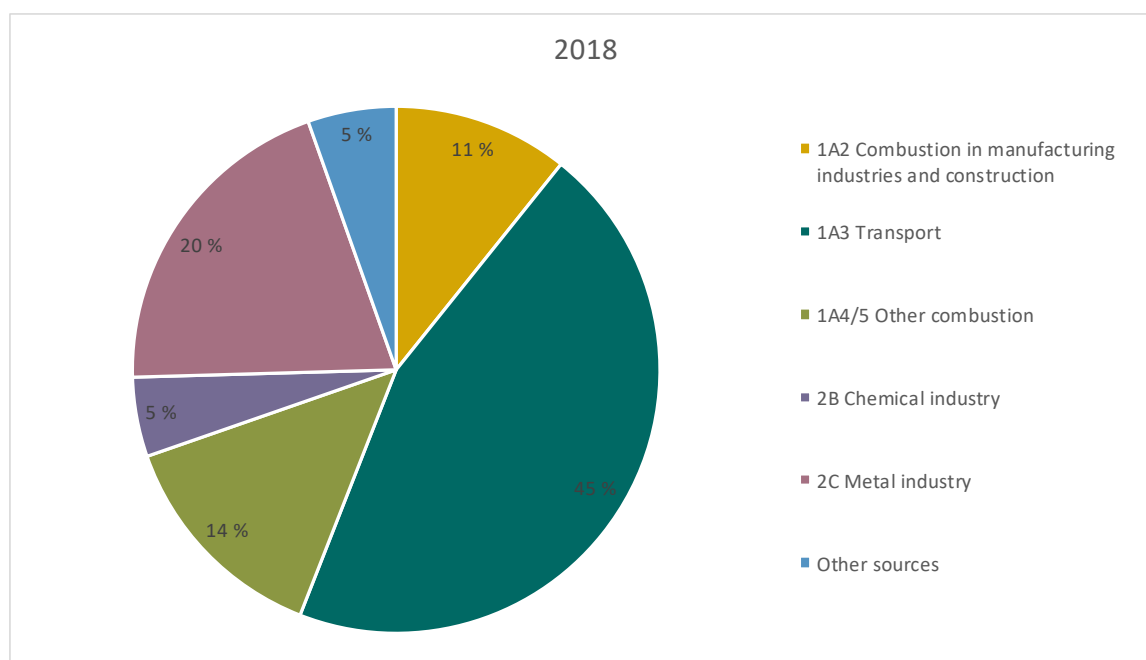


Figure 2.31. Distribution of lead emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

Transport has remained the largest source of lead emissions, but since 1996 tyre and brake wear has been the most important source within the transport sector, being responsible for 28 per cent of lead emissions in 2018.

In 2018, process emissions from metal industry and combustion activities within manufacturing industries and construction constituted 23 and 9 per cent of the total lead emissions, respectively. The category "other combustion" (NFR 1A4 and 1A5) accounted for 14 per cent of total lead emissions in 2018.

2.3.9 Cadmium

Emissions of cadmium totaled 0.5 tonnes in 2018, representing a 4 per cent increase from 2017, and a 69 per cent reduction from 1990. Emissions have primarily been reduced in manufacturing industries and field burning.

Process emissions from production of iron, zinc, steel and ferroalloys have been reduced due to reduction efforts and closing down of production plants. Metal industry was responsible for 20 per cent of cadmium emissions in 2018, compared to 33 per cent in 1990. Cadmium emissions from metal industry increased by 59 per cent from 2017 to 2018.

Cadmium emissions from field burning have been significantly reduced from 1990. In 2018, it contributed to 4 per cent of total Norwegian cadmium emissions, compared to 21 per cent in 1990.

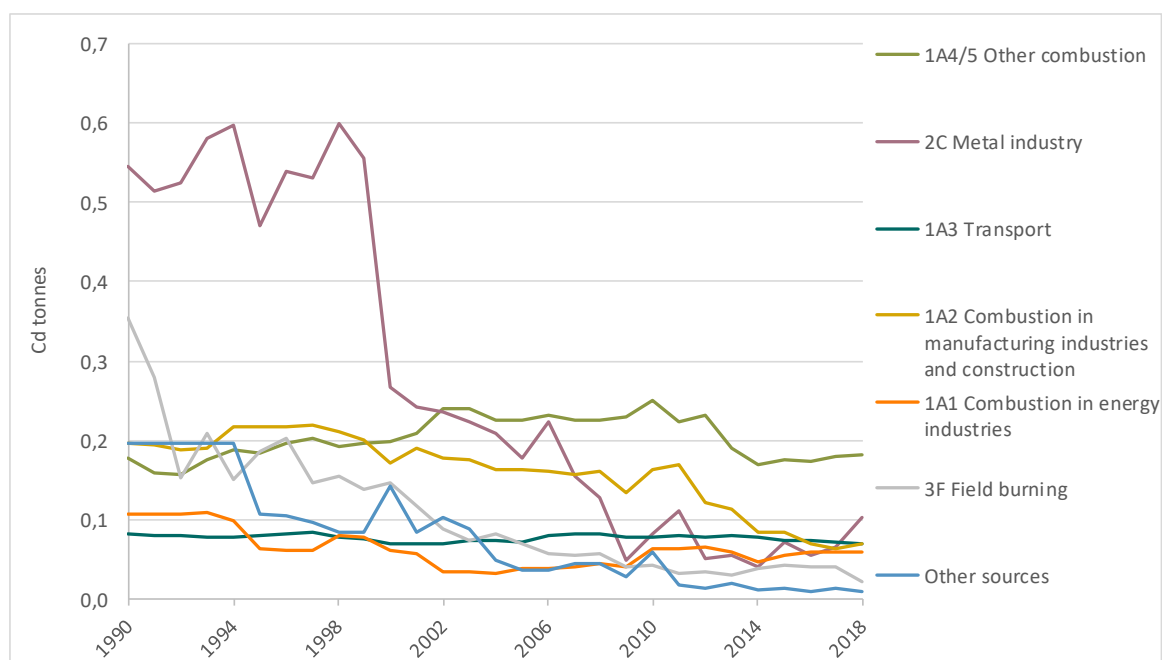


Figure 2.32. Trends in cadmium emissions, 1990-2018. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

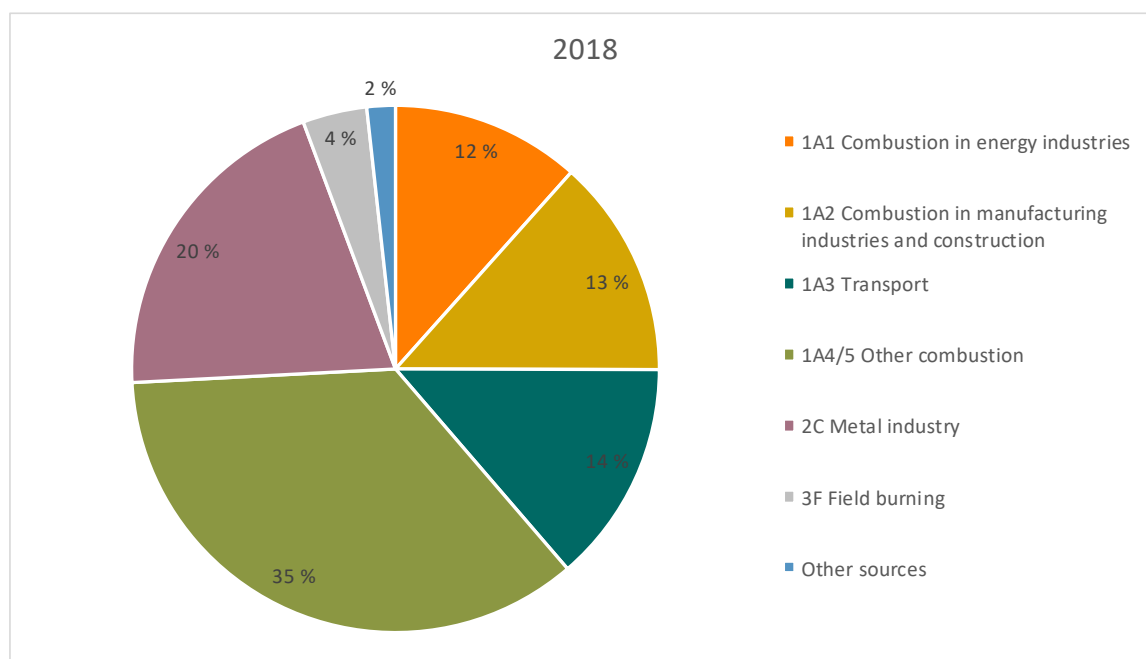


Figure 2.33. Distribution of cadmium emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

In 2018, the largest source of cadmium emissions was the category "other combustion" (NFR 1A4 and 1A5) contributing to 35 per cent. Within this category, stationary plants in the residential sector contributed to 88 per cent of emissions in 2018, primarily due to wood burning. Combustion in energy industries, combustion in manufacturing industries and construction and transport are also large sources of cadmium emissions. In 2018, they contributed to 12, 13 and 14 per cent of the total emissions, respectively.

2.3.10 Mercury

Emissions of mercury amounted to 0.2 tonnes in 2018, which is a 85 per cent reduction from 1990. The decrease is mainly due to reductions within metal industry sector (2C) and use of tobacco included in "Other product use" (NRF 2G). These sectors contributed, respectively, to 42 and 20 per cent of total mercury emissions in 1990, and have been reduced by 94 and 97 per cent, respectively, since then.

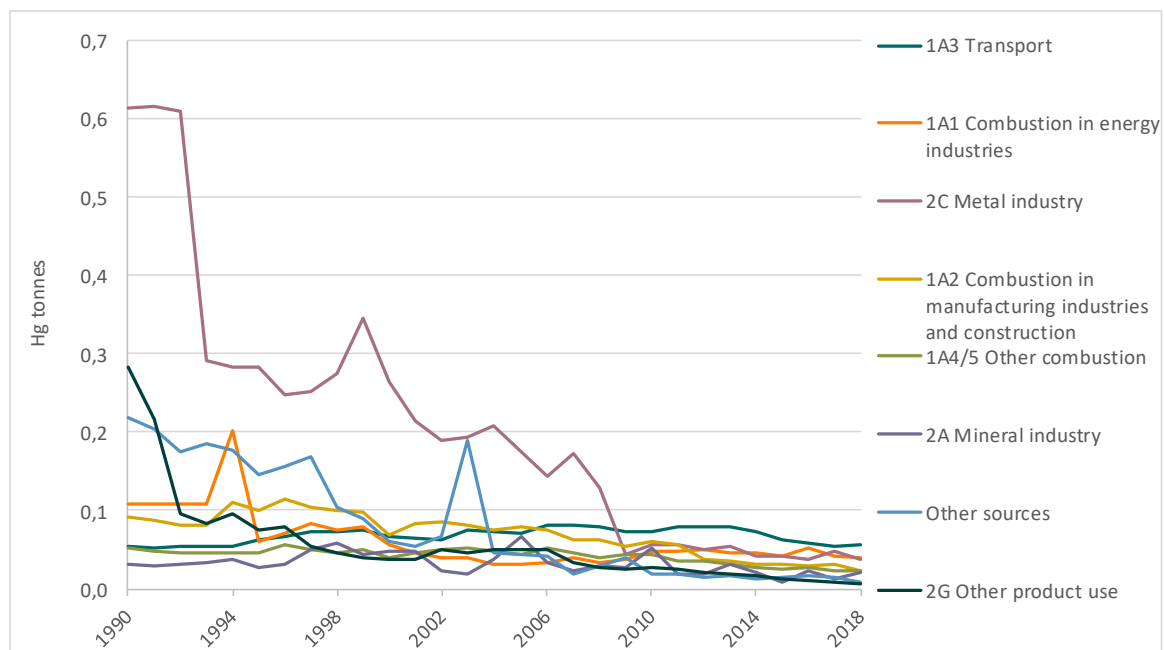


Figure 2.34. Trends in mercury emissions, 1990-2018. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

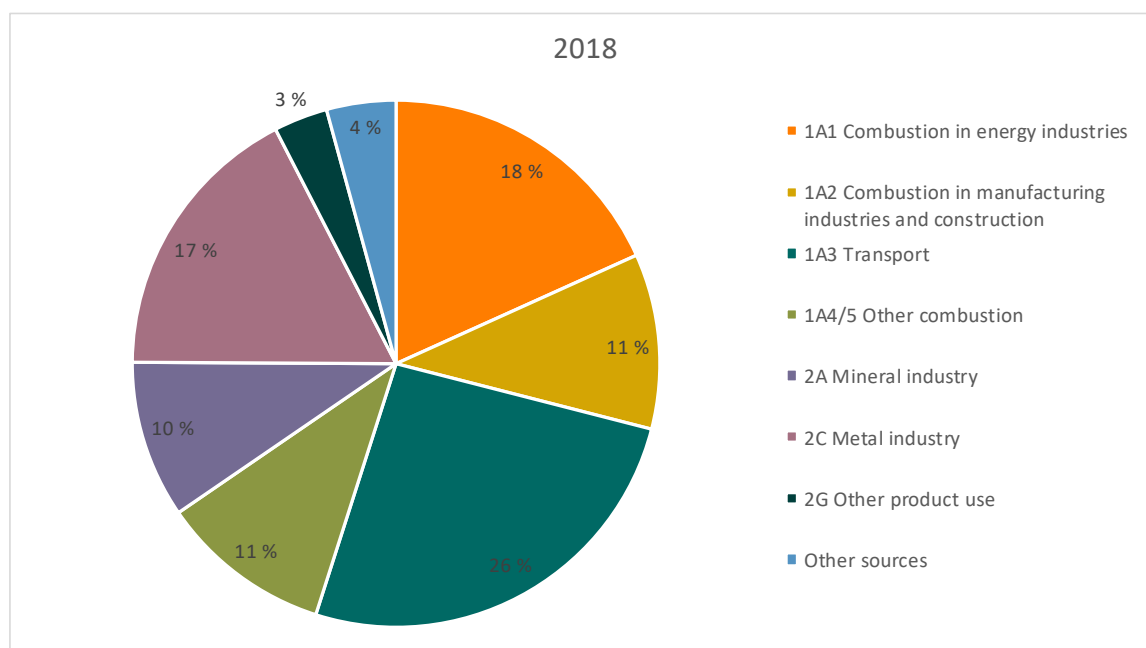


Figure 2.35. Distribution of mercury emissions between emission sources, 2018. Per cent

Source: Statistics Norway/Norwegian Environment Agency

Mercury emissions originate from a wide range of sources. The most important source of mercury emissions in 2018 were the transport sector. The emissions from this source have remained at the same level since 1990.

2.3.11 Copper, chromium, and arsenic

Emissions of copper were 27 tonnes in 2018, an increase of 14 per cent since 1990. Tyre and brake wear within road transport is the most dominant source of emissions of copper.

Emissions of chromium amounted to 3.7 tonnes in 2018 and have been relatively constant in recent years. Emissions have been reduced by 68 per cent since 1990.

In 2018, 1.3 tonnes of arsenic were emitted, an increase of 14 per cent from 2017 and a reduction of 62 per cent since 1990. For the past few years, the variation in arsenic emissions is due to varying arsenic content in raw materials and reducing agents used in metal production.

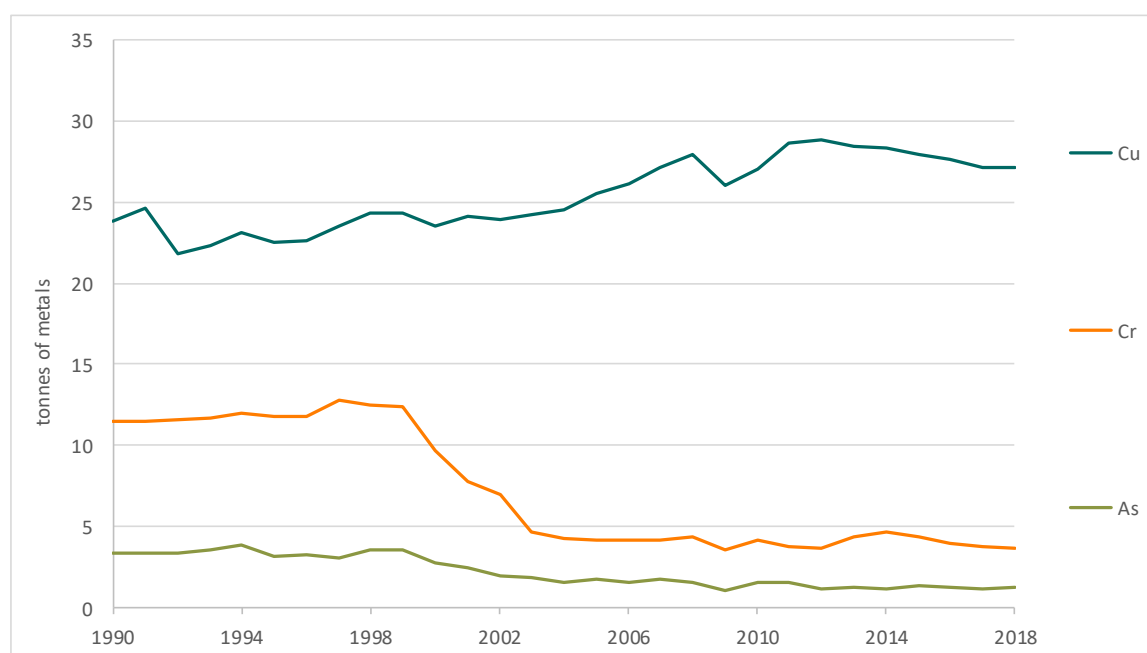


Figure 2.36. Trends in copper, chromium and arsenic emissions, 1990-2018. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

3 ENERGY (NFR sector 1)

3.1 Overview

This chapter provides descriptions of methodologies used to calculate emissions from the energy sector. The disposition of the chapter follows the NFR classifications of the emission sources. In section 3.2, emission estimations from energy combustion are described. This includes combustion emissions from energy industries, manufacturing industries and construction, transport, and other combustion sources. Section 3.2 also includes memo items about international bunker fuels.

In section 3.3, a description is given for fugitive emissions from fuels. This includes fugitive emissions from coal mining and handling, and from oil and natural gas.

3.2 Energy combustion

NFR 1A

Last update: 01.03.19

3.2.1 Overview

Combustion of fossil fuels and biomass leads to emissions of SO₂, NO_x, NMVOC, CO, particulate matter, heavy metals, PAH, dioxins and NH₃.

Table 3.1. Energy combustion emissions as per cent of total emissions, 2018

| Pollutant | Per cent of emissions |
|------------------|-----------------------|
| SO ₂ | 36 |
| NO _x | 88 |
| NMVOC | 29 |
| CO | 62 |
| NH ₃ | 2 |
| PM ₁₀ | 68 |
| BC | 89 |

Source: Statistics Norway/Norwegian Environment Agency

The emissions of SO₂, NMVOC and CO from the energy sector have been significantly reduced since 1990. The reduction of SO₂ emissions has taken place in all sectors due to reduced sulphur content in fuels. NMVOC and CO emissions have been reduced mainly due to reductions in emissions from petrol passenger cars.

Emissions of NO_x and particles were stable in the 1990s. Particles have been reduced since 2002 while most of NO_x reductions happened after 2007.

Catalysts in petrol passenger cars cause NH₃ emissions.

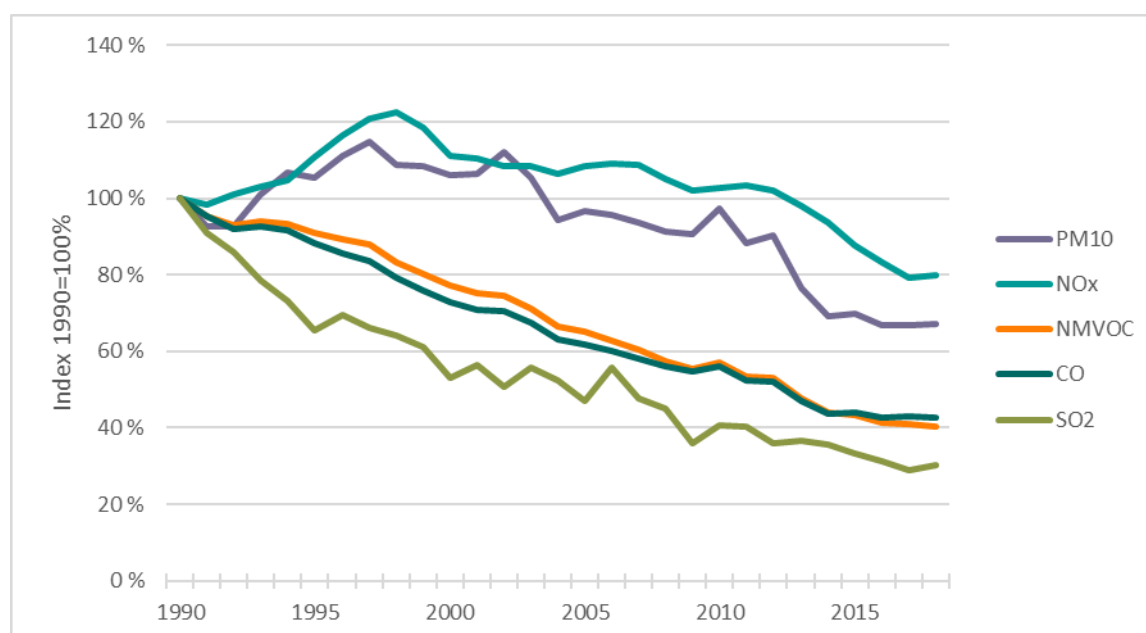


Figure 3.1. Trends for the emissions for most of the long-range transboundary air pollutants from energy combustion. Index 1990 = 100%

Source: Statistics Norway/Norwegian Environment Agency

Emissions from energy combustion include contributions from all sources addressed in the UNECE Guidelines. Emissions from waste incineration at district heating plants are accounted for under the energy sector, as the energy is utilised. Emissions from flaring in the energy sectors are described in section 3.3. Coal and coke used as reducing agents and gas used for production of ammonia (non-energy part) are accounted for under industrial processes (section 4). Flaring of natural gas and fuel gas in chemical industry is reported in section 4.3. Other flaring outside the energy sectors is described in section 6.4. The same applies to emissions from accidental fires etc. Emissions from burning of crop residues and agricultural waste are accounted for and described in section 5.6. Emissions from tobacco are described in section 4.5.6.

The main source for calculation of emissions from energy combustion is the energy balance, which is prepared annually by Statistics Norway. In 2018, a revised version of the energy balance for 1990-2017 was published. The technical system was upgraded, and several methodological changes and new input data were introduced. See section 8.1 Recalculations.

Many different sources are utilised in the preparation of the energy balance. E.g. energy use in extraction of oil and gas, which constitutes an important part of Norwegian energy use, is reported from by the companies to the Norwegian Petroleum Directorate. Other energy producers, such as oil refineries and district heating plants, also report their own energy use to Statistics Norway and the Norwegian Environment Agency.

For different oil products, the total frame for annual use is given by Statistics Norway's statistics on deliveries of petroleum products. These statistics are also used in the estimation of use in different economic sectors, together with other available information. The distribution between

sectors is of varying quality; in some cases surveys from earlier years are used to estimate current distribution of consumption between sectors. For manufacturing industries, however, Statistics Norway's annual survey on all types of energy use, based on reports from plants that are responsible for approximately 96 per cent of the energy use in these sectors, combined with estimations for the remaining plants, provide figures of high quality.

3.2.1.1 Method

3.2.1.1.1 General

Emissions from energy combustion are estimated at the sectoral level in accordance with the IPCC sectoral approach Tier 2/Tier 3. Total fuel consumption is often better known than the sectoral consumption.

The general method to estimate emissions from fuel combustion is multiplication of fuel consumption by source and sector by an appropriate emission factor. Exceptions are road and air transport where more detailed estimation models are used, involving additional activity data (see section 3.2.4.2 and 3.2.4.1 respectively). Fuel consumption figures are taken from the Norwegian energy balance. The mean theoretical energy content of fuels and their densities are listed in Table 3.2.

Table 3.2 Average energy content (NCV) and density of fuels*

| Energy product | Theoretical energy content | Density |
|------------------------------------|----------------------------|----------------------|
| | GJ/tonne | Tonne/m ³ |
| Coal | 28.1 | : |
| Coke | 28.5 | : |
| Petrol coke | 35 | : |
| Crude oil | 42.3 | 0.85 |
| Motor gasoline | 43.9 | 0.74 |
| Aviation gasoline | 43.9 | 0.74 |
| Kerosene (heating) | 43.1 | 0.81 |
| Jet kerosene | 43.1 | 0.81 |
| Auto diesel | 43.1 | 0.84 |
| Marine gas oil/diesel | 43.1 | 0.84 |
| Light fuel oils | 43.1 | 0.84 |
| Heavy distillate | 43.1 | 0.88 |
| Heavy fuel oil | 40.6 | 0.98 |
| Natural gas (dry gas) (land) | 47.97 | 0.74 ¹ |
| Natural gas (rich gas) (off shore) | 47.41 | 0.85 ¹ |
| LPG | 46.1 | 0.53 |
| Refinery gas | 48.6 | : |
| Blast furnace gas | : | : |
| Fuel gas ² | 50 | : |
| Landfill gas ³ | 50.4 | : |
| Biogas ³ | 50.4 | : |
| Fuel wood | 16.8 | 0.5 |
| Ethanol | 26.8 | 0.79 |
| Biodiesel | 36.8 | 0.88 |
| Wood waste | 16.8 | : |
| Black liquor | 7.2 - 9.2 | : |
| Charcoal | 29.5 | : |
| Municipal waste | 11.5 | : |
| Special waste | 40.6 | 0.98 |

*The theoretical energy content of a particular energy commodity may vary; Figures indicate mean values. All data are net calorific value (NCV)

¹ kg/Sm³ = standard cubic metre (at 15 °C and 1 atmospheric pressure)

² In this inventory, fuel gas is a hydrogen-rich excess gas from petrochemical industry

³ Landfill gas and other types of biogas are reported as methane content in the energy balance

Source: *Energy statistics, Statistics Norway*

Handbook of Emission Factors (HBEFA; (INFRAS 2010)) describes methodologies used for road traffic. Several documentation reports have been published describing the methodologies used for road traffic (Holmengen & Fedoryshyn 2015) and navigation (Tornsjø 2001, Flugsrud *et al.* 2010). The methodology for aviation is described in an internal document from Statistics Norway (Skullerud 2014).

3.2.1.1.2 *Delimitation towards industrial processes etc.*

The energy combustion sector borders to several other source categories. This section presents the demarcation with other sectors used in the inventory.

Energy consumption reported as activity data in the emission inventories is generally delimited in the same way as emissions. In cases where different substances are handled differently, the delimitation of energy consumption follows the delimitation of CO₂ emissions.

Flaring is not reported as energy use in 1A. Instead, flaring is reported in the following source categories:

- Flaring in refineries and in exploration/extraction is reported in 1B – Fugitive emissions.
- Flaring in manufacturing industries is reported in 2 – Industrial processes, particularly in 2B – Chemical industry. (In the energy balance, flaring in manufacturing is reported as "losses".)
- Flaring of landfill gas is reported in 6C – Waste incineration.

Combustion of *solid waste* and *hazardous waste* is reported in the energy section (district heating in 1A1a and in several manufacturing industries). No significant combustion of solid or hazardous waste occurs without energy recovery.

Combustion of *landfill gas* with energy recovery is reported in the energy section (mainly in 1A4a Commercial/Institutional). Flaring is reported in 6C waste incineration, as mentioned above.

Some special problems relating to allocation of reported total plant emissions are discussed in section 3.2.1.1.4.

3.2.1.1.3 *Emissions reported by plants: overview*

For some major manufacturing plants (in particular offshore activities, refineries, gas terminals, cement industry, production of plastics, ammonia production), emissions of one or more compounds, reported to the Norwegian Environment Agency from the plants, are used instead of figures calculated with general emission factors as described above. In these cases, the energy consumption at the plants in question is subtracted from the total energy use before the general method is used to calculate the remaining emissions of the compound in question, in order to prevent double counting.

Emissions are reported to the Norwegian Environment Agency under a number of different reporting obligations.

In the general equation (2.2), *Emissions* (E) = $[(A - A_{PS}) \cdot EF] + E_{PS}$, E_{PS} represents the reported emission data, while A_{PS} represents the energy consumption at the plants. Note that for most plants, reported emissions are used only for a limited number of compounds. For the remaining substances in the inventory, the general method with standard emission factors is used.

Reported figures are used for a relatively small number of plants, but as these contribute to a large share of the total energy use, a major part of the total emissions are based on such

reported figures. For the source categories petroleum refining, manufacture of solid fuels and other energy industries and iron and steel, more than 90 per cent of the sector emissions are based on reported data from plants. The reports are from the mandatory reporting obligation that is a part of the plants' permits given by the authorities.

3.2.1.1.4 *Emissions reported by plants: Energy data*

Energy data for plants with reported emissions (A_{PS} in equation (2.2)) should be consistent both with the energy balance that is used for activity totals A and with the reported emission data. Consistency with emission data means that the energy data should correspond to the same activity as the reported emissions.

In most cases, figures on plant energy use in the inventory are based on data reported from the plants to Statistics Norway. This ensures consistency with the energy balance. However, for some plants, some of the energy data may differ between reports to Statistics Norway and data reported together with emissions to the Norwegian Environment Agency. In some cases, this may lead to problems with consistency.

3.2.1.1.5 *Emissions reported by plants: Allocation to combustion/ processes*

In some cases, emissions are reported as a plant total which includes both combustion and process emissions. All emissions of particulates, heavy metals and POPs are entered into the inventory as process emissions. Emissions from combustion are set to 0 in order to avoid double counting.

3.2.1.1.6 *Emissions reported by plants: Allocation to fuels*

For some plants and substances, emissions are reported by fuel, but in most cases reported combustion emissions are entered as a plant total. The emissions are then allocated to fuels bases on standard EFs using equation 3.1:

$$(3.1) \quad E_{PS,f} = E_{PS} \cdot A_{PS,f} \cdot EF_f / \sum_f (E_{PS} \cdot EF_f)$$

where the subscript f denotes fuel type.

This means that any deviations in data will be distributed across all fuels at the plant. Typical situations include:

- Plants with atypical fuels which differ from standard emission factors.
- Plants with errors or other inconsistencies in energy data.

In such cases, implied emission factors may deviate from the standard range also for other fuels than the one which is really affected.

Plants/substances which are entered by fuel currently include among others:

- Particulate matter from manufacturing of wood products.
- Heavy metal and POP emissions from combustion of municipal solid waste and special waste.

3.2.1.2 Activity data

The annual energy balance, compiled by Statistics Norway, forms the framework for the calculation of emissions from energy use. The energy balance defines the total energy consumption for which emissions are accounted. However, a large part of the total emissions are based on reports from plants that use much energy, i.e. offshore activities and energy-intensive industries on shore. Energy consumption in these plants is included in the energy balance, but this consumption is subtracted before the remaining emissions are calculated by the standard method of multiplying energy use by emission factors. Energy figures reported from the plants to Statistics Norway, which are used in the energy balance, sometimes deviate from the energy figures used to estimate reported emission figures, and this may cause inaccuracies in implied emission factors.

The energy balance surveys the flow of the different energy carriers within Norwegian territory. It includes energy carriers used as raw materials and reducing agents, but these are presented in a separate item and are not included in the data used to estimate emissions from combustion. Some emissions vary with the combustion technology; a distribution between different sources is thus required. Total use of the different oil products is based on the Norwegian sales statistics for petroleum products. For other energy carriers, the total use of each energy carrier is determined by summing up reported/estimated consumption in the different sectors. A short summary of the determination of amounts used of the main groups of energy carriers and the distribution between emission sources is given below.

3.2.1.2.1 *Natural gas*

Most of the combustion of natural gas is related to extraction of oil and gas on the Norwegian continental shelf. The amounts of gas combusted, distributed between gas turbines and flaring, are reported annually to Statistics Norway by the Norwegian Petroleum Directorate (NPD). These figures include natural gas combusted in gas turbines on the various oil and gas fields as well as on Norway's four gas terminals on shore. Statistics Norway's annual survey on energy use in manufacturing industries and sales figures from distributors give the remainder. Some manufacturing industries use natural gas in direct-fired furnaces; the rest is burned in boilers and, in some cases, flared.

3.2.1.2.2 *LPG and other gases*

Consumption of *LPG* in manufacturing industries is reported by the plants to Statistics Norway in the annual survey on energy use. Figures on use of *LPG* in households are based on sales figures, collected annually from the oil companies. Use in agriculture and construction are not based on annual surveys. The figure for agriculture is interpolated for years not included in surveys, whereas the figure for construction is adjusted annually, based on employment figures.

Use of *refinery gas* is reported to Statistics Norway from the refineries. The distribution between

the sources direct-fired furnaces, flaring and boilers is based on information collected from the refineries in the early 1990's. However, the total emissions from the refineries included in the inventory are equal to emissions reported from the plants and are regarded being of high quality.

At some industrial plants, excess gas from chemical and metallurgical industrial processes is burned, partly in direct-fired furnaces and partly in boilers. These amounts of gases are reported to Statistics Norway. A petrochemical plant generates *fuel gas* derived from ethane and LPG. Most of the gas is burned on-site, but fuel gas is also sold to several other plants. All use of fuel gas is reported as energy consumption in the inventory.

Several metallurgical plants generate *CO-rich excess gas* that is either burnt on-site or sold to adjacent plants. Two ferroalloy plants sell parts of their CO gas to some other plants (one producer of ammonia, a district heating plant, iron and steel producers and mineral industry), where it is combusted for energy purposes. These amounts are reported as energy consumption.

One sewage treatment plant utilizes biogas extracted at the plant, and reports quantities combusted (in turbines). Emissions are estimated by Statistics Norway, using the same emission factors as for combustion of natural gas in turbines.

3.2.1.2.3 Oil products

Total use of the different oil products is based on Statistics Norway's annual sales statistics for petroleum products. The data are generally considered reliable (with some reservations which are accounted for further down in this section), since all major oil companies selling oil products report to these statistics and have an interest in the quality of the data. The statistics are corrected for direct import by other importers or companies. The use of sales statistics provides a total for the use of oil products. The use in the different sectors must sum up to this total. This is not the case for the other energy carriers. The method used for oil products defines use as identical to sales; in practice, there will be annual changes in consumer stocks, which are not accounted for.

However, since the late 1990s the distribution in the sales statistics between different middle distillates has not been in accordance with the bottom-up estimated consumption of the products. In particular, the registered sales of light fuel oil have generally been too low, and it is known that some auto diesel also is used for heating. In order to balance the accounts for the different products, it has since 1998 been necessary to transfer some amounts between products instead of using the sales figures directly. The most important transfer is from auto diesel to light fuel oil, but in addition some auto diesel has also been transferred to heavy distillate.

Due to inaccuracies in the reporting of sales of marine gas oil from approximately 2005, there is also some uncertainty connected to the distribution between domestic and international sea transport for the latest years.

Stationary use takes place in boilers and, in some manufacturing industries, in direct-fired furnaces. There is also some combustion in small ovens, mainly in private households. Mobile combustion is distributed between a number of different sources, described in more detail in section 3.2.4 Transport. In addition to oil products included in the sales statistics, figures on use of waste oil are given in Statistics Norway's statistics on energy use in the manufacturing industries. Statistics Norway also collects additional information directly from a few companies on the use of waste oil as a fuel source.

Generally, in Norway there is a continuous shift between use of oil and hydroelectricity, corresponding to changes in prices. Between years, this may cause changes in use of oil products and corresponding emissions which can be considerable.

3.2.1.2.4 *Coal, coke and petrol coke*

Use of coal, coke and petrol coke in manufacturing industries is reported annually from the plants to Statistics Norway. The statistics cover all main consumers and are of high quality. Combustion takes place partly in direct-fired furnaces, partly in boilers. Figures on some minor quantities burned in small ovens in private households are based on sales figures. In addition, the figure on an insignificant use of coal in the agricultural sector was formerly collected from one farmer. Since 2002, there has been no use of coal in Norwegian agriculture.

3.2.1.2.5 *Biofuels*

Use of wood waste and black liquor in manufacturing industries is taken from Statistics Norway's annual survey on energy use in these sectors. Use of wood in households is based on figures on the amount of wood burned from the annual survey on consumer expenditure for the years before 2005 and for 2012. The statistics cover purchase in physical units and estimates for self-harvest. The survey figures refer to quantities *acquired*, which not necessarily correspond to *use*. The survey gathers monthly data that cover the preceding twelve months; the figure used in the emission calculations (taken from the energy balance), is the average of the survey figures from the year in question and the following year. For the years 2005-2011, the figures are based on responses to questions relating to wood burning in Statistics Norway's Travel and Holiday Survey. The figures in this survey refer to quantities of wood *used*. The survey quarterly gathers data that cover the preceding twelve months. The figure used in the emission calculations is the average of five quarterly surveys. Figures on some minor use in agriculture and in construction are derived from earlier surveys for these sectors. Combustion takes place in boilers and in small ovens in private households. Consumption figures for wood pellets and wood briquettes are estimates, based on annual information from producers and distributors. Data on use of peat for energy purposes is not available, but according to the Energy Farm, the centre for Bioenergy in Norway, such use is very limited (Hohle 2005).

3.2.1.2.6 *Waste*

District heating plants and incineration plants report annually combusted amounts of waste (boilers) to Statistics Norway and the Norwegian Environment Agency. There is also some combustion in manufacturing industries, reported to Statistics Norway.

According to the Norwegian Pollution Act, each incineration plant has to report emission data

for SO₂, NO_x, CO, NH₃, particles, heavy metals and dioxins, and the amount of waste incinerated to the county governor. The county governor then reports this information to the Norwegian Environment Agency. If emissions are not reported, the general method to estimate emissions from waste incineration is to multiply the amount of waste used by an appropriate emission factor. Normally a plant specific emission factor is made for the component in question. This factor is based on the ratio between previous emission figures and quantities of waste burned. This factor is then multiplied with the amount of waste incinerated that specific year.

3.2.1.2.7 *Energy balance sheets vs. energy accounts*

There are two different ways of presenting energy balances: Energy balance sheets (EBS) and energy accounts. The energy figures used in the emission calculations are mainly based on the energy balance sheets.

The energy accounts follow the energy consumption in Norwegian economic activity in the same way as the national accounts. All energy used by Norwegian enterprises and households is to be included. Energy used by Norwegian aviation and navigation abroad is also included, while the energy used by foreign transport industries in Norway is excluded.

The energy balance sheet follows the flow of energy within Norway. This means that the figures only include energy sold in Norway, regardless of the users' nationality. This leads to deviations between the energy balance sheet and the energy accounts, especially for international shipping and aviation.

The energy balance sheet has a separate item for energy sources consumed for transportation purposes. The energy accounts place the consumption of all energy under the relevant consumer sector, regardless of whether the consumption refers to transportation, heating or processing.

Figures from the energy sources balance sheet are reported to international organisations such as the OECD and the UN. The energy balance sheet should therefore usually be comparable with international energy statistics.

Important differences between figures presented in the energy balance sheet (EBS) and figures used in the emission calculations (EC) are:

- *Air transport*: EC use only Norwegian domestic air traffic (excluding military), while EBS includes all energy sold in Norway for air transport, including military and energy used for international air transport.
- *Coal/coke for non-energy purposes*: This consumption is included in net domestic consumption in EBS, whereas EC include only energy used for combustion in the calculation of emissions from energy. Emissions from coal and coke used as reducing agents are accounted for in the IPPU sector.

3.2.1.3 **Emission factors**

Emission factors used for the energy sector are listed by Statistics Norway, see link in appendix B. Emission factors for SO₂ are independent of combustion technology. In cases where

technology for cleansing of SO₂ has been installed, this will be reflected in the emission figures reported from the respective plants. For the other emission components, further descriptions are also given for each source sector.

The emission factors of NO_x, CO, NMVOC and NH₃ for stationary combustion have been evaluated by Norsk Energi for the Norwegian Environment Agency. The evaluation is described in the report "Vurdering av utslippsfaktorer for beregning av NO_x-utslipp med mer fra stasjonær forbrenning i Norge" (Evaluation of NO_x emission factors etcetera from stationary combustion in Norway) (Norsk Energi 2003). The report focused mainly on NO_x, but emission factors for CO, NMVOC and NH₃ were also considered.

The conclusion in Norsk Energi (2003) was that there are significant discrepancies between the emission factors from literature and the factors used in the inventory. Some of the emission factors used in the national inventory are higher and some lower than the emission factors found in literature. To some extent the discrepancy is due to the fact that the emission factors from literature are not reflecting technology used in Norway and are therefore not valid for Norwegian conditions. In addition it is considered that some of the Norwegian emission factors are based on more reliable data than the factors from literature. However, Norsk Energi (2003) proposed to change some of the emission factors, due to the fact that the factors from literature were considered to be of better quality than those used in the Norwegian emission inventory. One of the factors was the NO_x emission factor for heavy fuel oil, see below. In general, for all other compounds the emission factors proposed in Norsk Energi (2003) were lower than the emission factors that are used in the Norwegian emission inventory. We consider that the effect on national totals of not replacing the emission factors with the proposed factors in Norsk Energi (2003) has led to overestimated emissions. However, Norway is continuously considering all aspects of the Norwegian emission inventory, including the emission factors, and with the updated EMEP Guidebook (EEA 2016) we now consider to evaluate the emission factors in our inventory.

3.2.1.3.1 NO_x

The NO_x emission factors used in the Norwegian emission inventory have, as mentioned above, been evaluated by Norsk Energi in Norsk Energi (2003) and also in "NO_x-utslipp i forbindelse med eventuell NO_x-avgift" (Evaluation of NO_x emissions in connection with implementing NO_x tax) (Norsk Energi 2006). The conclusion in both reports is that the NO_x emission factors used in the inventory are within the intervals Norsk Energi found in their own measured data and from literature.

Norsk Energi (2003) concluded that the general emission factor for heavy fuel oil should be considered to be changed from 4.2 to 5 kg NO_x per tonne fuel and for chemical and metal industry from 5 to 6 kg NO_x per tonne heavy fuel oil. The consumption of heavy fuel oil in stationary combustion in Norway is very small and NO_x emissions in the Norwegian inventory from the largest consumers of heavy fuel oil in industry are based on plant specific data. Due to this, the proposed emission factors from Norsk Energi (2003; 2006) are not included in the Norwegian emission inventory.

3.2.1.3.2 SO₂

The emission factors for SO₂ from oil products change yearly, in accordance with variations in the sulphur content in the products. The presented factors refer to uncleansed emissions; in cases where the emissions are reduced through installed cleansing measures, this will be reflected in emission figures reported from the respective plants.

3.2.1.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C, as well as under the individual underlying source categories.

Generally, the total energy use is less uncertain than the energy use in each sector. For some sectors (e.g. the energy and manufacturing industries) the energy use is well known, while it is more uncertain in households and the service sectors. The energy use in the most uncertain sectors has been adjusted in the official energy statistics, so that the sum of the energy use in all sectors equals the total sales.

3.2.1.5 QA/QC

The emission sources in the energy sector are subjected to the QA/QC procedures described in section 1.6.

3.2.2 Energy industries

NFR 1A1

Last update: 10.03. 15

3.2.2.1 Description

Energy industries include emissions from electricity and heat generation and distribution, extraction of oil and natural gas, coal production, gas terminals and oil refineries. Norway produces electricity mainly from hydropower, so emissions from electricity production are small compared to most other countries. Due to the large production of oil and gas, the emissions from combustion in energy production are high.

Emissions from drilling at moveable offshore installations are included here. Emissions from these installations while not in operation (during transport, etc.) are included with 1A3d Navigation.

3.2.2.2 Method

A general description of the method used for estimation of emissions from fuel combustion is given in section 3.2.1.1. For waste incineration, also a more detailed description of the methodology for some components is given in this section.

3.2.2.2.1 Waste incineration

NO_x

Emissions of NO_x are reported from each plant to the Norwegian Environment Agency. An estimated amount of 2.5 per cent of this NO_x is subtracted and reported to UNFCCC as N₂O (SFT 1996). Accordingly, the net NO_x emissions constitute 97.5 per cent of the emissions reported by

the plants. For some years, emissions of NO_x have not been reported for a number of plants. In these cases, specific emission factors for the plants have been made, based upon earlier emissions and amounts of waste incinerated. These new factors have been used to estimate the missing figures.

Particles

Emissions of particles from district heating plants are reported to the Norwegian Environment Agency. The different plants started to report particulate emissions at various points in time. Most of them started reporting from 1994. Emissions of particles in the years before reporting have been assumed to be the same as in the first year the plant reported. New control device systems (mainly wet scrubbers) were installed at the end of the 1980s at the largest plants. Around 1995 more control device systems were installed as a result of stricter emission requirements. Most plants today have fabric filter or electrofilter together with wet scrubbers. Only two plants do not have wet scrubbers.

The emission permits do not state which particle fraction that is going to be measured. It is common to measure total amount of particles. It is however presumed that the particles emitted are less than PM_{2.5}. TSP and PM₁₀ are therefore the same as PM_{2.5}.

Dioxins

Emissions of dioxins from waste burning at district heating plants are reported to the Norwegian Environment Agency from the period 1994/1995. Before 1994 there is only national totals. For estimating the emissions of dioxins for each plant before 1994 an emission factor is derived from total amount of waste burned together with the total dioxin estimate. The emissions of dioxins were estimated by multiplying the given emission factor of 20 µg/tonne waste by the amount of waste burned at each plant. This calculation was done for each of the missing years for plants that did not report emissions.

PCB

PCB emissions are not reported to the Norwegian Environment Agency. A country specific emission factor has been used to estimate PCB. To take into accounts emission reduction systems implemented in incinerators, this emission factor decreases during the period, following the trend of dioxins emission factor.

Heavy metals

Emissions of heavy metals from waste combustion at district heating plants is reported to the Norwegian Environment Agency. Before 1999, many emissions of heavy metals were reported together as one group. This made it difficult to use the data to estimate the emission of each component. From 1999, there are separate data for each component, but for As, Cr and Cu there are a few plants that have not detailed reporting. To calculate the emissions of heavy metals before 1999 we have estimated an emission factor for each plant with the aid of reported emission data and amount of waste burned at each plant. The emission factor derived has been used to calculate emissions for previous years by multiplying each specific emission factor with the amount burned for the corresponding year for each plant.

Every district heating plant faced stricter emission requirements for particles from 1995. It is

expected that the emissions of heavy metals, except for mercury, were reduced analogously. At the same time, the emission of mercury was regulated from 0.1 mg/Nm³ to 0.05 mg/Nm³. These regulations are considered while calculating emissions for previous years.

3.2.2.3 Activity data

3.2.2.3.1 Electricity and heat generation and distribution

The energy producers annually report their use of different energy carriers to Statistics Norway. There is only some minor use of oil products at plants producing electricity from hydropower. Combustion of coal at Norway's only dual-purpose power plant at Svalbard/Spitsbergen is of a somewhat larger size. The amount of waste combusted at district heating plants is reported annually both to Statistics Norway and the Norwegian Environment Agency. The data are considered to be of high quality.

3.2.2.3.2 Extraction of oil and natural gas

Production of oil and natural gas is the dominating sector for emissions from combustion in the energy industries in Norway. The Norwegian Petroleum Directorate annually reports the amounts of gas combusted in turbines and diesel burned in turbines and direct-fired furnaces on the oil and gas fields. The data are of high quality, due to the CO₂ tax on fuel combustion. These activity data are used for 1990-2002. From 2003 onwards, reported activity data from the field operators are used.

3.2.2.3.3 Coal production

Norway's coal production takes place on Svalbard. The only coal producing company annually reports its coal consumption and some minor use of oil products. In addition to emissions related to Norway's own coal production, emissions from Russian activities are also included in the Norwegian emission inventory. Russian activity data are scarce, and emissions from an estimated quantity of coal combusted in Russian power plants are calculated. Since 1999 there has been only one such plant, in earlier years there were two.

3.2.2.3.4 Gas terminals

Natural gas from the Norwegian continental shelf is landed, treated and distributed at gas terminals on shore. There are four gas terminals in Norway. The eldest started up before 1990, one in 1996 and two in 2007. Annual figures on natural gas combusted in turbines and flared are reported to the Norwegian Petroleum Directorate (figures on flaring at one plant is reported to the Norwegian Environment Agency).

3.2.2.3.5 Gas power plants

Norway has one major gas power plants and several minor ones. One large plants was opened in 2007 and ran intermittently, depending on electricity and gas prices, to 2011. The plant was permanently closed in 2014. The second large plant was opened in 2010. Several of the smaller plants are back-up plants that are run only in emergency situations. Thus, there will be large annual fluctuations in emissions. In addition, there are gas power plants within the oil and gas extraction industry which are reported there.

3.2.2.3.6 Oil refineries

The oil refineries annually report their use of different energy carriers to Statistics Norway and the Norwegian Environment Agency. Refinery gas is most important, but there is also some use of LPG and oil products. Burning of coke while regenerating the catalyst in cracker units is reported under 1B2A4 – Fugitive emissions – Refining/Storage.

3.2.2.4 Emission factors

Emission factors used for the energy industries are listed by Statistics Norway, see link in appendix B. For some industries and components more information about the derivation of the emission factors are given in this section.

3.2.2.4.1 SO₂

Russian electricity and heat production

Emissions from combustion of coal for electricity production in the Russian settlements on Svalbard are included in the Norwegian emission inventory. Up to 1998 there were two Russian settlements with electricity and heat production: Barentsburg and Pyramiden. Since the coal production at Pyramiden was closed down in 1998, the settlement was abandoned, and all activity now takes place in Barentsburg. For SO₂, emission factors are based on information from Trust Arktikugol in Moscow. From 1999, the factor 70 kg/tonne is used, and for earlier years 16 kg/tonne. The factor-estimated figures are reduced by 60 per cent, due to the assumption that such an amount of the sulphur is bound in the ash.

3.2.2.4.2 NO_x

Offshore installations

NO_x emissions from diesel engines at offshore installations were revised in 2014 based on Karlsson and Finborud (2012). The recommended factors are shown in Table 3.3.

Table 3.3. Recommended emission factors for NO_x for different engine types

| | Before 2000 kg NO _x /tonne fuel | After 2000 kg NO _x /tonne fuel | Previous default factor |
|--|---|--|-------------------------------|
| 200-1000 rpm: Medium speed | 54 | 53 | 70 |
| 1000-1500 rpm: High speed, lower range | 50 | 50 | 60 |
| > 1500 rpm: High speed, higher range | 45 | 44 | 55 |

Source: Karlsson and Finborud (2012)

From 2003, emissions at fixed installations and at moveable installations during drilling operations are taken from reports from operators. Some operators use default emissions factors, whereas an increasing fraction use plant-specific factors.

In the implementation of the factors from Karlsson and Finborud (2012), the following principles were used:

- Reported emissions with implied emission factors less than 1 per cent from the old default values (70/60/55 kg/t) were assumed to have used the default factors. These emissions were reduced to the new default values for engines from before 2000 (54/50/45 kg/t).

- Other reported emissions were assumed to have used plant-specific factors and were left unchanged.
- Emissions from other consumption in engines were calculated with a general factor of 54 kg NO_x/tonne. This applies to all emissions before 2003, and the remaining fraction of sales to the oil and gas industry from 2003 onwards.
- Emissions from use of marine gas oil for turbines have not been part of this revision.

3.2.2.4.3 TSP, PM₁₀ and PM_{2.5}

Electricity and heat generation

Emission factors for TSP, PM₁₀ and PM_{2.5} are based on emission data given in EPA (2002). EPA (2002) gives emission data based on measurements made from various boilers using different control device systems. The Norwegian power plant at Svalbard is equipped with a multicyclone, and emission factors derived from measurements from boilers controlled with multicyclone device systems are used.

Waste incineration

Emissions of particles from district heating plants are reported to the Norwegian Environment Agency.

3.2.2.4.4 BC

For energy industries, BC emissions are estimated as fractions of PM_{2.5}-emissions. The share of BC has been given by IIASA (Kupiainen & Klimont 2007) as it is used in the GAINS model.

Waste incineration

The share of BC among TSP proposed in Kupiainen and Klimont (2004) has been used to estimate BC emissions. It gives a mass share of 0.9 percent of BC in PM_{2.5}.

For incineration of special waste, the same BC share as for heavy fuel oil combustion in residential, commercial, services and agriculture has been used.

3.2.2.4.5 Dioxins and PAH

Electricity and heat generation

Dioxin emissions from coal combustion at the power plants at Svalbard are derived from emission factors found in literature. The emission factor used is the emission factor recommended in Bremmer et al. (1994). The same emission factor is also used in Parma *et al.* (1995) and Hansen (2000). Burning of coal at power plants is also expected to give particle-bound dioxin emissions, but because of the effective control device using multicyclone collector, the emissions are expected to be low.

Emission factors used for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are the emission factor recommended in Guidebook 2016. Chapter 5C1, table 3-1. Is used for the years 1990-1995, and table 3-2 is used for the years after 1995.

PAH emissions from waste incineration are calculated by emission factors and amount of waste burned. We have no plant or country specific emission profile of PAH from waste incineration at

district heating plants in Norway.

3.2.2.4.6 PCB

In order to take into account emissions regulation in waste incineration, which was first implemented in 1995 and reinforced in 2006, the PCB emissions factor decreases along the period. From 1990 to 1994, the value given in the EMEP/EEA guidebook (2007) has been used to estimate PCB emissions from municipal waste and special waste.

In 2010, a PCB emission factor for municipal waste was estimated from emission measurements of waste incinerators. This emission factor has been considered for the estimations of PCB emissions after 2006. For the period 1995-2005, an emission factor has been derived from the dioxins trend observed during the same period.

As no measurements have been done for incineration of special waste, emission factors for the years after 1995 has been estimated using the same trend as for municipal waste. Table 3.4 presents PCB emissions factors.

Table 3.4. Emission factors for PCB from waste incineration plants, mg/tonn of waste

| | 1990-1994 | 1995-2005 | 2006 -> |
|-----------------|-----------|-----------|----------|
| Municipal waste | 0.82 | 0.00064 | 0.000032 |
| Special waste | 5 | 0.0039 | 0.0002 |

Source: Statistics Norway, EA (2007)

3.2.2.4.7 Heavy metals

Electricity and heat generation

The emission factors for heavy metals used for calculating emissions from coal fired power plants are from EEA (2001). The factors are, however, not specific for coal fired power plants but standard factors recommended for calculating emissions from coal combustion in energy and transformation industries.

3.2.2.4.8 HCB

For energy industries, HCB emissions have been estimated for the use of coal, light fuel oil, wood waste, black liquor, municipal and other waste. Emission factors used for the energy sector are given in Appendix B.

Waste incineration

HCB emissions are not reported by waste incineration plants. In order to take into account emissions regulation in waste incineration, which was first implemented in 1995 and reinforced in 2006, the HCB emission factor decreases along the period. Most of installations have anticipated 2006 regulation. Therefore, the emission factor from waste incineration in Denmark (Nielsen et al. 2010) has been considered since the year 2005. For 1995, emission factor has been estimated using the reduction trend of dioxins emission factor observed between 1995 and 2004. This emission factor has been used for the period 1995-2004. For the period 1990-1994, an emission factor from a former guidebook (EEA 2007) has been considered. Emission factors are presented in the Table 3.5.

Table 3.5. Emission factors for HCB from waste incineration plants, mg/tonn of waste

| | 1990-1994 | 1995-2004 | 2005 -> |
|-----------------|-----------|-----------|---------|
| Municipal waste | 2 | 0.9 | 0.045 |
| Other waste | 10 | 4.5 | 0.2 |

Source: Statistics Norway, Nielsen et al. (2010), EEA (2007)

Extraction of oil and natural gas

HCB emissions from the use of marine gas oil in offshore platform have been estimated using the same emission factor as for the use of marine gas oil in ship (0.08 mg HCB/tonn of gas oil).

3.2.2.5 Uncertainties

Uncertainty estimates for long-range air pollutants are given in Appendix C. Since the energy use is well known for the energy industries, the uncertainty in the activity data is considered to be minor.

The uncertainty in the activity data is ± 3 per cent of the mean for oil, ± 4 per cent for gas and ± 5 per cent of the mean for coal/coke and waste.

3.2.2.6 Source specific QA/QC

The energy industries are subjected to the general QA/QC procedures described in section 1.6. Some source specific QA/QC activities were conducted in the following industries:

Heat generation in district heating plants

Emissions of heavy metals and POPs from waste incineration have been subject to detailed control. The estimates are based on measurements, but the values are uncertain due to high variability. Reported emission values can vary by orders of magnitude from year to year. Each historical value has been checked in the QA/QC process, and some data have been rejected and replaced by calculated values.

Extraction of oil and natural gas

For emissions of NO_x from turbines offshore, time series over the emissions calculated with field specific emission factors have been compared with the emissions given, using the earlier used average emission factor.

From 2003 onwards, field specific emission figures reported from the companies are used directly in the emission model. These figures are compared with emissions calculated on the basis of field specific activity data and emission factors.

3.2.3 Manufacturing industries and construction

NFR 1A2

Last update: 20.05.09

3.2.3.1 Description

Emissions from the sector of manufacturing industries and construction include industrial emissions that to a large extent originating from the production of raw materials and semi-

manufactured goods (e.g. metals, petrochemicals, pulp and paper and mineral products). These emissions are related to fuel combustion only, that is, emissions from use of fuel for heating purposes. Consumption of coal as feedstock and reduction medium is not included in this sector, but is accounted for under the industrial processes sector (chapter 4).

The interannual variation in implied emission factors for heavy metals and dioxins in 1A2 is in some cases considerable. Examples are emissions of Pb, Hg and dioxin in 1A2d. These emission estimates are based on a combination of reported figures from the plants to the Norwegian Environment Agency and emissions based on activity data multiplied with emission factors listed by Statistics Norway, see link in appendix B. Energy use from the same plants is reported to Statistics Norway. Whenever emissions are reported these figures are used in the inventory.

Emissions of Pb in 1A2d have increased since 2000 due to increased combustion of special waste at the plants. The EF used for burning of special waste is 14 g Pb/tonne while for instance the EF for burning of heavy fuel oil is 1 g Pb/tonne. Emissions have increased, while the total energy consumption (activity data) has been more or less at the same level since 1994, with a minor decrease the latest years. This has led to increased IEFs. For some glass production plants the reported emissions of Pb are not splitted into emissions from process and combustion. In these cases, the all emissions are placed on the source that is thought to be the most prominent.

For Hg the emissions in 1A2d follow the same trend as the amount of liquid fuels used. The IEFs increases the years the consumption of liquid fuels increases and decreases the years the consumption of liquid fuels decreases.

For dioxins, IEFs vary due to variations in reported figures for one plant. The plant burnt various waste fractions in addition to regular fuel. The plant was closed down during 2001.

3.2.3.2 Activity data

Most of the emission figures are calculated on the basis of activity data and emission factors. For some large plants, varying emission figures are based on reported figures from the plants.

Statistics Norway carries out annual sample surveys on energy use in manufacturing industries, which supply most of the data material for the calculation of combustion emissions in these sectors in cases when reported emission figures not are used. The energy use survey is assumed to cover approximately 96 per cent of the energy use in this sector. For the remaining companies, figures are estimated based on data from the sample, together with data on economic turnover, taking into account use of different energy carriers in the same industries and size groups. A change in methodology from 1998 has had minor consequences for the time series, since the energy use is mainly concentrated to a few major plants within the industry, from which data were collected in both the present and the earlier method. The data on energy use in manufacturing industries are considered to be of high quality. Information on use of waste oil and other hazardous waste is also collected through the energy use statistics. For the construction industry, the figures on use of the different energy carriers are partly taken from the annual sales statistics for petroleum products and partly projected from earlier surveys; the energy data are considered rather uncertain. In some sectors, autodiesel is mainly

used in machinery and off-road vehicles, particularly in mining and construction. This amount of fuel is based on reported consumption of duty-free autodiesel in the manufacturing industries and on reported sales of duty-free autodiesel to construction. The methods for calculating emissions from motorized equipment are discussed in section 3.2.4.7. Emissions from off-road machinery in manufacturing industries and construction are reported in NFR category 1A2g vii.

3.2.3.3 Emission factors

Emission factors used for the manufacturing industries are listed by Statistics Norway, see link in appendix B.

3.2.3.4 Uncertainties

Uncertainty estimates for long-range air pollutants are given in Appendix C. The energy use is considered well known for the manufacturing industries.

3.2.3.5 Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4 Transport

NFR 1A3

3.2.4.1 Aviation

NFR 1A3a

Last update: 12.02.2020

3.2.4.1.1 Method

The calculation methodology applied is described in Thovsen (2017). According to the IPCC Good Practice Guidance the methodology used is Tier 3a based on the detailed methodology in the EMEP/EEA (2001). The new method is based on Eurocontrols "Advanced Emission Model"- AEM, combined with data from all aircraft movements to and from Norwegian airports. The method has bottom up estimates and allows estimation of emissions and fuel consumption for different types of aircrafts according to the engine type (Eurocontrol 2015) and aircraft movements. All movements below 1000 metres are included in the "Landing Take Off" (LTO) cycle. Movements over 1000 metres are included in the cruise phase. All emissions from the LTO cycle from domestic and international aviation are included in national totals. Emissions from the national and international cruise cycle are reported separately as memo items (see section 3.2.6.3).

3.2.4.1.2 Activity data

The types of fuel used in aircrafts are both jet kerosene and aviation gasoline. The latter is used mostly in small aircrafts. The total sales of jet kerosene and aviation gasoline are retrieved from the sales statistics of petroleum products, and is believed to cover the actual sales of fuel at Norwegian airports. Helicopter data is collected from several Norwegian airlines as the data source with aircraft movements has incomplete helicopter data.

Domestic consumption prior to 1995 is estimated by extrapolation on the basis of domestic kilometres flown and is more uncertain. The time series for liquid fuels used in aviation is given in Table 3.6.

Table 3.6. Liquid fuels in aviation, 1990-2018. TJ

| | 1 A 3 a ii (i) Domestic aviation (LTO) | 1 A 3 a i (i) International aviation (LTO) |
|------|---|---|
| 1990 | 2 706,19 | 1 728,84 |
| 1995 | 3 438,85 | 1 393,28 |
| 2000 | 4 216,40 | 1 997,84 |
| 2005 | 3 605,36 | 1 850,01 |
| 2006 | 3 753,51 | 2 175,62 |
| 2007 | 3 829,49 | 2 200,73 |
| 2008 | 4 118,29 | 2 127,03 |
| 2009 | 3 992,53 | 2 007,35 |
| 2010 | 3 847,65 | 2 325,85 |
| 2011 | 3 957,20 | 2 509,32 |
| 2012 | 4 075,53 | 2 708,22 |
| 2013 | 3 948,38 | 2 852,79 |
| 2014 | 4 306,20 | 2 981,21 |
| 2015 | 4 105,69 | 2 858,11 |
| 2016 | 3 922,99 | 2 852,89 |
| 2017 | 3 835,24 | 2 935,74 |
| 2018 | 4 092,00 | 2 783,85 |

Source: Statistics Norway

3.2.4.1.3 Emission factors

Emission factors used are listed by Statistics Norway, see link in appendix B.

The Norwegian Petroleum Industry Association provides emission factors for SO₂ for the combustion of jet fuel and gasoline (Finstad et al. 2002a). The emission factor for SO₂ varies annually depending on the sulphur content of the fuel used.

New aircraft and flight phase specific emission factors for NO_x, CO, VOC and particles are given in EMEP/EEA 2016. All particles are found to be less than PM_{2.5} (EMEP/EEA, 2016). The detailed emission factors are combined with the specific fuel consumption for each aircraft and flight phase EMEP/EEA, 2016, flight data by aircraft type and route from Avinor and the airports (Air transport statistics (Statistics Norway (Annually-a); background data 2013) and route distances to give weighted emission factors on an aggregated level. There are separate factors for LTO and the cruise phase.

The share of BC is assumed to amount to 48 per cent of total PM (EMEP/EEA, 2016). Emission factors are given in Appendix B.

The weighted emission factors are combined with the activity data (fuel consumption) to estimate emissions from civil aircraft.

The new emission factors for civil aircraft except helicopters have been used in the inventory back to 1990. Emission factors for helicopters and military aircraft were kept unchanged (EEA 2001; Finstad et al. 2002a).

3.2.4.1.4 *Uncertainties*

Activity data

The uncertainty in the activity data for civil aviation is estimated to be ± 20 per cent of the mean, primarily due to the difficulty in separating domestic emissions from emissions from fuel used in international transport (Rypdal & Zhang 2000). As described above, data before 1995 are more uncertain than for later years. This may also to a certain degree affect the time series consistency.

3.2.4.1.5 *Source specific QA/QC*

In 2018, a methodology improvement was made in the emission calculations for civil aviation (Thovsen 2017). According to the IPCC Good Practice Guidance the methodology used is Tier 3a based on the detailed methodology in EMEP/EEA (2001). This methodology allows estimation of emissions and fuel consumption for different types of aircrafts according to aircraft movements. The estimation is provided with new emission and fuel consumption factors for civil aircraft, except helicopters, from EMEP/EEA (2001); EEA (2013).

3.2.4.2 Road transport

NFR 1A3b i-v

Last update: 19.02.2019

3.2.4.2.1 *Description*

Emissions from this source includes combustion emissions from vehicles driven on roads, i.e. the categories passenger cars, light duty vehicles, heavy duty vehicles (including buses and coaches) and mopeds and motorcycles, as well as NMVOC emissions from gasoline evaporation. The methodology used for calculating emissions is described in more detail i Holmengen and Fedoryshyn (2015). The methodology corresponds to a Tier 3 methodology from the EMEP/EEA 2016 Guidebook (EEA 2016), using detailed information on vehicle fleet composition and driving patterns.

For passenger cars and light duty vehicles there has been a marked shift from petrol to diesel vehicles. In the 1990's petrol consumption within road transport far exceeded auto diesel. Distance driven by petrol passenger cars equally exceeded distances for auto diesel passenger cars. From 1.1.2007, there was a change in the registration tax for new passenger cars, and CO₂ became one parameter in calculating the level in addition to curb weight and engine power. This led to an increase in the sale of new diesel passenger cars. In 2006, the share of diesel vehicles within new passenger cars was 48 per cent. In 2007, the same share had increased to 74 per cent and was steady on that level until 2011. Since 2012, a NO_x – component was added to the passenger cars taxation. The share of diesel cars sold has since then been strongly reduced.

Norwegian political incentives for many years has resulted in an expanding purchase of electrical vehicles and plug-in hybrid vehicles, mainly passenger cars. In 2017, the number of electric passenger cars rose by 43 per cent, and amounts to 5 per cent of the fleet of passenger cars. In Norway it is a requirement that 10 per cent of the turnover from fuel sales is biofuel. In the last years the share of biofuel in biofuel blends has increased, and amount to 16 per cent in 2017.

3.2.4.2.2 *Method*

The consumption of gasoline, including bioethanol, for road traffic is estimated as total sales minus consumption for other uses, i.e. a top-down approach. Other uses for gasoline including bioethanol are e.g. small boats, snow scooters and motorized equipment. For auto diesel, the total consumption in road traffic is all auto diesel, including bio-diesel, charged with auto diesel tax. Consumption on CNG is based on a survey reported by suppliers of CNG. Consumption of LPG is estimated based on figures from the sales statistics on petroleum products and figures from “Drivkraft Norge”, a Norwegian association for the fuel and energy sector in Norway.

Pollutants other than CO₂ and fuel-related substances such as SO₂ and heavy metals, are estimated by the emission model of the Handbook of Emission Factors (HBEFA; (INFRAS 2009)). The current inventory uses HBEFA version 3.3. The model uses a mileage approach: Emissions = mileage * emission per km. The model results are used directly without any adjustment for discrepancies between estimated and registered fuel consumption.

The HBEFA model provides emission factors and possibilities for calculating emissions for segments and sub-segments for six vehicle classes: passenger cars, light commercial vehicles, heavy commercial vehicles, urban buses, coaches and motorcycles (including mopeds). The segments are based on engine volume for passenger cars and motorcycles, total weight for heavy commercial vehicles, urban buses and coaches, and tare weight for light commercial vehicles. The segments are further disaggregated to subsegments based on fuel type and technology type (e.g. Euro-1 – Euro-6). The segments used for Norway in the HBEFA model are given in Table 3.7.

Table 3.7. Segments used for Norway in the HBEFA

| Vehicle class | Segment | Fuel type | Segment split based on | Engine volume/weight class |
|---------------------------|----------------------------|-----------|------------------------|----------------------------|
| Passenger car | PC petrol <1,4L | Petrol | Engine volume | < 1.4 litres |
| | PC petrol 1,4-<2L | Petrol | Engine volume | 1.4- 2.0 litres |
| | PC petrol >=2L | Petrol | Engine volume | >= 2.0 litres |
| | PC Hybrid petrol/el medium | Petrol | - | All engine volumes |
| | PC PHEV petrol | Petrol | - | All engine volumes |
| | PC diesel <1,4L | Diesel | Engine volume | < 1.4 litres |
| | PC diesel 1,4-<2L | Diesel | Engine volume | 1.4- 2.0 litres |
| | PC diesel >=2L | Diesel | Engine volume | >= 2.0 litres |
| | PC hybrid diesel/el medium | Diesel | - | All engine volumes |
| | PC PHEV diesel | Diesel | - | All engine volumes |
| | PC LPG | LPG | - | All engine volumes |
| Light commercial vehicles | LCV petrol M+N1-I | Petrol | Tare weight | < 1305 kilos |
| | LCV petrol N1-II | Petrol | Tare weight | >= 1305-1760 kilos |
| | LCV petrol N1-III | Petrol | Tare weight | >= 1760-3859 kilos |
| | LCV diesel M+N1-I | Diesel | Tare weight | < 1305 kilos |
| | LCV diesel N1-II | Diesel | Tare weight | >= 1305-1760 kilos |
| Heavy goods vehicles | LCV diesel N1-III | Diesel | Tare weight | >= 1760-3859 kilos |
| | RT petrol | Petrol | - | All gross weights |
| | RigidTruck <7,5t | Diesel | Gross weight | <= 7.5 tonnes |
| | RigidTruck 7,5-12t | Diesel | Gross weight | > 7.5 - 12 tonnes |
| | RigidTruck >12-14t | Diesel | Gross weight | > 12 - 14 tonnes |
| | RigidTruck >14-20t | Diesel | Gross weight | > 14 - 20 tonnes |
| | RigidTruck >20-26t | Diesel | Gross weight | > 20 - 26 tonnes |
| | RigidTruck >26-28t | Diesel | Gross weight | > 26 - 28 tonnes |
| | RigidTruck >28-32t | Diesel | Gross weight | > 28 - 32 tonnes |
| | RigidTruck >32t | Diesel | Gross weight | > 32 tonnes |
| | Tractor for AT <=7,5t | Diesel | Gross weight | <= 7.5 tonnes |
| | Tractor for AT>7,5-14t | Diesel | Gross weight | > 7,5 - 14 tonnes |
| | Tractor for AT>14-20t | Diesel | Gross weight | > 14 - 20 tonnes |
| | Tractor for AT>20-28t | Diesel | Gross weight | > 20 - 28 tonnes |
| | Tractor for AT >34-40t | Diesel | Gross weight | > 34 - 40 tonnes |
| | Tractor for AT >40-50t | Diesel | Gross weight | > 40 - 50 tonnes |
| | Tractor for AT >50-60t | Diesel | Gross weight | > 50 - 60 tonnes |
| Coach | Coach Std <=18t | Diesel | Gross weight | <= 18 tonnes |
| | Coach 3-Axes >18t | Diesel | Gross weight | > 18 tonnes |
| Urban bus | Ubus Midi <=15t | Diesel | Gross weight | <= 15 tonnes |
| | Ubus Std >15-18t | Diesel | Gross weight | >15 - 18 tonnes |
| | Ubus Artic >18t | Diesel | Gross weight | > 18 tonnes |
| | Ubus Std >15-18t CNG | CNG | Gross weight | >15 - 18 tonnes |
| | Ubus Artic >18t CNG | CNG | Gross weight | > 18 tonnes |
| Motorcycles and mopeds | Moped <=50cc (v<50kmh) | Petrol | Engine volume | <= 50 cc |
| | MC 2S <=150cc | Petrol | Engine volume | <= 150 cc |
| | MC 2S >150cc | Petrol | Engine volume | >150 cc |
| | MC 4S <=150cc | Petrol | Engine volume | <= 150 cc |
| | MC 4S 151-250cc | Petrol | Engine volume | 151-250 cc |
| | MC 4S 251-750cc | Petrol | Engine volume | 251-750 cc |
| | MC 4S >750cc | Petrol | Engine volume | > 750 cc |

Source: Statistics Norway/Norwegian Environment Agency

The model combines the number of vehicles within each segment with driving lengths for the same segments to produce annual national mileage per subsegment. For heavy goods vehicles, the vehicle number is corrected for vehicles driving with trailers, and the driving is split into three load classes (empty, half loaded and fully loaded).

The annual national mileage is split between shares driven in different traffic situations. The traffic situations are a combination of area (urban/rural), road type (e.g. trunk road and access road), speed limit and level of service (freeflow, heavy, saturated, and stop and go). The traffic situations are further disaggregated by gradients, where the amount of driving on roads with slopes ranging from -6 per cent to 6 per cent is specified for each traffic situation.

Hot emission factors are provided on the disaggregated level of subsegments and traffic situations with different gradients, and the emissions are estimated after these steps of disaggregation.

The HBEFA model provides emission factors for cold emissions and evaporative emissions (soak, running losses and diurnal), in addition to hot emission factors. In order to calculate cold and evaporative emissions, information on diurnal variation in curves of traffic, trip length distributions, parking time distributions and driving behaviour distributions must be provided, in addition to variations in mean air temperature and humidity.

3.2.4.2.3 Activity data

All activity data are, as far as possible, updated for every year of the inventory. Data are taken primarily from official registers, public statistics and surveys. However, some of the data are based on assumptions. Many of the data sources are less comprehensive for the earliest years in the inventory. The sources of activity data are listed below:

- **Total fuel consumption:** the total amounts of fuels consumed are corrected for off-road use (in boats, snow scooters, motorized equipment, etc.). These corrections are estimated either from assumptions about the number of units, annual operation time and specific fuel consumption, or from assumptions about and investigations of the fraction of consumption used off-road in each sector. Statistics Norway's sales statistics for petroleum products supply the data for total fuel consumption (Statistics Norway Annually-b). Fuel consumption in road transport is given in Table 3.8.
- **Number of vehicles:** the number of vehicles in the various categories and age groups is taken from the statistics on registered vehicles, which receives data from the official register of the Norwegian Directorate of Public Roads. The model input is number of vehicles per vehicle class for each inventory year, and the share of vehicles for any given combination of segment and fuel type. These data are combined with information on the introduction of technology classes to provide number of vehicles within each subsegment. The information on introduction of technology classes are for recent years based on information from the official register of the Norwegian Directorate of Public Roads, and on legislation for the years in which the information in the register is insufficient.
- The HBEFA model distinguishes between two types of buses: urban buses, mainly used for urban driving, and coaches, mainly used for rural and motorway driving. Due to lack of specific information to make this split in the national vehicle register, the distinction between

urban buses and coaches are based on a methodology used in Sweden (Swedish environmental protection agency 2011), where the split is made based on the ratio p/w . Here, p is equal to the maximum allowed number of passengers (number of seats plus number of allowed standing passengers), and w is equal to the gross vehicle weight. These data are available from the national vehicle register. Buses with a p/w -value above 3.75 are classified as urban buses, whereas buses with a p/w -value below 3.75 are classified as coaches.

- *Average annual mileage*: Mileages for passenger cars, light commercial vehicles, heavy goods vehicles, coaches and urban buses are from 2005 onwards based on odometer readings taken during annual or biannual roadworthiness tests. The readings are collected by the Directorate of Public Roads and further processed by Statistics Norway (Statistics Norway 2010a). For earlier years, most figures are determined from surveys by Statistics Norway or the Institute of Transport Economics. In some instances assumptions are needed.
- The statistics on number of vehicles depict the vehicle fleet per December 31st of the inventory year, while the statistics on mileages represents annual driving for the entire year, including vehicles that have been scrapped or in other ways been in the vehicle fleet for only parts of the inventory year. To adjust for this discrepancy for the years 2005-2011, mean annual driving lengths for each vehicle category have been adjusted upwards in such a way that the totals correspond to the total annual traffic activity from the statistics on annual driving lengths.
- The average annual mileages vary as a function of age, with older vehicles generally driving shorter annual distances than newer vehicles. The correction of driving as a function of vehicle age is based on odometer readings taken during the roadworthiness test. The functions are calculated as the mean of the years 2005-2011, and the same correction curve is used for all years.
- Motorcycles and mopeds are not subject to roadworthiness tests in Norway. Average annual mileages are taken from a report on transport volumes in Norway (Vågane & Rideng 2010). Due to lack of data, corrections of annual mileage as a function of age for motor cycles and mopeds are taken from a Swedish survey (Bjørketun & Nilsson 2007) under the assumption that annual mileages as a function of age are comparable in Norway and Sweden.
- *Load data* are taken from the Road goods transport survey (Statistics Norway 2010b).
- *Transformation patterns* are calculated using information from Statistics Norway' Road goods transport survey on use of trailers and trailer size (Statistics Norway 2010b).
- *Traffic situations*: The Directorate of Public Roads has data on the annual number of vehicle-kilometres driven on national and county roads. The data are allocated by speed limits, road type, area type (urban/rural), and vehicle size (small/ large). Traffic on municipal roads (approx. 15 per cent) is estimated by Statistics Norway based on road lengths, detailed population data, traffic on adjoining roads, etc. The HBEFA model has emission factors for different situations of traffic flow (freeflow, heavy traffic, saturated traffic, and stop and go). Assumptions have been made as to this distribution for the different combinations of area type, road type and speed limits for Norway. Effects of road gradients are included, based primarily on Swiss data supplied to the HBEFA.
- *Ambient conditions* (air temperature and humidity) are included in the model to calculate cold and evaporative emissions. An average of five larger Norwegian cities has been used for spring, summer, autumn and winter separately. The data are based on measurements from

the Norwegian Meteorological Institute.

- *Trip length and parking time distributions* are calculated from the Norwegian travel survey (Institute of transport economics 1993). The distributions are given on an hourly basis.

Table 3.8. Fuel consumption in road transport, 1990-2017. TJ

| Year | Petrol | Diesel | LPG | Gaseous fuels | Biofuels |
|------|--------|---------|-----|---------------|----------|
| 1990 | 71 509 | 26 532 | - | - | - |
| 1995 | 64 664 | 34 588 | - | - | - |
| 2000 | 64 376 | 46 539 | - | 22 | 92 |
| 2005 | 62 647 | 64 231 | 200 | 69 | 116 |
| 2006 | 60 019 | 70 570 | 234 | 106 | 231 |
| 2007 | 56 697 | 76 900 | 269 | 109 | 1 273 |
| 2008 | 52 831 | 78 706 | 259 | 137 | 3 383 |
| 2009 | 49 382 | 79 997 | 277 | 145 | 3 991 |
| 2010 | 46 321 | 86 152 | 360 | 186 | 4 849 |
| 2011 | 41 800 | 89 895 | 253 | 213 | 4 791 |
| 2012 | 38 780 | 93 057 | 215 | 485 | 5 506 |
| 2013 | 35 986 | 95 764 | 294 | 588 | 5 309 |
| 2014 | 34 135 | 99 506 | 305 | 614 | 5 307 |
| 2015 | 31 987 | 102 257 | 246 | 431 | 5860 |
| 2016 | 30 077 | 100 568 | 359 | 445 | 12 889 |
| 2017 | 29 106 | 88 882 | 341 | 531 | 20 515 |
| 2018 | 27 935 | 93 342 | 312 | 483 | 15 080 |

Source: Statistics Norway

3.2.4.2.4 Emission factors

Emission factors are taken from the Handbook of Emission Factors (HBEFA). Factors are given as emission per vehicle kilometres for detailed combinations of subsegments and traffic situations.

BC is estimated as a fraction of PM_{2.5} emissions. BC emission factors depend on the vehicle category, the type of fuel and the existence of a treatment before exhaust. Shares of BC in PM_{2.5} were given by IIASA (Kupiainen & Klimont 2004).

HCB emissions have been estimated using an emission factor extracted from a former version of the EMEP guidebook (EEA 2007).

PCB emissions factors from Andrijewski et al. (2004) has been used to estimate emission from road transport. Table 3.9 presents PCB emissions factors.

Table 3.9. PCB emissions factors for gasoline and diesel combustion

| | | |
|------------------------|------------|----------|
| Leaded gasoline | 106 | mg/tonne |
| Unleaded gasoline | 0.02 | mg/tonne |
| Diesel light vehicles | 0.00000005 | mg/tonne |
| Diesel, heavy vehicles | 0.00000539 | mg/tonne |

Source: Andrijewski et al. (2004)

It has been assumed that PCB emissions vary with the gasoline lead content. Therefore, PCB emissions factor for combustion of gasoline in cars varies from 1990 to 1997 as it is presented in Table 3.10.

Table 3.10. PCB emissions factors for gasoline combustion for the period 1990-1997. Mg PCB/ tonne

| | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
|----------|------|------|------|------|------|------|------|------|
| mg/tonne | 106 | 82 | 74 | 49 | 8 | 8 | 0.2 | 0.02 |

Source: Statistics Norway

Biofuels for transport are handled as separate fuels.

Average factors are listed by Statistics Norway, see link in appendix B.

3.2.4.2.5 Uncertainties

The uncertainty estimates are given in Appendix C.

The comparison of bottom up estimates of fuel consumption from HBEFA with total sales (source specific QA/QC) reveals a discrepancy of 5-15 per cent. This is deemed to be a reasonable difference. This discrepancy is handled differently for different emission components. The total consumption of each type of fuel is the most important parameter in relation to the reporting requirements of the UNFCCC, as this forms the basis for the calculation of CO₂ from road traffic (IPCC 2006). One kilogram of gasoline or autodiesel yields a fixed amount of CO₂ irrespective of vehicle type.

Guidelines for greenhouse gas reporting, the IPCC guidelines (IPCC 2006), states that CO₂ emissions should be calculated using fuel consumption, and that sold amount of fuel should form the basis. Calculations of emissions of CH₄, N₂O and many of the components of emissions reported to CLRTAP (e.g. NO_x and particulates), however, depends on more detailed information about vehicle types and driving patterns, and here a more detailed model (for example HBEFA) should be applied. The relationship between emissions and fuel consumption must be considered differently for the emission components that are directly dependent on the composition and quantity of fuel (CO₂, SO_x and heavy metals) and those who to a larger extent depend on the type of vehicle and driving mode (e.g. NO_x, CH₄, N₂O, NH₃, CO, particles).

Fuel consumption is not an input to HBEFA, where emissions are calculated based on mileage and number of vehicles in each subsegment of vehicle classes, as well as other data sets, such as cold start and age distribution of mileage. Fuel consumption is however calculated in the model similarly to emission calculations. Biofuels are not handled as separate fuels in HBEFA. The estimated fuel consumption for the country as a whole can be compared with sold fuel sales statistics for petroleum products and the energy balance (including biofuels). After the revision in the energy balance, the comparison shows that the the fuel consumption calculated in HBEFA is more equal to the fuel consumption in the energy balance, especially for petrol. The difference in consumption of auto diesel is fluctuating in the time series, but is less than before the revision. From 1999 the consumption in HBEFA is lower than in the energy balance.

It is not known why there is a discrepancy between the consumption in the energy balance and bottom-up calculation in HBEFA, but there are several possible explanations as to why fuel sold does not match the fuel consumption calculated from the road transport emission model:

1. *Fuel purchased by foreign vehicles. Foreign vehicles are not included in the vehicle register statistics, even though they drive on Norwegian roads. Similarly, no fuel bought by Norwegian vehicles abroad is sampled. It is likely that there is no systematic "fuel tourism" across the Norwegian border, as there are no significant price differences between fuel prices in Norway and Sweden. The current calculations are based on the assumption that driving in Norway by foreign vehicles equals the driving of Norwegian vehicles abroad.*
2. *Driving patterns.* There may be elements in the driving patterns that causes fuel consumption per kilometer per vehicle to be higher than what the model calculates. One possible reason here is that the fuel consumptions stated in the vehicle type approvals are used as part of the input to the model, and there is an ongoing discussion about whether these systematically underestimate consumption. These data are however available only for the latter part of the series, and can not explain the discrepancies in the 1990s.

Whether the emission calculations should be corrected for differences in fuel consumption depends on the pollutants in question. For those components that are directly dependent on the amount of fuel (CO₂, SO₂, heavy metals) it will always be appropriate to use the fuel consumption from the energy balance as a basis for calculation. For the other emission components, the decision on whether to correct for total fuel consumption or not will depend on what is causing the discrepancy between fuel consumption calculated in the model and fuel consumption in the energy balance. If the reason is that the total mileage is underestimated in the model, and that the energy balance represents a "truer" picture of the consumption of fuels, emissions should be corrected. If the discrepancy, however, is due to an underestimation of the fuel consumption per kilometer, the emission estimates should not be corrected unless one finds a clear correlation between changes in consumption per kilometer and emissions per kilometer for the relevant emission components. As long as we do not know the reason for the discrepancy, an assessment of data quality in the various input data is crucial to determining whether emissions should be reconciled against fuel sales or not.

In the previous model (SFT 1993; SFT 1999b), the emissions of all substances were corrected to account for the discrepancy between the energy balance and the model calculations, because the energy balance was considered the most secure data source. When HBEFA was introduced as the computational model, a new data source was also introduced, namely the mileage statistics from Statistics Norway. These statistics are based on data from periodical technical inspections, and goes back to 2005. This important new data source is considered to be of good quality, and it has changed the assessment of whether the emissions shall be corrected for the consumption in the energy balance or not. There is no reason to believe that the total run lengths are underestimated, and we consider it likely that the reason for the discrepancy lies in the estimates of fuel consumption per kilometer. The energy balance is based on the assessment that Norwegian purchases abroad correspond to foreign purchases in Norway, and the same assessment is applied to the emission calculations. We have not found any reason to believe that the reasons for the discrepancies in fuel consumption are directly correlated with driving behavior. It has therefore been assessed that HBEFA estimated emissions of pollutants

that are not directly related to fuel consumption should not be reconciled with fuel consumption.

There are currently no comprehensive statistics on foreign vehicles driving in Norway. One possible explanation for the discrepancy between the calculated fuel consumption in HBEFA and sold quantity of fuel is that foreign driving in Norway exceeds Norwegian driving abroad. There has been an issue that the proportion of heavy vehicles with foreign vehicles increases. Better data related to foreign driving in Norway and the Norwegian driving vehicles abroad would strengthen or refute the current assumption that these two balance each other out.

3.2.4.2.6 Source specific QA/QC

Top-down and bottom-up data on fuel consumption are compared for gasoline and diesel vehicles on an annual basis. The consumption of gasoline and auto diesel for road traffic is estimated as total sales minus consumption for other uses, i.e. a top-down approach. The HBEFA emission model also makes bottom-up estimates of consumption, which can be compared with the top-down data. The causes are on one hand uncertainties in the amount of non-road use and on the other hand uncertainties in mileage and specific consumption in road transport.

However, the total consumption of auto diesel is well known. The uncertainty concerns the allocation between road and non-road use, connected to illegal use of diesel without road tax in road traffic. The total emissions may be sensitive to this allocation, due to different emission calculation methodologies. When inspected, taxed and tax free diesel can be identified by colour. The fine for illegal use of tax free diesel is currently from NOK 20 000,- and upwards. There is no reason to believe that this is a major problem.

3.2.4.3 Railways

NFR 1A3c

Last update: 19.02.2019

3.2.4.3.1 Description

Railway traffic in Norway uses mainly electricity. Auto diesel is used at a small number of lines, for shunting etc. There is also a minor consumption of coal in museum railways.

3.2.4.3.2 Method

General estimation methodology for calculating combustion emissions from consumption figures and emission factors is used.

3.2.4.3.3 Activity data

Consumption figures for auto diesel used in rail transport is based on sales statistics for petroleum products. Consumption of coal is estimated based on information from different museum railways; the same figure is used for all years from 1990. There has been a continuous electrification of the railroads in Norway, and the diesel consumption is more than halved from the level in the early 1990's.

Table 3.11. Fuel consumption in railways, 1990-2017. TJ

| | Liquid fuels | Solid fuels |
|------|--------------|-------------|
| 1990 | 1265 | 3.9 |
| 1995 | 1359 | 3.9 |
| 2000 | 835 | 3.9 |
| 2005 | 670 | 3.9 |
| 2006 | 641 | 3.9 |
| 2007 | 665 | 3.9 |
| 2008 | 685 | 3.9 |
| 2009 | 626 | 3.9 |
| 2010 | 803 | 3.9 |
| 2011 | 796 | 3.9 |
| 2012 | 819 | 3.9 |
| 2013 | 718 | 3.9 |
| 2014 | 586 | 3.9 |
| 2015 | 547 | 3.9 |
| 2016 | 549 | 3.9 |
| 2017 | 589 | 3.9 |
| 2018 | 593 | 3.9 |

Source: Statistics Norway

3.2.4.3.4 Emission factors

Emission factors for NO_x, HC, CO, and PM₁₀ were estimated by Bang (1993) based on a literature survey and data on Norwegian usage profiles. The HC factor of 4 g/kg was used directly for NMVOC.

The other emission factors are the same as for diesel machinery in mining and quarrying (see section 3.2.4.7.4), with the following exception:

- NH₃: 0.007 g/kg vs 0.005 g/kg for machinery (EEA 2013)
- BC: emission factors as share of PM_{2.5} from IIASA (Kupiainen & Klimont 2004) have been used. They are given in Appendix B.

General emission factors for coal are used in the calculations.

3.2.4.3.5 Uncertainties

The consumption data are of high quality. Their uncertainty is estimated to be ± 5 per cent of the mean.

3.2.4.3.6 Source specific QA/QC

Foreign railways in Norway are expanding, and the sale statistics for petroleum products include sales to foreigners. Therefore, the sale statistics is considered a better source for fuel consumption than the consumption data from the Norwegian State Railways.

3.2.4.4 Electric railway conductors

NFR 1A3c

Last update: 01.09.05

3.2.4.4.1 Method

Electric railway conductors contain copper that is emitted in contact with trains. In the inventory, copper emissions are calculated by emission factors and activity data.

3.2.4.4.2 Activity data

The activity data used for calculating emissions of copper from electric wires are annual train kilometers given by the Norwegian State Railways (NSB).

3.2.4.4.3 Emission factors

According to Norwegian State Railways (Rypdal & Mykkelbost 1997) the weight of a contact wire is 0.91 kg/meters. The weight is reduced by 20 per cent after 3 million train passes. This gives an emission factor of 0.06 g/train kilometers. It is, however, uncertain how much of this is emitted to air. In the inventory it is assumed that 50 per cent is emitted to air. This gives an emission factor of 0.03 g/ train kilometer.

Table 3.12. Emission factor for electric railway conductors. g/km

| Emission factor (g/train kilometers) | |
|--------------------------------------|------|
| Cu | 0.03 |

Source: Norwegian Environment Agency

3.2.4.4.4 Uncertainties

The emission factor used is uncertain. First, there is an uncertainty connected to the reduction of 20 per cent after 3 million train passes. Secondly, there is uncertainty regarding the assumption that 50 per cent are emissions to air (Finstad & Rypdal 2003).

3.2.4.4.5 Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.5 Navigation

NFR 1A3d

Last update: 14.02.20

3.2.4.5.1 Description

According to CLRTAP Norwegian national sea traffic is defined as ships moving between two Norwegian ports. In this connection, installations at the Norwegian part of the continental shelf are defined as ports.

Fishing is described in section 3.2.5

3.2.4.5.2 Method

Emissions from navigation are estimated according to the Tier 2 IPCC methodology. Emissions from moveable installations used in oil and gas exploration and extraction are split between 1A1 – energy industries (section 3.2.2) and navigation: Emissions from drilling are reported under 1A1, while emissions from transport and other activities are reported under navigation. Emissions from international marine bunkers are excluded from the national totals and are

reported separately (section 3.2.6.2), in accordance with the IPCC Good Practice Guidance.

Annual emissions are estimated from sales of fuel to domestic shipping, using average emission factors in the calculations.

For 1993, 1998, 2004 and 2007 emissions have also been estimated based on a bottom-up approach. Fuel consumption data were collected for all categories of ships (based on the full population of Norwegian ships in domestic transport); freight vessels (bulk and tank), oil loading vessels, supply/standby ships, tug boats, passenger vessels, fishing vessels, military ships and other ships. Emissions were estimated from ship specific emission factors and fuel use. From this information, average emission factors were estimated for application in the annual update based on fuel sales. This approach is unfortunately too resource demanding to perform annually.

3.2.4.5.3 *Activity data*

The annual sales statistics for petroleum products gives figures on the use of marine gas oil, heavy distillates and heavy fuel oil in domestic navigation. Information on fuel used in the ship categories in the bottom-up analysis is mainly given by data from the Business Sector's NO_x fund for 2007 and by earlier SSB analyses for 1993 and 1998 (Tornsjø 2001), and 2004. Data on fuel consumed by public road ferries are available from the Directorate of Public Roads.

Information on fuel use at mobile drilling rigs is taken from the sales statistics, but information on use (whether it is used for drilling, stationary combustion etc.) is taken from Environmental Web/EPIM Environment Hub (EEH) (reported from oil companies to the Norwegian Environment Agency and the Norwegian Petroleum Directorate). Consumption during drilling activities are reported under "Energy industries" (CRF 1A1c). Only the remaining part of sales, assumed to be for drilling rigs during transit etc., is included with Navigation.

For marine gas oil, the amount used for navigation is equal to total sales figures except bunkers, after the deduction of estimated stationary use, mainly in oil and gas extraction, but also some minor use in manufacturing industries and construction. Due to inaccuracies in the reporting of distribution of marine gas oil between domestic and international shipping from approximately 2005, there is some uncertainty connected to the figures for the latest years.

Use of natural gas in navigation, which was introduced in 2000 and has increased considerably from 2007, is based on sales figures reported to Statistics Norway from the distributors. Fuel consumption in national navigation is given in Table 3.13.

Table 3.13. Fuel consumption in national navigation, 1990-2018. TJ

| | Liquid fuels | Gaseous fuels |
|------|--------------|---------------|
| 1990 | 27 644 | - |
| 1995 | 35 944 | - |
| 2000 | 37 795 | - |
| 2005 | 39 065 | 332 |
| 2006 | 41 260 | 329 |
| 2007 | 40 763 | 1 577 |
| 2008 | 38 512 | 1 856 |
| 2009 | 38 674 | 1 980 |
| 2010 | 43 536 | 2 077 |
| 2011 | 46 839 | 2 465 |
| 2012 | 47 239 | 3 173 |
| 2013 | 45 447 | 3 736 |
| 2014 | 41 711 | 4 392 |
| 2015 | 36 297 | 4 546 |
| 2016 | 33 474 | 4 074 |
| 2017 | 30 839 | 4251 |
| 2018 | 32 283 | 3817 |

Source: Statistics Norway

3.2.4.5.4 Emission factors

Emission factors used for navigation are listed by Statistics Norway, see link in appendix B.

SO₂

The emission factors are determined from the sulphur content of the fuel.

NO_x

NO_x factors for different engine types (slow, medium and high speed) have been estimated by Marintek based on data from a comprehensive measure programme for NO_x emissions from ships, which has been implemented under the leadership of the Business Sector's NO_x fund. The new basis factors from Marintek apply to emissions from different engine types built before and after emission restrictions were implemented in 2000 (Bremnes Nielsen & Stenersen 2009).

Table 3.14. Recommended emission factors for NO_x for different engine types

| | Engine building year | |
|-------------------------------------|---|--|
| | Before 2000 kg NO _x /tonne fuel | After 2000 kg NO _x /tonne fuel |
| Slow speed NO _x factor | 82 | 78 |
| Medium speed NO _x factor | 54 | 53 |
| High speed NO _x factor | 47 | 41 |

Source: Bremnes Nielsen and Stenersen (2009)

The factors were weighted in two steps: First, by engine type distribution within ship categories (passenger, general cargo, offshore, fishing, etc.). Secondly, by estimated fuel consumption among categories. The fuel consumption weights were calculated based on data for 1993, 1998,

2004 and 2007, which are years with good availability of activity data. Average factors for other years were interpolated. In the interpolation of the average factors over the time series, a peak in the use of shuttle tankers has been taken into consideration. The fact that we have reported data for public road ferries for some years, and a gradual change to new engines with lower emissions starting in 2000 due to new restrictions, has also been taken into consideration. The factors from Marintek are valid for engines with no particular NO_x reduction measures. The NO_x factors used in the inventory are documented in Flugsrud et al. (2010).

The method outlined above is used for the years up to 2007. From 2008 onwards, a large number of NO_x reducing technologies have been installed, funded through the NO_x fund and certified by emission measurements. Annual emissions are reported by companies to the NO_x fund and/or to the Norwegian Tax Administration as part of a national NO_x tax system.

In 2016, data on NO_x emissions and consumption of fuel in the third quarter of 2016 were collected from companies participating in the NO_x fund with ships operating in Norwegian coastal traffic. The data were made available to Statistics Norway for inventory preparation. The NO_x emissions in these data are mainly based on ship-specific measurements. For ships without measurements, a slightly adjusted version of the Marintek factors were used³. Based on these data, an average NO_x factor for domestic navigation in 2016 was calculated. Emission factors for 2008-2015 were obtained by linear interpolation.

For gas engines the NO_x factor 5.6 kg NO_x/ tonne LNG is established based on the mass of LNG consumed (Bremnes Nielsen & Stenersen 2010).

For offshore drilling rigs, the factor 54 kg NO_x /tonne is used (Karlsson & Finborud 2012). See further discussion on NO_x from offshore installations in the section on stationary combustion.

Average NO_x factors for fishing and for general shipping are listed by Statistics Norway, see link in appendix B.

NH₃

Emissions of NH₃ from navigation are reported as "Not Estimated". The EMEP/EEA Guidebook (EEA 2013) has no emission factors, and in table 2-2 over "Contributions to total emissions" NH₃ is stated as "No emissions reported".

Particles

Factors for particulate matter are based on measurements performed by MARINTEK and literature sources. The factors are presented in Table 3.15.

³ Medium speed engines 200-1000 rpm and high speed engines >1500 rpm used factors from table 3.14. Slow speed engines <200 rpm used 100 kg NO_x/tonne. Medium/high speed engines 1000-1500 rpm used 50 kg NO_x/tonne. (Norwegian Tax Administration, <http://www.skatteetaten.no/globalassets/saravgifter/avgiftsrunskriv/2016-nox.pdf>)

Table 3.15. Particulate matter emission factors for oil and gas operated vessels

| Fuel | Emission factor | |
|---|-------------------|------------------------|
| | PM _{2.5} | PM ₁₀ , TSP |
| Marine gas oil, light fuel oils (kg/tonne) | 1.5 | 1.6 |
| Heavy fuel oil, heavy distillate (kg/tonne) | 5.1 | 5.4 |
| LNG (kg/1000 Sm ³) | 0.032 | 0.032 |

Source: Bremnes Nielsen and Stenersen (2010).and Bremnes Nielsen (pers.comm.4)

For oil based fuels it is assumed that all particles are included in PM₁₀ and 95 per cent of the particles are included in PM_{2.5} (Finstad et al. 2003).

Emission factors for particle emissions from gas operated vessels are based on measurements made by MARINTEK (Bremnes Nielsen, *pers.comm*), which show 95-99 per cent emission reduction compared to marine gas oil.

BC

BC emissions are estimated using shares of PM_{2.5} as emission factors. Factors from the IIASA (Kupiainen & Klimont 2007) have been used.

Table 3.16. BC emission factors for oil operated vessels

| Fuel | Emission factor | |
|----------------------------------|--------------------------|-----|
| | PM _{2.5} (kg/t) | BC |
| Marine gas oil, light fuel oils | 1.5 | 40% |
| Heavy fuel oil, heavy distillate | 5.1 | 43% |

Source: GAINS, IIASA (2010)

As no share for BC was found in the literature for the use of natural gas in navigation, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

HCB

HCB emissions from the use of heavy fuel oil and marine gas oil have been estimated using the EMEP-EEA guidebook (2013).

Table 3.17. HCB emission factors for oil operated vessels

| Fuel | Emission factor | |
|----------------|-----------------|--|
| | HCB (mg/t) | |
| Marine gas oil | 0.08 | |
| Heavy fuel oil | 0.14 | |

Source: EEA guidebook (2013).

PCB

PCB emissions from the use of heavy fuel oil and marine gas oil are considered higher in the navigation sector due to the presence of chlorine in the air. Emission factors determined by

⁴ Bremnes Nielsen, J. (2010): Personal information, email from Jørgen Bremnes Nielsen, 11 Nov. 2010, Marintek.

Cooper (2004) have been used to estimate PCB emissions.

Table 3.18. HCB emission factors for oil operated vessels

| Fuel | Emission factor PCB (mg/t) |
|----------------|---------------------------------------|
| Marine gas oil | 0.36 |
| Heavy fuel oil | 0.60 |

Source: Cooper (2004)

3.2.4.5.5 Uncertainties

The estimation of fuel used by fishing vessels is assumed to be rather uncertain. There is also uncertainty connected to the fuel use for other domestic sea traffic due to uncertainty in the sales statistics for petroleum products. Particularly, the delimitation between sales of marine gas oil for national use and bunkers has become more uncertain from approximately 2005, due to new and less accurate reporting routines in some oil companies.

Some uncertainty is also connected to the emission factors.

The uncertainty in the activity data for navigation is assessed to be ± 20 per cent. The uncertainty in the NO_x factors depends both on the uncertainty in the basis factors from Marintek (Bremnes Nielsen & Stenersen 2009) and on the uncertainty in the allocations that are made of the factors between ship types and years. Marintek has estimated the uncertainty in their basis NO_x factors for different engine types to ± 5 per cent. Uncertainties in emission factors are shown in Table 3.19.

Table 3.19. Uncertainties in emission factors for ships and fishing vessels. Per cent

| | Standard deviation (2σ) |
|----------------------------|--|
| SO_2 | ± 25 |
| NO_x ¹ | ± 15 |
| NM VOC | ± 50 |

¹ It is assumed that the uncertainty might be lower now than in this estimate from Rypdal and Zhang (2001) since more measures have been performed in connection with the Business Sector's NO_x fund.

Source: Rypdal and Zhang (2001)

3.2.4.5.6 Source specific QA/QC

As mentioned, emission estimates for ships have been made bottom-up for 1993 and 1998 (Tornsjø 2001) and for 2004 and 2007. These results have been compared with top-down data (from sales) on fuel consumption used in the annual estimates.

The outcome showed that data from sales were only 1 per cent higher than data from reported consumption in 2007. For 2004 the sales data were 27 per cent higher than the consumption data in the bottom-up analysis. This can be explained by the fact that the bottom-up method does not cover all ships, but it may also be that the domestic/international distinction is not specified precisely enough in the sales statistics. Another element, which not has been taken into account, is possible changes in stock. For the years 1993 and 1998 a deviation of -12 and -15 per cent, respectively, has been found. In the calculations, sales figures are used, as they are assumed to be more complete and are annually available.

3.2.4.6 Pipeline

NFR 1A3e i

Last update: 22.03.10

Figures on natural gas used in turbines for pipeline transport at two separate facilities are reported annually from the Norwegian Petroleum Directorate to Statistics Norway. However, energy generation for pipeline transport also takes place at the production facilities. Specific data on consumption for transport are not available. Thus, the consumption at the two pipeline facilities does not give a correct picture of the activity in this sector. As a consequence, all emissions from pipelines are reported under NFR 1A1.

3.2.4.7 Motorized equipment

NFR 1A2 g-vii etc.

Last update: 19.02.2019

3.2.4.7.1 Description

The category "motorized equipment" comprises all mobile combustion sources except road, sea, air, and railway transport. Farm and construction equipments are the most important categories. Other categories include mines and quarries, forestry, snow scooters, small boats (leisure) and miscellaneous household equipment.

Emissions from motorized equipments are reported under several categories:

Table 3.20. Motorised equipment categories

| | NFR |
|--------------------------------|----------|
| Manufacturing and construction | 1A2g-vii |
| Commercial and institutional | 1A4a-ii |
| Households | 1A4b-ii |
| Agriculture/Forestry/Fishing | 1A4c-ii |
| Military | 1A5b |

Source: Statistics Norway/Norwegian Environment Agency

Primarily consumption of gasoline (including bioethanol) and auto diesel is considered.

3.2.4.7.2 Method

Emissions are estimated through the general methodology described in section 3.2.1.1, involving consumption figures and appropriate emission factors.

3.2.4.7.3 Activity data

Consumption of gasoline and auto diesel are handled differently. They are both based on data from the energy balance. Auto diesel used in off-road vehicles are tax-free from 1994, and tax-free auto diesel in the years 1990-1993 are extrapolated based on the split between diesel with or without tax in 1995-1998.

Small boats (leisure): The consumption of gasoline and tax-free auto diesel is estimated based on a model using data on size of the fleet, type of fuel, 2- and 4-stroke engine, size of engine. The data is collected from a survey (Båtlivsundersøkelsen) in 2010, and the time series are

extrapolated.

Other motorized equipments on tax-free auto diesel: is given as the difference between total sales of tax-free diesel and estimated use in railway transportation and small boats (leisure).

Snow scooters: A portion of 1 per cent of the gasoline consumption (including bioethanol) in households (mobile combustion) is assigned to snow scooters in the years 1990-2017.

Other motorized equipments on gasoline: 2 per cent of the gasoline consumption (including bioethanol) in households (mobile combustion) is assigned to other motorized equipments in the years 1990-2017. 97 per cent of the consumption in forestry is assigned to other motorized equipments.

3.2.4.7.4 *Emission factors*

Emission factors used are listed by Statistics Norway, see link in appendix B.

For diesel machinery, emission factors for HC, CO, and PM₁₀ were estimated by Bang (1993), based on a literature survey and data on Norwegian usage profiles. Source for emission factor for NO_x from diesel machinery is from Bang (1993) for motor gasoline and light fuel oils. For auto diesel, emission factors from a Danish report (Winther & Nielsen 2006) is used. NMVOC factors were calculated by subtracting an assumed CH₄ fraction of 0.3 g/kg diesel.

The emission factors used in the emission model for tractor and construction machinery are calculated from the basic factors in Winther and Nielsen (2006), weighted by the age and engine rating distribution of the tractor and construction machinery populations, as well as assumptions on motor load and operating hours and the introduction scheme for emission regulations by the EU (Stage I, II, III and IV).

3.2.4.7.5 *Uncertainties*

The estimates of consumption are considered quite uncertain. The total consumption of tax-free autodiesel is well known, while the distribution between the motorized equipments are not. Consumption in small boats (leisure) is the main factor to the uncertainty.

3.2.4.7.6 *Source specific QA/QC*

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.8 **Automobile tyre and brake wear**

NFR 1A3b vi

Last update: 22.12.16

3.2.4.8.1 *Tyre wear*

Description

Tyre wear is a source for emission of particles, heavy metals and persistent organic pollutants. The tyres are worn down by 10 to 20 per cent of its total weight during its lifetime. Most of the rubber is lost during acceleration and braking. All rubber lost is assumed to be particles containing heavy metals and PAH.

Method

A Tier 1 method is used to estimate emissions from tyre wear as speed-dependency is not accounted for.

Particles

All rubber lost is assumed to be small particles. The emissions of particles are calculated based on emission factors and annual mileage.

Heavy metals

Rubber particles contain heavy metals. Emissions of the heavy metals As, Cd, Cu, Cr, Pb and Hg are calculated based on annual mileage and emission factors.

PAH

The particles emitted from tyre wear contain PAH. Emissions are calculated based on emission factors and annual mileage.

Activity data

Annual mileage is used for calculating the emissions from tyre wear. Annual mileage is given by the road traffic model, see section 3.2.4.2.

Emission factors**Particles**

The emission factors used for calculating the emission of particles are given by TNO (Institute of environmental and energy technology 2002). The emission factors are based on several Dutch and British studies. Recommended emission factors for TSP and PM₁₀ are taken from 'Institute of environmental and energy technology (2002). Emission factor for PM_{2.5} was set to be zero. A new report from TNO (TNO 2008) presents emission factors for all three fractions of particulate matter. The emission factors for TSP and PM₁₀ are in the same range as the emissions factors given in 'Institute of environmental and energy technology (2002). In the Norwegian inventory, it has been chosen to include PM_{2.5} emissions using the same ratio between PM₁₀ and PM_{2.5} as the ratio between PM₁₀ and PM_{2.5} from TNO (2008). The emission factors used are given in Table 3.21. Institute of environmental and energy technology (2002).. Emission factor for PM_{2.5} was set to be zero. A new report from TNO presents emission factors for all three fractions of particulate matter. The emission factors for TSP and PM₁₀ are in the same range as the emissions factors given in '. In the Norwegian inventory, it has been chosen to include PM_{2.5} emissions using the same ratio between PM₁₀ and PM_{2.5} as the ratio between PM₁₀ and PM_{2.5} from . The emission factors used are given in .

Table 3.21. Emission factors for particles from tyre wear. kg/mill. km

| | TSP | PM₁₀ | PM_{2.5} |
|---------------------|------------|------------------------|-------------------------|
| Private cars | 69 | 3.45 | 0.69 |
| Van | 90 | 4.5 | 0.9 |
| Heavy duty vehicles | 371,25 | 18.563 | 3.71 |
| MC | 34,5 | 1.725 | 0.35 |

Source: TNO (Institute of environmental and energy technology 2002)

BC

BC is estimated as a fraction of $PM_{2.5}$. Emission factors depend on the type of vehicle. IIASA (Kupiainen & Klimont 2004) gives emission factors for Black Carbon and Organic Carbon as share of TSP. Since the sum of emissions of BC and OC has to be lower than $PM_{2.5}$ emissions, the emission factors have been adjusted.

Table 3.22. Emission factors for BC from tyre wear in share of $PM_{2.5}$. Particles are shown in kg/mill. km

| | TSP | $PM_{2.5}$ | BC |
|---------------------|------------|------------------------------|-----------|
| Passenger cars | 69 | 0.69 | 30% |
| Light duty vehicles | 90 | 0.9 | 30% |
| Heavy duty vehicles | 371,25 | 3.71 | 30% |
| MC | 34,5 | 0.35 | 30% |

Source: IIASA (Kupiainen & Klimont 2004)

Heavy metals

The emission factors used for the heavy metals As, Cd, Cu, Cr and Pb are derived from a particle-heavy metal distribution given by Dutch studies (van den Brink 1996). The content of heavy metals in the particles, given by this distribution, is multiplied by the PM_{10} emission factor (Table 3.21). This gives the emission factors for the heavy metals As, Cd, Cu, Cr and Pb from tyre wear (Table 3.23).

Table 3.23. Emission factors for heavy metals from tyre wear. g/mill. Km

| | As | Cd | Cu | Cr | Pb |
|---------------------|-----------|-----------|-----------|-----------|-----------|
| Private cars | 0.003 | 0.007 | 1.691 | 0.014 | 0.552 |
| Van | 0.005 | 0.009 | 2.205 | 0.018 | 0.720 |
| Heavy duty vehicles | 0.019 | 0.037 | 9.096 | 0.074 | 2.970 |
| MC | 0.002 | 0.003 | 0.845 | 0.007 | 0.276 |

Source: van den Brink (1996)

The emission factor used for the estimation of the emissions of Hg is 0.079 g/ mill. km. This emission factor is derived from a study of heavy metal content in tyres (Bækken 1993) and an estimate of the amount of tyre in Norway in 1993 of 6000 tonnes (Finstad et al. 2001).

PAH

Emission factors for PAH are given in Finstad et al. (2001), but there is no information about how much of the emissions that are emitted to air, and how much that goes to soil and to water. All emissions are therefore supposed to be emitted to air. There is also no PAH profile available, so in lack of other data the same PAH profile as for burning of tyres is used (EPA 1998). PAH emission factors for tyre wear are given in Table 3.24. There are no data available for Benzo(a)pyrene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene. All PAH-4 emissions are assumed to be benzo(b)fluoranthene. The PAH-4 profile is given in table 3.25.

Table 3.24. Emission factors for PAH from tyre wear. g/mill. Km

| | PAH | PAH-4 |
|---------------------|------------|--------------|
| Light duty vehicles | 10.4 | 6 |
| Heavy duty vehicles | 0.1 | 0,035 |

Source: Finstad et al. (2001)

Table 3.25. PAH profile road dust, also used for tyre wear (only PAH-4 is shown)

| | Per cent |
|------------------------|----------|
| Benzo(a)pyrene | .. |
| benzo(b)fluoranthene | 100 |
| benzo(k)fluoranthene | .. |
| indeno(1,2,3-cd)pyrene | .. |

Source: Finstad et al. (2001)

Uncertainties

The calculation of emissions from tyre wear is uncertain. First, the emission factors for particles used are based on international studies and not on Norwegian conditions. There is also uncertainty concerning how much of the particles that are emitted to air. According to a Dutch judgement, all particles emitted to air are PM₁₀. This is however only a judgement, and not based on scientific research.

The heavy metal emission factors are based on the particle emission factors for PM₁₀, and since this factor is uncertain, the heavy metal emission factors will also be uncertain. The content of heavy metals in the particles emitted from tyre wear is based on a Dutch study and can therefore differ from Norwegian conditions and type of tyres used.

Until 2004, different methods for calculating the emissions of heavy metals from tyre wear were used. One method was used for calculating emissions of Pb, Cd and Hg (Finstad et al. 2001) and another for calculating emissions of Cu, Cr and As (Finstad & Rypdal 2003). From 2004, the same method has been used for all the heavy metal components.

Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.8.2 Brake wear

Description

Brake blocks will wear during braking and this generates dust containing various metals. In the inventory, emissions of particles and heavy metals are included from this source.

Method

A Tier 1 method is used to estimate emissions from brake wear as speed-dependency is not accounted for.

Particles

Emissions of particles are calculated based on emission factors recommended by an annual mileage.

Heavy metals

Emissions of lead, copper and chromium are calculated after a method described in SLB (Stockholms luft- och bulleranalys 1998a). The calculations are based on annual brake wear, driven kilometers and the brake blocks' metal content.

Brake wear, private cars and vans

To calculate emissions, brake wear first has to be estimated. It is assumed that private cars change brake blocks every fourth year. The background for this assumption is that private cars, by normal driving, change brake blocks at the front after 30 – 40 thousand kilometers and at the back after 60 – 80 thousand kilometers. A private car drives in average 12 thousand kilometers each year. Assuming that the brake blocks are changed after 60 thousand kilometers, the car will be four years old when blocks first are changed.

The brake blocks at front weigh 0.13-0.15 kg and 0.09-0.11 kg at the back. It is assumed in the calculations that the brake blocks weigh 0.15 kg at the front and 0.11 kg at the back, that the brake blocks are worn 70 per cent before they are changed and that the front and back blocks are changed after 40 and 60 thousand kilometers, respectively. Brake wear per kilometer are given equations (3.4) and (3.5):

$$(3.4) \quad \text{Front brake blocks (private cars): } 0.7 \cdot 4 \cdot 0.15 / 40\,000$$

$$(3.5) \quad \text{Back brake blocks (private cars): } 0.7 \cdot 4 \cdot 0.11 / 60\,000$$

The same method is used for calculating emissions from brake wear for vans and minibuses.

Brake wear, heavy duty vehicles

The number of brake blocks at a heavy duty vehicle varies with both brand and model. It is assumed that each front brake block weighs 2.5 kg and 3.5 kg at the back (Stockholms luft- och bulleranalys 1998a). This means that a truck with four wheels have 12 kg of brake blocks. It is assumed that the blocks are changed after 100 thousand kilometers when the brake blocks are worn 70 per cent.

Metal content

The metal content in the brake blocks for cars have been tested (Stockholms luft- och bulleranalys 1998a). For calculating the emissions from brake blocks, annual brake wear has been multiplied by the metal content. The metal content in the brake blocks in front of the car differs from the content in the brake blocks at the back (Table 3.26). For heavy duty vehicles, the metal content is independent of age or type of brake block.

Table 3.26. Metal content in brake blocks. mg/kg

| | Private cars | | Heavy duty vehicles |
|----|--------------|-------|---------------------|
| | Front | Back | Front and back |
| Cr | 137 | 73.4 | 165 |
| Cu | 117941 | 92198 | 9031 |
| Pb | 9052 | 18655 | 457 |

Source: Stockholms luft- och bulleranalys (1998a); Stockholms luft- och bulleranalys (1998b)

How much of the heavy metal emissions that are emitted to air were investigated by Sternbeck *et al.* (2001). Tunnel experiments showed that approximately 20 per cent of the brake wear emissions were emitted to air. This result is used in the calculations of brake wear emissions.

Activity data

For calculating the emissions of particles, are annual mileage given by the road traffic model, see section 3.2.4.2.

For calculating the emissions of heavy metals, annually driven kilometres are also given by the road traffic model.

Emission factors**Particles**

Emission factors recommended by TNO (Institute of environmental and energy technology 2002), based on different European studies, are used (Table 3.27).

Table 3.27. Particle emission factors for brake wear. kg/mill. km

| | PM _{2.5} | PM ₁₀ | TSP |
|------------------------|-------------------|------------------|-------|
| Private cars (BM1+DM1) | 6 | 6 | 6 |
| Van (BN1+DN1) | 7.5 | 7.5 | 7.5 |
| Heavy duty vehicles | 32.25 | 32.25 | 32.25 |
| MC | 3 | 3 | 3 |

Source: TNO (Institute of environmental and energy technology 2002)

BC

BC is estimated as a fraction of TSP from emission factors depending on the vehicle type, given by IIASA (Kupiainen & Klimont 2004).

Table 3.28. Emission factors for BC from tyre wear in share of TSP. Particles are shown in kg/mill. km

| | TSP | BC |
|---------------------|-------|----|
| Passenger cars | 6 | 1% |
| Light duty vehicles | 7.5 | 1% |
| Heavy duty vehicles | 32.25 | 1% |
| MC | 3 | 1% |

Source: IIASA (Kupiainen & Klimont 2004)

Heavy metals

Emission factors for Cr, Cu and Pb are derived based on the above information and are given in Table 3.29.

Table 3.29. Heavy metal emission factors for brake wear. g/mill. km

| | Private cars and vans | Heavy duty vehicles |
|----|-----------------------|---------------------|
| Cr | 0.36 | 2.77 |
| Cu | 342.33 | 151.72 |
| Pb | 38.16 | 7.68 |

Source: Statistics Norway

Uncertainties

There is high uncertainty in different steps in the emission calculations of heavy metals from brake wear, since many assumptions have been done. For example, there is uncertainty connected to the weight and the metal content of the brake blocks, and to the number of driven kilometres before blocks are changed.

No other major emission components are assumed missing.

Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.9 Automobile road abrasion

NFR 1A3bvii

Last update: 22.12.16

3.2.4.9.1 Description

Asphalt dust is emitted to air while using studded tires. The abrasion layer on asphalt roads can contain approximately 90 per cent stones (rock/minerals) and 5 per cent filler. The rest is bitumen. During studded tyre abrasion, stone materials are worn down to minor particles and will together with detached filler and bitumen whirl up and become airborne. How much dust/particles studded tires generate depends on:

- Weight of the stud
- The road surface resistance against abrasion
- Vehicle velocity
- Share of heavy vehicle
- If the road surface is dry, wet or ice coated

A great share of the dust from studded tyres will bind up to the water film when the road surface is wet. Some of it will however whirl up again when the road surface dries up. This is not included in the calculation.

Bitumen is a mixture of a great number of organic components, including PAH components. The emissions of benzo(b)fluoranthene, from road abrasion are calculated and included in the emission inventory. Calculated emissions of Cd are also included.

PM emissions from road abrasion are declining due to implementation of measures. In the largest cities there is a tax to pay when you drive with studded tyres in the city. This, together with information from the authorities about problems caused by PM, has reduced the numbers of cars with studded tyres both in the cities and all over the country. In addition, the weight of the stud has been reduced and hence also the emissions of PM. Consequently, the emissions are decreasing even though the annual total driving length is increasing. In contrast, emissions from automobile tyre and brake wear are calculated by multiplying the driving length with an emission factor, not taking into account the type of tyres. Since the driving length is increasing, the emissions from this source category also increase.

3.2.4.9.2 Method

A Tier 1 method is used to calculate emissions from road abrasion as speed-dependency is not accounted for.

Particles

PM₁₀

The method is prepared by TI/SINTEF and documented in SFT (1999b). For calculating average emission Q (ton/year) of PM₁₀ formula (3.6) is used:

$$(3.6) \quad Q_{PM10} \text{ (ton/year)} = \sum_{\text{All vehicle categories}} SPS * n * l * m * p * w * \alpha / 10^6$$

SPS: The specific wear of studded tyres (SPS). Gives an estimate of how much of the road surface that is worn off on one road kilometer of a vehicle with studded tyres

n: Number of cars of a vehicle category in the area

l: Annual mileage for a vehicle category in the area

m: Part of the year with studded tyres in the area (between 0 and 1)

p: Share of the vehicle category using studded tyres

w: Correction factor for wet and frozen road surface. In the calculation of w, frozen surface is given 0, wet surface 0.5 and dry surface 1. If the mileage with studded tyres on a wet and frozen surface respectively is v and x, $w = (0.05*v) + (1*(1-v-x))$

α: Share of the road dust in air that is PM₁₀. There is no data for this factor. The share of PM₁₀ on ground is used as a reference. There is very varied data for the size of this factor (Hedalen 1994). Hedalen gives a PM₁₀ share of 3-4 per cent. In the calculations 3 per cent is used as a first estimate. Hedalen (1994) states further that the PM_{2.5} share of total road dust is 0.5-1 per cent.

The road surface has stronger wear resistance on roads with heavy traffic than on roads with little traffic. The SPS value can therefore vary with the amount of traffic. SPS values for different ÅDT⁵ intervals were estimated based on analysis of track depths over the years 1988-1995 (Norwegian public roads administration 1996).

SPS is also dependent on the weight of the studs. The studs have in the recent years become lighter. The requirement in 1988 was that the stud on light vehicles should not exceed 2.0 gram, in 1990 this was changed to 1.8 gram, and it changed again in 1992 to 1.1 gram (Norwegian public roads administration 1997). The so-called "light studs" has a weight on 0.7 gram. Studs used on tyres for heavy vehicles could until 1992 weigh 8.0 gram, but this demand was changed to 3.0 gram. There are also other factors influencing the SPS values, for example the road surface wear resistance and the quality of the stone materials used.

SPS values used in the calculations are given in Table 3.30. The SPS values are divided on classes of ÅDT (Evensen, *pers. comm.*⁶). Values are given for 1993-1997 and a prediction for 2002. For the years in between, a moving average is calculated. For the years after 2002, the 2002 SPS values are used. In the calculations average values for SPS, weighted after the size of traffic load

⁵ ÅDT = Average annual daily traffic

⁶ Evensen, R. (2007). Note for Johnny Johansen. 14/12 1997. Bærum: ViaNova.

on roads with different ÅDT, are used. The values are given in g/km and are valid for all vehicles. To estimate how much of the emissions that originate from heavy vehicles, it is provided that heavy vehicles wear 5 times more than light vehicles. The vehicle velocity is not given as an own factor, since it is included in the calculation of SPS.

Table 3.30. SPS values. g/km

| ÅDT | 1973-1980 | 1981-1987 | 1988-1992 | 1993-1997 | 2002- |
|----------------------|-----------|-----------|-----------|-----------|-------|
| 0-1500 | 22 | 20 | 20 | 18 | 16 |
| 1500-3000 | 20 | 20 | 18 | 16 | 14 |
| 3000-5000 | 16 | 15 | 14 | 12 | 10 |
| >5000 | 14 | 12 | 11 | 10 | 9 |
| Average ¹ | 17.1 | 15.6 | 14.7 | 13.1 | 11.6 |

¹ Weight after traffic load on roads with different ÅDT.

Source: Evensen, pers.comm.⁶

Annual traffic load (trafikkarbeid) ($n \cdot l$ in the formula) used in the calculations are annual mileage given by the road traffic model, see section 3.2.4.2.

Use of studded tyres is forbidden in Norway from the first Monday after Easter and until 31st of October. There is an exception from this rule in the three northern counties, Nordland, Troms and Finnmark. In these counties, use of studded tyres is forbidden between 1st of May to 15th of October. It is assumed in the calculations that studded tyres are used the whole period when it is allowed. This means that m is 6.5/12 in the northern counties and 5.5/12 for rest of the country.

Shares of traffic load on studded tyres in the five largest towns in Norway are given in Table 3.31. There has been a decrease in use of studded tyres in Norway during the latest years. The factor p in the formula will therefore vary from one year to another. Information regarding the share of studded tyres originates from the Norwegian Public Roads Administration. There is also national data on share of the car fleet with studded tyres. The data material is based on interviews of car drivers (Norwegian public roads administration 1995a; Norwegian public roads administration 1995b; Norwegian public roads administration 1998). The questionnaires were given out at daytime and caused that most of the answers were from local car drivers. Accordingly, the survey included too many car drivers with annual mileage over 20 000 km. The survey from 1997 was however done differently. In the calculation program, the studded tyre share was decided to be 0.2. This value was adjusted by the different local road administrations, based on interviews or other available knowledge. In 2000, the Norwegian Public Roads Administration made a new investigation over local use of studded tyre (Johansen & Amundsen 2000). In 2006, Gjensidige made a survey over the use of studded tyres in different counties in Norway, winter 05/06 (Vaaje 2006). For 2001-2004 averages of the two investigations are calculated for the counties. For the five largest cities, data from the Norwegian Public Roads Administration was used also for 2001-2005, but for the rest of the country the results from Gjensidige (Vaaje 2006) was used. The data are given in Table 3.32. For the period 1973-1990 it is assumed that the studded tyre share was 90 per cent.

Table 3.31. Use of studded tyres in five prioritized communities. Share of traffic load with studded tyres.
Light duty vehicles

| | 1998/ 1999 | 1999/ 2000 | 2000/ 2001 | 2001/ 2002 | 2002/ 2003 | 2003/ 2004 | 2004/ 2005 | |
|-----------|---------------|---------------|---------------|---------------|---------------|---------------|---------------|---------------|
| Oslo | 51.9 | 32.4 | 21.2 | 31.3 | 29.2 | 28.4 | 24.0 | |
| Drammen | 49.6 | 48.7 | 52.1 | 41.8 | 42.3 | 40.6 | 31.5 | |
| Stavanger | 38.1 | 31.3 | 26.8 | 29.3 | 28.8 | 35.2 | 30.1 | |
| Bergen | 37.0 | 29.4 | 28.3 | 31.0 | 30.7 | 30.4 | 30.3 | |
| Trondheim | 67.0 | 64.4 | 62.1 | 44.4 | 40.2 | 38.8 | 38.1 | |
| | 2005/ 2006 | 2006/ 2007 | 2007/ 2008 | 2008/ 2009 | 2010/ 2011 | 2011/ 2012 | 2012/ 2013 | 2013/ 2014 |
| Oslo | 19.9 | 20.3 | 17.0 | 16.4 | 14.4 | 16.1 | 15.2 | 15.2 |
| Drammen | 27.0 | 28.0 | 27.3 | 22.9 | 25.2 | 25.0 | 20.6 | 20.6 |
| Stavanger | 32.2 | 28.4 | 33.2 | 19.6 | 27.9 | 28.9 | 26.8 | 26.8 |
| Bergen | 29.6 | 21.4 | 10.5 | 14.7 | 12.3 | 18.0 | 16.6 | 16.6 |
| Trondheim | 32.9 | 31.2 | 19.4 | 28.6 | 25.8 | 28.4 | 35.3 | 35.3 |
| | 2014/ 2015 | 2015/ 2016 | 2016/ 2017 | 2017/ 2018 | 2018/ 2019 | | | |
| Oslo | 15 | 14 | 12 | 9 | 9 | | | |
| Drammen | 20 | 20 | 19 | 16 | 17 | | | |
| Stavanger | 22 | 24 | 25 | 14 | 15 | | | |
| Bergen | 14 | 13 | 13 | 14 | 12 | | | |
| Trondheim | 36 | 36 | 34 | 28 | 26 | | | |

Source: The Norwegian Public Roads Administration

Table 3.32. Averaged studded tyre share in Norway weighted by traffic load in the different counties. Light duty vehicles

| Year | |
|------|------|
| 1990 | 0.90 |
| 1991 | 0.87 |
| 1992 | 0.88 |
| 1993 | 0.88 |
| 1994 | 0.87 |
| 1995 | 0.86 |
| 1996 | 0.83 |
| 1997 | 0.79 |
| 1998 | 0.70 |
| 1999 | 0.63 |
| 2000 | 0.58 |
| 2001 | 0.56 |
| 2002 | 0.55 |
| 2003 | 0.53 |
| 2004 | 0.51 |
| 2005 | 0.49 |
| 2006 | 0.48 |
| 2007 | 0.46 |
| 2008 | 0.45 |
| 2009 | 0.44 |
| 2010 | 0.43 |
| 2011 | 0.43 |
| 2012 | 0.42 |
| 2013 | 0.41 |
| 2014 | 0.40 |
| 2015 | 0.39 |
| 2016 | 0.38 |
| 2017 | 0.37 |
| 2018 | 0.36 |

Source: Statistics Norway based on data from the Norwegian Public Roads Administration and Gjensidige

To calculate the correction factor for humid road surface, traffic load data is used. This is divided into different road conditions after Evensen (*pers. comm.*⁷) (Table 3.33). Share of wet and dry road surface will change some as a consequence of varied share of studded tyres. In the calculations for 1990-1997 a correction factor is used, based on the estimation that 80 per cent of light duty vehicles and 60 per cent of heavy duty vehicles use studded tyres.

⁷ Evensen, R. (1997) Personal information, telephone call 20/11 1997. Bærum: ViaNova.

Table 3.33. Grouping of wet, dry and icy road surface

| | In the Norwegian emission inventory |
|-------------|--|
| Wet | Wet |
| Dry | Dry |
| Slush | Wet |
| Loose snow | Wet ¹ |
| Hard snow | Hard snow/ice |
| Bare tracks | 80 per cent dry and 20 per cent wet ² |

¹ Assumption made of NILU and Statistics Norway.

² Assumption made by Evensen (pers. comm.,7).

Source: Statistics Norway

TSP

Hedalen and Myran (1994) analysed road dust depots from Trondheim and found that 30 weight percentage of the particles were below PM₁₀. This gives a distribution where PM₁₀ is 0.3*TSP. This distribution is used in the inventory.

Cd

Emissions of Cd are calculated based on emission factors from Bækken (1993) and annually generated road dust of PM₁₀.

Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4)

Emissions of PAH are calculated based on emission factors from Larssen (1985) and annually generated road dust of PM₁₀.

3.2.4.9.3 Activity data

Cd and PAH

The activity data used for calculating the emissions of Cd and PAH are annually generated PM₁₀ of road dust, see section 3.2.4.9.2.

3.2.4.9.4 Emission factors

Particles

The emission factors can be derived from the factors given under section 3.2.4.9.2. The emission figures are calculated as a product of SPS values for the given year, the number of kilometers driven, part of the cars with studded tyres, part of the year with winter season, correction for icy surface and the PM₁₀ share of the emission (α). The emission factors do not reflect the whirl up of road dust. Heavy duty vehicles whirl up much more than light duty vehicles.

BC

BC is estimated as a fraction of PM_{2.5}. Emission factors depend on the type of vehicle. IIASA (Kupiainen & Klimont 2004) gives emission factors for Black Carbon and Organic Carbon as share of TSP. Since the sum of emissions of BC and OC has to be lower than PM_{2.5} emissions, the emission factors have been adjusted.

Table 3.34. Emission factors for BC from tyre wear in share of TSP

| | BC |
|---------------------|-------|
| Light duty vehicles | 0.83% |
| Heavy duty vehicles | 0.83% |

Source: IIASA (Kupiainen & Klimont 2004)

Cd

The Cd content in the bitumen is uncertain. According to Bækken (1993), the Cd content varies between 1.9 and 43 g Cd per tonne road dust. Statistics Norway has chosen an average emission factor of 22.5 g/ton, see Table 3.35.

Table 3.35. PAH and Cd emission factors from road dust¹. g/tonne. PM₁₀ of road dust

| | Emission factor (g/tonne PM ₁₀ from road dust) |
|--------------------------------|---|
| Norwegian standard (PAH-total) | 61.7 |
| Benzo(a)pyrene, | .. |
| benzo(b)fluoranthene | 5,5 |
| benzo(k)fluoranthene | .. |
| indeno(1,2,3-cd)pyrene, | .. |
| Cd | 22.5 |

¹ Dry road surface.

Source: Finstad et al. (2001)

PAH

The PAH content in the bitumen is uncertain and can vary over time. According to Larssen (1985), the PAH content in airborne dust from wet roads is 330 ppm and 75 ppm from dry roads. Statistics Norway has chosen 85 ppm. In Table 3.35, the emission factor of 85 g/ton is converted to correspond to the PAH components included in NS9815. This gives an emission factor of 61.7 g/ton for PAH-total.

3.2.4.9.5 Uncertainties

Particle distribution of road dust has also been investigated by others than Hedalen and Myran, among them the Norwegian Institute for Air Research (NILU). The results from these measurements show another distribution than Hedalen and Myran, with a PM₁₀-fraction much lower than 30 weight percentage. In the calculation of PM₁₀, data from Hedalen and Myran (1994) are used, and for consistency reasons the same source is used for estimating TSP, despite the uncertainty and the discrepancy with NILU's estimations.

The value of α (PM₁₀ share in road dust) is very uncertain. An average velocity is assumed in the calculations. This is further complicated when road surface on roads with high velocities have another wear resistance than other road surfaces.

The emission factor used for calculating Cd emissions is uncertain since it is based on two measurements.

The estimation of the PAH content in road dust from Larssen (1985) is very uncertain, since it is based on only one measurement in Oslo, but it is the only estimate available, and is used in lack of other data.

3.2.4.9.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

3.2.5 Other sectors

NFR 1A4/1A5

Last update: 10.03.15

3.2.5.1 Description

The source category "Other sectors" includes *all* military combustion, *stationary* combustion in agriculture, forestry, fishing, commercial and institutional sectors and households, motorized equipment and snow scooters in agriculture and forestry, and ships and boats in fishing.

3.2.5.2 Activity data

Motorized equipment is described in section 3.2.4.7.

Households

Use of wood in households for the years 2005-2011 are based on responses to questions relating to wood-burning in Statistics Norway's Travel and Holiday Survey. The figures in the survey refer to quantities of wood *used* (given in tonne). The survey quarterly gathers data that cover the preceding twelve months. The figure used in the emission calculations is the average of five quarterly surveys. For the years before 2005 and for 2012, figures are based on the amount of wood burned from the annual survey on consumer expenditure (Statistics Norway). For the years after 2012, the figures are again based on Statistics Norway's Travel and Holiday Survey, but now only data from three quarterly surveys.

The statistics cover purchase in physical units and estimates for self-harvest. For this survey, the figures refer to quantities *acquired*, which not necessarily correspond to *use*. The survey gathers monthly data that cover the preceding twelve months; the figure used in the emission calculations (taken from the energy accounts), is the average of the survey figures from the year in question and the following year. It is assumed that all combustion of wood in private households takes place in small ovens.

Figures on use of coal and coal coke are derived from information from the main importer. Formerly, Norway's only coal producing company had figures on coal sold for residential heating in Norway. From about 2000, this sale was replaced by imports from abroad. Figures for LPG are collected from the domestic suppliers. Heavy fuel oil is taken from the sales statistics for petroleum products. As the consumption of each energy carrier shall balance against the total sales in the sales statistics, use of fuel oil, kerosene and heavy distillates in households is given as the residual after consumption in all other sectors has been assessed. Use of natural gas is based on sales figures reported to Statistics Norway from the distributors.

Agriculture

Data on energy use in hothouses are collected in surveys performed regularly by Statistics Norway. Sales figures are used to project the figures for consumption of oil products in the years between. For biofuels and LPG, figures are interpolated for years not included in surveys. The

Agricultural Budgeting Board has figures on the use of gasoline, auto diesel and fuel oil in agriculture excluding hothouses. A figure on the minor use of coal was previously collected annually from the only consumer. Since 2002, however, there has been no known use of coal in the Norwegian agricultural activities. Use of natural gas in agriculture, which has increased considerably since it was first registered in 2003, is based on sales figures reported to Statistics Norway from the distributors.

Fishing

Figures on the use of marine gas fuel, heavy distillate and heavy fuel oil are identical with the registered sales to fishing in the sales statistics for petroleum products. In addition to these figures on use in large fishing vessels, a minor figure on estimated use of gasoline in small fishing boats is also included.

Commercial and institutional sectors

Figures on energy use in wholesale and retail trade, and hotels and restaurants, are based on a survey for 2000, performed by Statistics Norway. For the following years, figures from this survey have been adjusted proportionally to the development in employment in the industries in question. For earlier years, the figures are based on a survey from the mid-1980s (Sagen 1987). LPG figures for the whole period from 1990 have, however, been estimated separately after consultation with an oil company.

For most other commercial and institutional sectors, the total use of fuel oil appears as a residual after the use in all other sectors has been estimated; the distribution of this residual between sub-sectors is done by using figures on energy use per man-labour year from the energy survey from the mid-1980s.

Use of heating kerosene in commercial industries is calculated by projecting a figure on use from the mid-1980s proportionally with the registered sales to buildings in industrial industries outside the manufacturing industries. The estimated total amount is distributed between sub-sectors by using figures on energy use per man-labour year from the mid-1980s survey.

Use of natural gas is based on sales figures reported to Statistics Norway from the distributors.

Calculated emissions from combustion of biogas at a sewage treatment plant are included for all years since 1993.

Military

Figures on fuel oil are annually collected directly from the military administration, while figures from the sales statistics for petroleum products are used for other energy carriers.

3.2.5.3 Emission factor

Emission factors used are listed by Statistics Norway, see link in appendix B.

Emission factors for fuelwood are based on data for different oven technologies. Ovens made in 1998 and later have significantly improved combustion and reduced emissions. The factors are

weighted based on information from the surveys of the amount of wood burned in ovens with the different technologies. The yearly weighted factors are listed by Statistics Norway, see link in appendix B.

The country specific emission factor for PAH-4 is split into benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene based on information from Guidebook 2013, chapter 1A4.

Table 3.36. Emission factors for fuelwood, g/kg dry matter, BC in share of PM_{2.5} emissions

| | Open fireplaces | Ovens -1997 | Ovens 1998- |
|---------------------------------|----------------------------|------------------------|------------------------|
| NO _x | 1.3 | 0.97 | 0.97 |
| CO | 126.3 | 150 | 50.5 |
| TSP | 17.3 | 22.7 | 13.4 |
| TSP large cities | 17.3 | 17.4 | 12.2 |
| PM ₁₀ | 17.0 | 22.2 | 13.1 |
| PM ₁₀ , large cities | 17.0 | 17.1 | 12.0 |
| PM _{2.5} | 16.435 | 21.6 | 12.7 |
| PM _{2.5} large cities | 16.435 | 16.5 | 11.6 |
| BC | 9% of PM _{2.5} | 0.96 | 0.86 |
| BC large cities | 9% of PM _{2.5} | 1.01 | 0.9 |
| PCB | 0.1184 | 0.1184 | 0.0156 |
| Benzo(a)pyrene, | 0.82 | 0.74 | 0.006 |
| benzo(b)fluoranthene | 1.29 | 1.16 | 0.01 |
| benzo(k)fluoranthene | 0.30 | 0.27 | 0.003 |
| indeno(1,2,3-cd)pyrene, | 0.59 | 0.53 | 0.005 |

Source: PAH : Finstad et al. (2001), TSP, PM₁₀, PM_{2.5} and BC from ovens : SINTEF (2013), TSP, PM₁₀, PM_{2.5} and BC from fireplaces: Aasestad (2013), PCB : (Nielsen et al. 2015), other pollutants : Haakonsen and Kvingedal (2001).

3.2.5.4 Uncertainties

Uncertainty in *fishing* is described together with navigation in section 3.2.4.5.5.

The method used for finding the use of fuel oil, kerosene and heavy distillates in households implies a great deal of uncertainty regarding the quality of these figures, particularly for fuel oil, which is the most important of these three energy carriers. Since the late 1990s, it has also been necessary to adjust figures for other sectors in order to get consumption figures for households that look reasonable. Hopefully, new surveys will improve the quality of these figures in the future.

As the total use of the different oil products is defined as equal to the registered sales, use in some sectors are given as a residual. This applies to use of heating kerosene and heavy distillates in households, and total use of fuel oil in commercial and institutional sectors. Accordingly, these quantities must be regarded as uncertain, as they are not based on direct calculations. This uncertainty, however, applies only to the distribution of use between sectors - the total use is defined as equal to registered sales, regardless of changes in stock.

There have been large variations in annual sales of military aviation kerosene; as stock changes are not taken into account, the actual annual use is uncertain.

3.2.5.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

3.2.6 International bunkers

NFR - memo item

Last update: 31.03.06

3.2.6.1 Description

Emissions from international bunkers (marine and aviation) have been estimated and reported separately from national estimates, in accordance with the IPCC Guidelines. Differences between the IEA (International Energy Agency) data and the data reported to UNFCCC in sectoral data for marine shipping and aviation, are due to the fact that different definitions of domestic use are employed. In the Norwegian inventory, domestic consumption is based on a census in accordance with the IPCC good practice guidance. On the other hand, the IEA makes its own assessment with respect to the split between the domestic and the international market.

3.2.6.2 Shipping

3.2.6.2.1 Method

The sales statistics for petroleum products, which is based on reports from the oil companies to Statistics Norway, has figures on sales for bunkers of marine gas oil, heavy distillates and heavy fuel oil. The same emission factors as in the Norwegian national calculations are used.

3.2.6.2.2 Activity data

Sales figures for international sea transport from Statistics Norway's sales statistics for petroleum products are used for marine gas oil, heavy distillates and heavy fuel oil.

3.2.6.2.3 Emission factor

Emission factors used for *Shipping* are described under *Navigation* in section 3.2.4.5.

3.2.6.3 Aviation

3.2.6.3.1 Method

The consumption of aviation bunker fuel in Norway is estimated as the difference between total purchases of jet kerosene in Norway for civil aviation and reported domestic consumption. Figures on total aviation fuel consumption are derived from sales data reported to Statistics Norway from the oil companies. These data do not distinguish between national and international uses. Data on domestic fuel purchase and consumption are therefore collected by Statistics Norway from all airline companies operating domestic traffic in Norway. The figures on domestic consumption from airlines are subtracted from the total sales of jet kerosene to arrive at the total fuel sales for international aviation. The bottom-up approach of Norway is the detailed Tier 2 CORINAIR methodology. The methodology is based on detailed information on types of aircraft and number of LTOs, as well as cruise distances.

3.2.6.3.2 Activity data

Statistics Norway annually collects data on use of fuel from the air traffic companies, including specifications on domestic use and purchases of fuel in Norway and abroad.

3.2.6.3.3 Emission factor

Emission factors used for *Aviation* are described under *Aviation* in section 3.2.4.1.

3.3 Fugitive emissions from fuels

NFR 1B

3.3.1 Overview

Emission sources included in the inventory from the sector *Fugitive emissions from fuels* are fugitive emissions from coal mining and handling, and from oil and natural gas.

Fugitive emissions from oil and natural gas include emissions from loading and refining of oil, gasoline distribution, fugitive emissions from the gas terminals on shore and fugitive emissions in connection with venting and flaring offshore.

3.3.2 Coal mining and handling

NFR 1B1

Last update: 06.06. 18

3.3.2.1 Description

Coal has been shipped from Svalbard since 1907. There are today one coal mine at Spitsbergen (the largest island in the Svalbard archipelago) operated by a Norwegian company. They opened a second mine in 2001. As the Norwegian emission inventory, according to official definitions, shall include emissions from all activities at Svalbard, also emissions from Russian coal production have been estimated. Until 1998, there was production in two Russian coal mines, Barentsburg and Pyramiden, but since then, production takes place only in the Barentsburg mine. The Norwegian mines and Pyramiden are defined as surface mines, whereas Barentsburg is an underground mine.

In 2005 there was a fire in one of the Norwegian coal mines and this caused that the production was almost halved from 2004 to 2005.

Russian production has since 2001 been considerably smaller than the Norwegian production. In 2008 a fire started in the Russian mine at Barentsburg. Shortly after the fire started, the mine was filled with water and hence there were no significant emissions from the fire. This is the reason why emissions from the fire are not estimated. The production in 2008 and 2009 was therefore very small. In autumn 2010, ordinary production was restarted. Russian activity data are more uncertain than the Norwegian, which causes a correspondingly higher uncertainty in the emission figures.

At Svalbard there was a smouldering fire in Pyramiden, the Russian mine that was closed down in 1998. At an inspection in 2005, no emissions were registered, which indicates that the fire had burnt out. Due to lack of data, emissions for earlier years from this fire have not been estimated. However, Norwegian authorities assume that these emissions are limited.

Emissions from NMVOC and particles from handling of coal are included in the inventory.

3.3.2.2 Method

NMVOC

NMVOC emissions from handling of coal are estimated by multiplying the amount of coal extracted (raw coal production) with Tier 2 emission factors from EMEP/EEA Guidebook 2013 (EEA 2013).

Particles

Emissions of particles from handling of coal are estimated by multiplying the amount of coal extracted (raw coal production) with Tier 1 emission factors from EMEP/EEA Guidebook 2016 (EEA 2013).

3.3.2.3 Activity data

Figures on Norwegian production (raw coal production) are reported by the plant to Statistics Norway. Russian figures are reported to the Norwegian authorities on Svalbard; these figures are, however, regarded as highly uncertain, consisting of a mixture of figures on production and shipments.

3.3.2.4 Emission factors

NMVOC

Emission factors for NMVOC are taken from EMEP/EEA Guidebook 2016 (EEA 2013). The Tier 2 factors used are 3 kg NMVOC per tonne coal for surface mines and 0.2 kg NMVOC per tonne coal for underground mines.

Particles

Emission factors for particles are taken from EMEP/EEA Guidebook 2016 (EEA 2013). The same Tier 1 factors are used for both surface and underground mines. The factors are 0.089 kg particles per tonne coal for TSP, 0.042 kg particles per tonne coal for PM₁₀ and 0.005 kg particles per tonne coal for PM_{2.5}.

3.3.2.5 Uncertainties

The uncertainty in the activity data concerning Norwegian coal production is regarded as being low. The uncertainty in Russian data is considerably higher.

3.3.3 Uncontrolled combustion and burning coal dumps

NFR 1B1b

Last update: 07.06.11

3.3.3.1 Description

In 2005, a fire broke out in one of the Norwegian coal mines at Spitsbergen, causing minor emissions.

3.3.3.2 Method

Emissions have been calculated by multiplication of the quantity of coal combusted by standard emission factors for combustion of coal.

3.3.3.3 Activity data

The company operating the mine has provided an estimate on the quantity of coal combusted in the fire.

3.3.3.4 Emission factors

Emission factors for direct-fired furnaces (see link in appendix B to factors listed by Statistics Norway) have been used in the calculations.

Emissions of BC have been estimated using the same share of PM_{2.5} as used for coal burning.

3.3.3.5 Uncertainties

The uncertainty in the activity data, that is the quantity of coal combusted, is unknown. However, as the emissions are small, the uncertainty is insignificant. Emission from fires in Russian coal mines on Svalbard are not included.

3.3.3.6 Source specific QA/QC

There is no specific QA/QC procedure for this source.

3.3.4 Oil and natural gas

NFR 1B2

Last update: 06.03.18

3.3.4.1 Description

1B2a covers emissions from loading and storage of crude oil, refining of oil and distribution of gasoline. Loading, unloading and storage of crude oil on the oil fields off-shore and at oil terminals on-shore cause emissions of NMVOC. Non-combustion emissions from Norway's two oil refineries (a third was closed down in 2000) include NO_x, NMVOC, SO₂ and particulates. Gasoline distribution causes emissions of NMVOC, however, these emissions has decreased due to a reduced demand on gasoline from passenger cars. Especially from 2007 there has been a shift in the consumption of fuels for road traffic from gasoline to auto diesel. This is mainly due to the introduction of a CO₂ tax on passenger cars (PC) from January 1st 2007. This resulted in diesel driven cars becoming less expensive than gasoline driven cars. From 2007, approximately 70 per cent of all new PCs were diesel driven.

1B2b covers fugitive emissions of NMVOC from gas terminals on shore.

1B2c covers fugitive emissions from venting and flaring. Venting emissions include emissions of NMVOC from exploration and production drilling of gas and oil. The major source is cold vent and leakage of NMVOC from production drilling.

Most of the emissions in *1B2c* come from flaring of natural gas offshore (during both well testing, extraction and pipeline transport) and at gas terminals and flaring of refinery gas at the refineries. This flaring causes emissions of NO_x, NMVOC, SO₂, CO, particulates, BC, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins. There is also some flaring of oil in connection with well testing - amounts flared and emissions are reported to NPD (the Norwegian Petroleum Directorate) and the Norwegian Environment Agency.

The major source in sector 1B2 is flaring of natural gas on the Norwegian continental shelf. Table 3.37 gives an overview over the calculations of the fugitive emissions of NMVOC and other gases.

Table 3.37 Fugitive emissions from oil and natural gas. Emission sources, compounds, methods, emission factors and activity data included in the Norwegian GHG Inventory

| B Fugitive emissions from fuels | CO₂ | CH₄ | N₂O | NMVOC | Method | Emission factor | Activity data |
|--|-----------------------|-----------------------|-----------------------|--------------|---------------|------------------------|----------------------|
| 1.B.2.a Oil | | | | | | | |
| i. Exploration | IE | IE | NO | IE | Tier II | CS | PS |
| ii. Production | IE | IE | NO | IE | Tier II | CS | PS |
| iii. Transport | E | R/E | NO | R/E | Tier II | CS | PS |
| iv. Refining/Storage | R/E | R | NO | R | Tier I/II | CS | PS |
| v. Distribution of oil products | E | NE | NO | R/E | Tier I | C/CS | CS/PS |
| vi. Other | NO | NO | NO | NO | | | |
| 1.B.2.b Natural gas | | | | | | | |
| i. Exploration | IE | IE | NO | IE | Tier II | CS | PS |
| ii. Production | IE | IE | NO | IE | Tier II | CS | PS |
| iii. Processing | IE | IE | NO | IE | Tier II | CS | PS |
| iv. Transmission | IE | IE | NO | IE | Tier II | CS | PS |
| v. Distribution | IE | E | NO | IE | Tier II | OTH | CS/PS |
| vi. Other | E | R | NO | R | Tier II | CS | PS |
| 1.B.2.c Venting | | | | | | | |
| i. Oil | IE | IE | NO | IE | Tier II | CS/PS | PS |
| ii. Gas | IE | IE | NO | IE | Tier II | CS/PS | PS |
| iii. Combined | R/E | R/E | NO | R/E | Tier II | CS/PS | PS |
| Flaring | | | | | | | |
| i. Oil (well testing) | R/E | E | E | R/E | Tier II | CS | PS |
| ii. Gas | | | | | | | |
| Gas and oil fields | R/E | R/E | E | R/E | Tier II | CS | PS |
| Gas terminals | R | R | E | R/E | Tier I | CS | CS |
| Refineries | R | R | R/E | E | Tier I | CS | CS |
| iii. Combined | IE | IE | IE | IE | Tier I | CS | CS |

R = emission figures in the national emission inventory are based on figures reported by the plants. E = emission figures are estimated by Statistics Norway (Activity data * emission factor). IE = Included elsewhere, NO = Not occurring, CS = Country specific, PS = Plant specific, Tier = the qualitative level of the methodology used, C=Corinair, OTH=Other.

3.3.4.2 Method

3.3.4.2.1 Loading and storage of crude oil off shore and on shore

NMVOC

Included in the inventory is emission from loading and storage of crude oil produced at the Norwegian continental shelf. This means also those oil fields that is on both the Norwegian and UK continental shelf and is loaded on the Norwegian side of the shelf is included as a whole in the Norwegian inventory and opposite.

For the years 1990-2002 the emissions of CH₄ and NMVOC is calculated by Statistics Norway. The calculation is based on the field specific amounts of crude oil loaded and stored multiplied with field specific emission factors. Field specific activity data and emission factors used in the calculation were annually reported by the field operators to Statistics Norway and the Norwegian Environment Agency (emission factors are only reported to the Norwegian

Environment Agency). Since year 2000 an increasing share of the shuttle tankers have had installed vapor recovery units (VRU), and emissions from loading of crude oil on shuttle tankers with and without VRU are calculated separately for each field. In addition emission figures were annually reported to the Norwegian Environment Agency and used in the QC of the emission figures calculated by Statistics Norway.

From 2003, emission of CH₄ and NMVOC from loading and storage of crude oil on shuttle tankers included in the GHG Inventory are based on reported emission figures from the oil companies. Emissions, activity and emissions factors with and without VRU are reported from each field operator into the database EPIM Environment Hub (EEH), previously Environmental Web. The database is operated by the Norwegian Petroleum Directorate, the Norwegian Environment Agency and ¹The Norwegian Oil Industry Association. The method for calculating the emissions is the same as for 1990-2002.

An agreement was established 25 June 2002 between the Norwegian Pollution Control Authority (now Norwegian Environment Agency) and VOC Industrisamarbeid (a union of oil companies operating on the Norwegian continental shelf) aiming to reduce NMVOC emissions from loading and storage of crude oil off shore. So in addition, also from 2003, the emission of CH₄ and NMVOC from loading and storage of crude oil on shuttle tankers is reported annually to the Norwegian Environment Agency by the "VOC Industrisamarbeid" in the report "VOC Industrisamarbeid. NMVOC reduksjon bøyelasting norsk sokkel" (VOC Cooperation. Reduction of NMVOC from buoy loading on the Norwegian continental shelf). The report include e.g. details of ships buoy loading and which oil fields the oil has been loaded /stored at, amount of oil loaded, EFs with and without VRU. The method for calculating the emissions is the same as for 1990-2002.

Norway considers that the method for calculating the CH₄ and NMVOC emissions from loading and storage of crude oil is consistent for the period 1990-2014.

For the two Norwegian oil terminals on shore, the emissions from loading of crude oil are reported annually from the terminals to the Norwegian Environment Agency. At one of the terminals VRU for recovering NMVOC was installed in 1996. The calculation of the emissions of CH₄ and NMVOC at both terminals is based upon the amount of crude oil loaded and oil specific emission factor dependent of the origin of the crude oil loaded.

3.3.4.2.2 Oil refineries

NO_x, NMVOC, SO₂ and particulates

Emission figures from the oil refineries are reported to the Norwegian Environment Agency, and are after QA/QC procedures used in the emission inventory.

3.3.4.2.3 Gasoline distribution

NMVOC

Emissions from gasoline distribution are calculated from the amounts of gasoline sold from the sale statistics and emission factors for, respectively, loading of tanker at gasoline depot, loading of tanks at gasoline stations, and loading of cars.

3.3.4.2.4 *Gas terminals*

NMVOC

Fugitive emissions of NMVOC from gas terminals are annually reported from the terminals to the Norwegian Environment Agency.

The emissions are calculated based on the number of sealed and leaky equipment units that is recorded through the measuring and maintenance program for reducing the leakage. The number of sealed and leaky equipment units is collected two times a year and the average number of the counting is used in the calculation. It is assumed in the calculation that a leakage has lasted the whole year if not the opposite is documented.

3.3.4.2.5 *Venting*

NMVOC

Emissions of NMVOC from cold venting and diffuse emissions for each field are reported annually to the Norwegian Environment Agency from the field operator. The emissions are mostly calculated by multiplying the amount of gas produced with an emission factor for each emission source identified at the field.

3.3.4.2.6 *Flaring*

NO_x, NMVOC, CO, particulates, BC, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins.

Emissions from flaring of natural gas off shore are calculated by Statistics Norway on the basis of field specific gas consumption data and emission factors. For NO_x, NMVOC and SO₂, calculated emissions are used in the inventory for the years until 2002. From 2003, emissions of these pollutants from flaring offshore have been reported by the oil companies to NPD and the Norwegian Environment Agency are used in the inventory. The same method is used in the calculation of emissions from flaring in connection with well testing.

Emissions of NO_x from flaring at gas terminals are reported for all years. For NMVOC, emissions are calculated for one gas terminal and reported figures used for the others. Other emissions from the gas terminals are based on activity data and emission factors.

3.3.4.3 Activity data

3.3.4.3.1 *Loading and storage of crude oil off shore and on shore*

The amount of oil buoy loaded and oil loaded from storage tankers is reported by the field operators in an annual report to the Norwegian Environment Agency and the Norwegian Petroleum Directorate (NPD). The amount of oil loaded on shuttle tankers with or without VRU is separated in the report.

Before 2003, Statistics Norway gathered data on amounts of crude oil loaded at shuttle tankers and stored at storage vessels from the NPD. The data from each field are reported monthly by the field operators to NPD on both a mass and a volume basis. The allocation of the amount of crude oil loaded at shuttle tankers and stored at storage vessels with or without VRU is from the annual report the field operators are committed to deliver to the Norwegian Environment Agency and NPD.

The amount of oil loaded at on shore oil terminals is also reported to the Norwegian Environment Agency and NPD.

3.3.4.3.2 Oil refineries

The crude oil throughput is annually reported by the plant to the Norwegian Environment Agency.

3.3.4.3.3 Gasoline distribution

Gasoline sold is annually collected in Statistics Norway's sales statistics for petroleum products.

3.3.4.3.4 Gas terminals

Activity data that the terminals use in their emission calculations are sampled through the terminals measuring and maintenance program, whose aim is to reduce leakage.

3.3.4.3.5 Venting

Amounts of gas produced or handled at the platforms are reported to the Norwegian Environment Agency from the operators and used in the QC of the reported emissions.

3.3.4.3.6 Flaring

Amounts of gas flared at offshore oil and gas installations are reported monthly by the operators to the Norwegian Petroleum Directorate (NPD). Amounts flared at the gas terminals are reported to NPD and the Norwegian Environment Agency. Amounts of refinery gas flared are found by distributing the total amounts between different combustion technologies by using an old distribution key, based on data collected from the refineries in the early 1990s. This distribution was confirmed in 2003.

3.3.4.4 Emission factors

3.3.4.4.1 Loading and storage of crude oil offshore and on shore

For the years before 2003, emission factors used in the calculation of NMVOC emissions offshore are field specific and were reported to the Norwegian Environment Agency and NPD in an annual report. The Norwegian Environment Agency forwarded the emission factors to Statistics Norway. From 2003, the emission figures reported by the field operators are used in the inventory.

The evaporation rate varies from field to field and over time, and the emission factors are dependent on the composition of the crude oil as indicated by density and Reid vapour pressure (RVP). The VOC evaporation emission factors are obtained from measurements, which include emissions from loading and washing of shuttle tankers. For some fields the emission factors are not measured, only estimated. The CH₄ content of the VOC evaporated is also measured so that total emissions of VOC are split between CH₄ and NMVOC.

The emission factors that the field operators use in their calculations are reported to the Norwegian Environment Agency and NPD. They report emission factors with and without VRU and the split between CH₄ and NMVOC.

Loading on shore: The emission factors are considerably lower at one of Norway's two oil terminals than at the other, because the oil is transported by ship and therefore the lightest fractions have already evaporated. At the other terminal the oil is delivered by pipeline. The latter terminal has installed VRU, which may reduce NMVOC emissions from loading of ships at the terminal by about 90 per cent. NMVOC emissions at this terminal are estimated to be more than 50 per cent lower than they would have been without VRU. However, the VRU technology is not designed to reduce methane and ethane emissions.

3.3.4.4.2 Oil refineries

The emission factor used in the calculation of NMVOC and methane emissions from the largest refinery is based upon measurements using DIAL (Differential absorption LIDAR). A new measurement program was initiated in 2009 and followed up with new measurements later. An annual EF is deduced from the measured NMVOC and methane emissions and the crude oil throughput. The average EF for the period 2009-2013 is used for the years before the current program was initiated, *i.e.* 1990-2008.

BC emissions have been estimated as a fraction of the PM_{2.5}. IIASA (Kupiainen & Klimont 2004) gives a fraction of 0.16 per cent.

3.3.4.4.3 Gasoline distribution

The emission factor for NMVOC from refuelling of gasoline in cars (1.48 kg NMVOC/tonne gasoline) is taken from inventory guidebook EMEP/EEA Guidebook 2016 (EEA 2013).

3.3.4.4.4 Venting

The emission factors used are listed in Table 3.38.

Table 3.38. Emission factors for cold vents and leakage at gas fields offshore

| | NMVOC Emission factor | Calculation method |
|--|--------------------------|--------------------|
| Emission source | [g/Sm ³] | |
| Glycol regeneration | 0.065 | |
| Gas dissolved in liquid from K.O. Drum | 0.004 | |
| Gas from produced water system | 0.03 | |
| Seal oil systems | 0.015 | |
| Leaks through dry compressor gaskets | 0.0014 | |
| Start gas for turbines ¹ | 0.4 | Tonne per start up |
| Depressurisation of equipment | 0.005 | |
| Instrument flushing and sampling | 0.00021 | |
| Purge and blanket gas ¹ | 0.032 | |
| Extinguished flare | 0.014 | |
| Leaks in process | 0.007 | |
| Depressurisation of annulus | 0.0000005 | |
| Drilling | 0.55 | Tonne per well |

¹ The gas source is standard fuel gas.

Source: Aker Engineering (1992)

3.3.4.4.5 Flaring

NO_x: A NO_x emission factor at 1.4 g NO_x/Sm³ flared gas at off shore installations is based upon studies conducted by Bakken et al. (2008). In the study two new experimental laws have been

compared with DIAL-measurements of NO_x emissions made on onshore flares.

PM₁₀: The emission factor is based on McEwen and Johnson (2011). In fig. 7, this paper gives a regression formula for the emission factor as a function of the heating value (GCV) as $EF = 0.0578(HV) - 2,09$. For Norwegian offshore flaring a heating value of 48 MJ/Sm³ is suggested in Bakken et al. (2008). This gives an emission factor of 0.856 g PM₁₀/Sm³.

BC: Emissions are estimated using the same methodology as PM₁₀ emissions. The regression formula for the BC emission factor, given as a function of the heating value (GCV) is $EF = 0.0578(HV) - 2,09$. This gives an emission factor of 0.684 g BC/Sm³.

Other emission factors from flaring of gas are listed by Statistics Norway, see link in appendix B. The same factors are used for flaring of gas in connection with *well testing*. For flaring of *oil*, the emission factors are shown in Table 3.39.

Table 3.39. Emission factors for flaring of oil in connection with well testing

| Compounds (unit) | unit/tonnes flared oil | Source |
|----------------------------|------------------------|--|
| NO _x (tonnes) | 0.0037 | (OLF 2009) |
| NMVOC (tonnes) | 0.0033 | |
| CO (tonnes) | 0.018 | |
| TSP (tonnes) | 0.025 | Measurements (OLF ¹) |
| PM ₁₀ (tonnes) | 0.0215 | Use the same distribution as for combustion of heavy fuel oil in industry (EPA 2002) |
| PM _{2.5} (tonnes) | 0.014 | |
| PAH (kg) | 0.012 | (OLF 1991) |
| Benzo(a)pyrene, | 0.00024 | Use the same distribution as for combustion of heavy fuel oil in industry (EPA 1998) |
| benzo(b)fluoranthene | .. | |
| benzo(k)fluoranthene | .. | |
| indeno(1,2,3-cd)pyrene, | .. | |
| Dioxins (mg) | 0.01 | Measurements (OLF) |
| PCB (mg) | 220 | Langøren and Malvik (2010) |

¹The Norwegian Oil Industry Association (OLF) now Norwegian Oil and Gas Association.

3.3.4.5 Uncertainties

The uncertainty in the emission factors for NMVOC (Rypdal & Zhang 2001) from *oil loading* is estimated to be ± 40 per cent and in the activity data ± 3 per cent.

The uncertainty in the amount of gas flared is regarded as being low, ± 1.4 per cent, based on data reported in the emission trading scheme (Climate and Pollution Agency 2011) and assumptions in Rypdal and Zhang (2000). The uncertainty in NMVOC emissions from venting is much higher than for flaring.

The emission factors for both storage and transmission of natural gas are uncertain, since Austrian factors are used in lack of country specific Norwegian factors.

All uncertainty estimates for this source are given in Appendix C.

3.3.4.6 Source-specific QA/QC and verification

Statistics Norway gathers activity data on oil and gas activities from the Norwegian Petroleum Directorate (NPD). These data are reported monthly by the field operators to NPD. The activity data are quality controlled by comparing them with the figures reported in the field operator's annual report to the Norwegian Environment Agency and NPD. The emissions calculated by Statistics Norway for 1990-2002 were compared with the emission data that the field operators reported to the Norwegian Environment Agency and NPD. From 2003, Statistics Norway estimate emissions based on activity data that the field operators monthly report to NPD, and reported emission factors. When discrepancies are found between the two sets of data these are investigated and corrections are made if appropriate. If errors are found, the Norwegian Environment Agency contacts the plant to discuss the reported data and changes are made if necessary.

The reported emissions from the gas terminals are compared with previous years' emissions.

The Norwegian Environment Agency collects the activity data used for venting and flaring in the calculation from the NPD. The figures are quality controlled by comparing them with the figures reported in the field operators' annual report to the Norwegian Environment Agency and NPD, and time series are checked.

The Norwegian Environment Agency perform internal checks of the reported data for venting from the field operators. Some errors in the time-series are usually found and the field operators are contacted and changes are made. The same procedure is followed to check the amount of gas reported as flared. The quality of the activity data is considered to be high, due to the fact that there is a tax on gas flared offshore. NPD has a thorough control of the amount of gas reported as flared.

4 INDUSTRIAL PROCESSES AND PRODUCT USE (NFR sector 2)

NFR 2

4.1 Overview

This chapter provides descriptions of the methodologies employed to calculate emissions of greenhouse gases and long-range transboundary air pollutants from industrial processes and product use. Only non-combustion emissions are included in this chapter. Emissions from fuel combustion in the manufacturing industries are reported in the Energy chapter. Emission figures are either reported by plants to the Norwegian Environment Agency or calculated by Statistics Norway, based on emission factors and activity data. The emission factors are collected from different sources, while the activity data used in calculations carried out by Statistics Norway mainly come from official statistics collected by Statistics Norway.

4.2 Mineral products

NFR 2A

Last update: 25.02.19

The sector category Mineral products in the Norwegian inventory includes emissions from fourteen different products. SO₂, NO_x, NH₃, particles, BC, heavy metals dioxins, PAHs benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and HCB are components that are emitted during the production of mineral products and included in the inventory.

Table 4.1 shows the various components emitted from the different activities, and for which components the emission figures in the national inventory are based on figures reported by the plants and for which the figures are estimated by Statistics Norway.

Table 4.1. Mineral products. Components emitted and included in the Norwegian inventory¹

| Mineral products | SO ₂ | NO _x | NH ₃ | Particles | BC | Heavy metals | Dioxins | PAH | HCB |
|---|-----------------|-----------------|-----------------|-----------|----|--------------|---------|-----|-----|
| Cement production | R | NE | NE | R | E | R | R | NE | R |
| Lime production | NE | NE | NA | R | E | R | NA | NA | R |
| Glass and glassfibre production | NE | R | R | R | E | R | NA | NE | NE |
| Mining and extraction of stones and minerals | NA | NA | NA | R | NA | NA | NA | NA | NA |
| Construction and demolition | NA | NA | NA | E | NA | NA | NA | NA | NA |
| Ceramics | NA | NA | NA | R | E | NA | NA | NA | NE |
| Non-metallurgical magnesia production | R | NA | NA | R | E | NA | R | NA | NE |
| Sandpit and rock-crushing plants | NA | NA | NA | E | NA | NA | NA | NA | NA |
| Concrete pumice stone | R | NA | NA | R | E | NA | NA | NA | NA |
| Rock wool production | NA | R | R | R | E | R | NA | NA | NE |
| Production of mineral white | NA | NA | NA | R | E | R | NA | NA | NA |
| Construction /repairing of vessels - Sandblasting | NA | NA | NA | R | NA | NA | NA | NA | NA |

¹ R means that emission figures in the national emission inventory are based on figures reported by the plants. E means that the figures are estimated by Statistics Norway (Activity data * emission factor). NA = Not Applicable. NE = Not Estimated.

Source: Statistics Norway/Norwegian Environment Agency

4.2.1 Cement production

NFR 2A1

Last update: 31.01.20

4.2.1.1 Description

Two plants in Norway produce cement. Production of cement gives rise to both non-combustion and combustion emissions of SO₂. The emission from combustion is reported in chapter 3 Energy. The non-combustion emissions originate from the raw material calcium carbonate (CaCO₃). The resulting calcium oxide (CaO) is heated to form clinker and then crushed to form cement. The emissions of SO₂ from non-combustion are reported to The Norwegian Environment Agency.

SO₂ from cement production is emitted from sulphur in the fuel (reported under Energy) and in the raw materials, especially pyrite in limestone. Only the SO₂ from the raw materials should be counted as non-combustion emissions. Particles as well as heavy metals are emitted during the production process. More than 90 per cent of the emission of mercury is due to mercury in the limestone, while the emissions of Pb, Cd, Cu, Cr and As originate both from processes and combustion of fuel. Emissions of dioxins are due to the thermal process in the clinker production.

4.2.1.2 Method

SO₂

The plants annually report emissions of SO₂ to the Norwegian Environment Agency. Figures are based on measurements at the plants.

SO₂ emissions from production of cement come from energy carriers like e.g. coal and oil and from limestone. The sulphur from the energy carriers is to a large extent included in the clinker during the process. The emissions are distributed between combustion and non-combustion

emissions based on studies conducted by Institute for Energy Technology in 1970 and 1999. Both studies indicate that 80-99 per cent of the sulphur from energy carriers is included in the clinker.

The total SO₂ emissions from the two plants are based on measurements. When the SO₂ emissions reported from the plant are not distributed between combustion and non-combustion emissions, the Norwegian Environment Agency distributes the total emissions, using the same percentage distribution as in the last year with reported distributed SO₂ emissions. The production technology is to some extent different for the two plants. In the last years, the distribution between combustion and non-combustion emissions is about 10/90 for one plant and 18/82 for the other plant. The difference is assumed to be due to the fact that one plant has a "by-pass" system where some of the flue gas is not in contact with the raw materials.

The amount of energy carriers used in cement production is subtracted from the energy balance to avoid double counting, see section 3.2.1.2.

Particles

Emissions have been reported to the Norwegian Environment Agency since 1991 for one plant and since 1992 for the other. It is believed that the reported figures also include emissions from combustion. Therefore, emissions from combustion of coal, coke and waste oil used in cement production are not calculated, to avoid double counting. The plants have installed particle filter.

Particle size distribution for emitted particles from cement production is found in EEA (2019). PM₁₀ and PM_{2.5} are assumed to be 90 and 50 per cent of TSP, respectively.

BC

Emissions have been estimated from a share of PM₁₀ emissions given by EEA (2019). As a share of PM₁₀, BC emission factor is 0.3 per cent.

Heavy metals and POPs

Emission figures for heavy metals are reported to the Norwegian Environment Agency. It is believed that these figures also include emissions from combustion. Therefore emissions from combustion of coal, coke and waste oil used in cement production are not calculated, to avoid double counting.

Dioxin figures are reported to the Norwegian Environment Agency. It is also here assumed that the reported figures include emissions from fuel combustion, therefore emissions from combustion are not calculated.

HCB emissions were reported by the plant to the Norwegian Environment Agency in 2010. This reporting has been used to estimate emissions in 2010. For other years in the time series, emissions have been estimated using a Tier 2 emission factor from the EEA (2016).

4.2.1.3 Activity data

The activity data is the production of clinker and this is reported annually to the Norwegian

Environment Agency. These are reported in the NFR table and in Appendix F.

4.2.1.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Reported emission figures for particles have varied a great deal as a result of changes the plants have undergone to reduce emissions. There are also uncertain measurements due to annual variations.

Regarding the heavy metals, it has varied when the two plants started reporting the various components, and therefore estimations have been necessary for the years when reporting have been insufficient. The reported figures also vary from a year to another due to process technical conditions, variations in the metal content in the limestone used, and uncertain measurements.

4.2.1.5 Source specific QA/QC

The emissions are reported according to their permits to the Norwegian Environment Agency (NEA). The case handler assesses the reported emissions and activity data and contacts the plants if needed. Emissions from these plants also include GHGs (green house gases), so the inventory team at NEA undertakes QA/QC procedures that also are relevant for other components. See section 1.6 for the description of the general QA/QC procedure.

4.2.2 Lime production

NFR 2A2

Last update: 13.02.20

4.2.2.1 Description

Emissions of particles, black carbon and HCB from lime production are included in the Norwegian inventory.

4.2.2.2 Method

Particles

One plant has reported emission figures for particulate matter to the Norwegian Environment Agency since 1990. Emission figures from 1990 to 1995 are based on calculations, using emission factors and production volume. Since 1996, the figures are a result of measurements at the plant. The plant has installed particle filter.

In the inventory, a particle size distribution suggested by EEA (2019). PM₁₀ and PM_{2.5} are assumed to be 38.9 and 7.78 per cent of TSP, respectively.

BC

For the same plant that reports particles, BC emissions have been estimated from a share of TSP emissions given by EEA (2019). As a share of TSP, BC emission factor is 0.46 per cent.

HCB

HCB may unintentionally be formed in the production and extracation of lime in the thermic

process. One plant has reported emissions in 2010. Emissions for the rest of the timeseries are estimated based on lime production data. Emissions for two other plants that do not report emissions are also estimated based on lime production. The emission factor used is 0.008 mg HCB per tonne lime from Japan (Toda 2006). It is also assumed that, for this category, the reported figures include emissions from fuel combustion, therefore emissions from combustion are not calculated.

4.2.2.3 Activity data

The activity data is the production of lime and this is reported annually to the Norwegian Environment Agency. These are reported in the NFR table and in Appendix F.

4.2.2.4 Uncertainties

The particle distribution used is not specified for the plants, and the particles emitted might therefore have another distribution than the one suggested from EEA (2019).

4.2.2.5 Source specific QA/QC

The emissions are reported according to their permits to NEA. The case handler assesses the reported emissions and activity data and contacts the plants if needed. Emissions from these plants also include GHGs (green house gases), so the inventory team at NEA undertakes QA/QC procedures that also are relevant for other components. See section 1.6 for the description of the general QA/QC procedure.

4.2.3 Glass and glassfibre production

NFR 2A3

Last update: 17.01.20

4.2.3.1 Description

Five plants producing glass, glass wool or glass fibre are included in the emission inventory, with figures based on emission reports to the Norwegian Environment Agency. PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxin emissions are neither calculated nor measured. Production of glass can be a source for dioxin emissions, but no reported figures are available. Emission factors are found in literature, but since activity data (production rate) is not available and it is assumed that the emission factor is dependent on type of glass produced, emissions are not calculated.

Emissions of particles are also reported from three other glass-producers in Norway, but since annual emissions are low (less than 1 tonne), they are not included in the inventory

4.2.3.2 Method

NO_x

The two glass wool producing plants and the one producing glass fibre annually report emission figures for NO_x to the Norwegian Environment Agency. The emission figures are based on calculations.

NH₃

The two glass wool producing plants annually report emission figures for NH₃ to the Norwegian Environment Agency. The emission figures are based on measurements.

Particles

The two plants producing glass wool have reported emission figures to the Norwegian Environment Agency since 1990. The glass fibre producing plant has reported emissions from 1996; for the period 1990-1995, the 1996 figure is used in the inventory. One glass-producer with particle emissions reported figures from 1995. Emission figures from 1990 to 1994 were assumed to be the same as the reported 1995 figure. This plant was closed down in 1999.

The size distribution from EEA (2019) is used for particle size distribution, that is PM_{2.5} is 80 per cent of TSP and PM₁₀ is 90 per cent of TSP.

BC

Emissions have been estimated from a share of PM_{2.5} emissions given by EEA (2019), that is 0.062 per cent of PM_{2.5} emissions.

Heavy metals and POPs

Emission of lead has been reported from two glass-producers to the Norwegian Environment Agency. One of them was closed down in 1999. The emission of lead is due to the lead content in the raw material used. Emissions of other heavy metals are reported under Energy.

4.2.3.3 Uncertainties

For the years where reported emission figures for particles do not exist, Statistics Norway has assumed that emissions are in the same order of magnitude as for the first year of reporting. This is uncertain and only an estimate, since it does not consider annual changes in raw materials, production rates, nor possible cleaning devices.

4.2.3.4 Source specific QA/QC

The emissions are reported according to their permits to NEA. The case handler assesses the reported emissions and activity data and contacts the plants if needed. Emissions from these plants also include GHGs (green house gases), so the inventory team at NEA undertakes QA/QC procedures that also are relevant for other components. See section 1.6 for the description of the general QA/QC procedure.

4.2.4 Mining and extraction of stones and minerals

NFR 2A5A

Last update: 10.01.18

4.2.4.1 Description

Mining and extraction of stones and minerals are done by several plants. Particles are emitted during these processes.

4.2.4.2 Method

Particles

Emission figures are reported to the Norwegian Environment Agency. Reported figures exist from 1992. Emission figures for 1990 and 1991 are assumed by Statistics Norway, in accordance with the Norwegian Environment Agency, to be the same as reported figures in 1992. An exception is one plant, which only reported emissions for 1992. For this plant, Statistics Norway has calculated emissions based on production rates for previous and later years.

It is given for most plants that they use fabric filter or textile fibre to clean their particle emissions. It is assumed by the Norwegian Environment Agency that the particles emitted are larger than PM₁₀. Thus, BC has not been considered for this category.

The Norwegian inventory uses the size distribution recommended by TNO (Institute of environmental and energy technology 2002) for sandpits and rock-crushing plants (Table 4.3).

Emission of particles is often a source of heavy metal emissions since particles often contain heavy metals. Type of metals will however depend on the origin of the particles. Metals might therefore be emitted during mining and extraction of stones and minerals. There are, however, no data available for calculating emissions of heavy metals.

4.2.4.3 Uncertainties

For years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order of size as for the first year of reporting. This is uncertain and a result of lack of better data. The size of the particles emitted from mining and extraction will also depend on the type of stone/mineral and production process. The particle size distribution used in the inventory does not consider these differences.

4.2.4.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.5 Construction and demolition

NFR 2A5B

Last update: 13.02.20

4.2.5.1 Description

Construction and building includes a lot of different activities that will generate particle emissions. Building of roads, railways, tunnels and demolition of buildings is also a source of particle emissions, but no emission factors are found in the literature, and therefore such emissions are not included in the inventory.

4.2.5.2 Method

Particles

Emission factors and activity data are used to estimate the Norwegian emissions.

4.2.5.3 Emission factors

The emission factors used are shown in Table 4.2.

Table 4.2. Particle emission factors for building and construction. Tonne/hectare/year

| Component | Tonne/hectare/year |
|-------------------|--------------------|
| TSP | 2.9 |
| PM ₁₀ | 0.86 |
| PM _{2.5} | 0.086 |

Source: EEA (2019)

Statistics Norway assumes that none of the processes used in building and construction will lead to BC emissions. Hence, BC has not been considered for this activity.

4.2.5.4 Activity data

The activity data used is the annual area of completed buildings from the building statistics at Statistics Norway. These are reported in the NFR table and in Appendix F.

4.2.5.5 Uncertainties

The particle emissions depend on climate conditions as well as building traditions and building materials. Since the emission factors used are based on surveys in other countries than Norway, these factors might not be ideal for Norwegian conditions.

4.2.5.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.6 Ceramics

NFR 2A6

Last update: 17.01.20

4.2.6.1 Description

One plant producing bricks reports emission of particulate matter from limestone and dolomite use to the Norwegian Environment Agency. The plant was closed in 2014.

4.2.6.2 Method

Particles

Emissions have been reported to the Norwegian Environment Agency since 2000. Reported figure for 2000 have been used for all years since 1990. The same particle size distribution is used as for production of cement as given in TNO (Institute of environmental and energy technology 2002). PM₁₀ and PM_{2.5} are assumed to be 85 and 30 per cent of TSP, respectively.

BC

Emissions has been estimated from a share of PM_{2.5} emissions. Values for bricks production are given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 37.5 per cent.

4.2.6.3 Uncertainties

Uncertainty estimates are given in Appendix C.

4.2.6.4 Source specific QA/QC

The emissions are reported according to their permits to NEA. The case handler assesses the reported emissions and activity data and contacts the plants if needed. Emissions from these plants also include GHGs (green house gases), so the inventory team at NEA undertakes QA/QC procedures that also are relevant for other components. See section 1.6 for the description of the general QA/QC procedure.

4.2.7 Non-metallurgical Magnesia Production

NFR 2A6

Last update: 27.10.16

4.2.7.1 Description

One plant whose main activity is producing magnesium oxide and calcium oxide from limestone and dolomite is included in the emission inventory. The plant was established in 2005.

4.2.7.2 Method

Particles

Emissions have been reported to the Norwegian Environment Agency for the years 2005-2008 and 2013 and onwards. Linear interpolation has been used for the intervening years. No information is found regarding the particle size distribution for particles emitted during production. In lack of other data we used the same distribution as for aluminium production. PM₁₀ and PM_{2.5} are assumed to be 100 and 43 per cent of TSP, respectively.

BC

Emissions has been estimated from a share of PM_{2.5} emissions. Values for bricks production are given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 37.5 per cent.

SO₂

Emissions have been reported to the Norwegian Environment Agency since 2006.

Dioxins

Emissions have been reported to the Norwegian Environment Agency for the years 2011, 2013 and onwards.

PCB

Emissions have been measured and reported to the Norwegian Environment Agency for 2010. An emission factor has been built from these emissions measurements to estimate emissions for the whole timeseries.

4.2.7.3 Activity data

The amount of limestone and dolomite used by the plant in their calculation is annually reported

to the Norwegian Environment Agency.

4.2.7.4 Uncertainties

Uncertainty estimates are given in Appendix C.

The particle distribution used is not specified for the plants, and the particles emitted might therefore have another distribution than the one suggested.

4.2.7.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.8 Sandpit and rock-crushing plant

NFR 2A6

Last update: 13.02.20

4.2.8.1 Method

Particles will be emitted during crushing of rocks and at sandpits. In the inventory, emissions are estimated based on the production of sand and crushed stone from the production statistics at Statistics Norway, and emission factors recommended by EEA.

Emission of particles is often a source of heavy metal emissions since particles often contain heavy metals. Type of metals will however depend on the origin of the particles. Metals might therefore be emitted during crushing at sandpits and rock-crushing plants. There are however no data available for calculating emission of heavy metals.

4.2.8.2 Activity data

The production of sand and crushed stone is annually given by Statistics Norway's production statistics. These are reported in the NFR table and in Appendix F.

4.2.8.3 Emission factors

The emission factors used are shown in Table 4.3.

Table 4.3. Particle emission factors for sandpits and rock-crushing plants. Ratio X^1 /TSP

| Component | g/tonne produced |
|-------------------|------------------|
| TSP | 102 |
| PM ₁₀ | 50 |
| PM _{2.5} | 5 |

¹ X is either PM_{2.5}, PM₁₀ or TSP.

Source: EEA (2019)

All particles are assumed to be larger than PM_{2.5}. Thus, no emission of BC has been estimated.

4.2.8.4 Uncertainties

This emission source is highly uncertain since the emissions will vary from one place to another depending on the different processes in use, type of raw materials and of course the activity level. Little information is available in the literature. The emission factors used are only based on one source and are uncertain. In addition, there is uncertainty regarding the activity data. The

PRODCOM codes used in the production statistics include total production of sand and crushed stone in Norway, but some of it might not be relevant for these calculations.

4.2.8.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.9 Concrete pumice stone

NFR 2A6

Last update: 27.10.16

4.2.9.1 Description

Two factories have reported emissions of SO₂ and particles from concrete pumice stone production to the Norwegian Environment Agency until 2004 when one of them was closed down. Non-combustion emissions of SO₂ originate from the clay used in the production process.

Particles often contain heavy metals, but type of metals and volumes will depend on the origin of the particles. Metals might therefore be emitted during production of concrete pumice stone. Statistics Norway and the Norwegian Environment Agency have, however, no data available for calculating emissions of heavy metals from this source.

4.2.9.2 Method

SO₂

Emission figures for SO₂ are reported to the Norwegian Environment Agency, based on measurements at the two manufacturing plants in Norway. The plants have installed flue gas desulphurisation equipment.

Particles

The plants have reported emissions of particles to the Norwegian Environment Agency since 1990. It is assumed that the reported figures include both process and combustion emissions, so emission calculations from fuel combustion are not done for these two plants. The plants have installed particle filters.

No information concerning particle size is found in national or international literature, but the Norwegian Environment Agency assumes that most of the particles emitted from these plants are smaller than PM₁₀. Statistics Norway has decided to use the same particle size distribution for production of cement as given in TNO (Institute of environmental and energy technology 2002). PM₁₀ is therefore assumed to be 0.85*TSP and PM_{2.5} is 0.3*TSP.

BC

Emissions has been estimated from a share of PM_{2.5} emissions. Values for bricks production are given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 37.5 per cent.

4.2.9.3 Uncertainties

The particle size distribution used is not specific for production of concrete pumice stone, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate.

4.2.9.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.10 Rock wool production

NFR 2A6

Last update: 12.01.17

4.2.10.1 Description

Three plants in Norway produced rock wool until 2003 when one of them was closed down. In the inventory, emission figures for NH₃, particles and heavy metals are included. Particles originate from the cutting of the mineral wool and from fuel used in the production. The emissions of heavy metals are partly due to use of coal/coke, but mainly due to the stone used in the production. Emissions of dioxins and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) are neither reported nor calculated since emissions of these components are minor or not occurring.

4.2.10.2 Method

NO_x

Emission figures are reported to the Norwegian Environment Agency.

NH₃

Emission figures are reported to the Norwegian Environment Agency. Figures exist from 1992. It is assumed in the inventory that emission figures for 1990 and 1991 are the same as the reported figure in 1992.

Particles

Emission figures are reported to the Norwegian Environment Agency. Most of the emissions come from the spin chamber, and the particle size is assumed to be less than 1 µm. Particles emitted from the fabric filter are also assumed to be smaller than 1 µm. All emissions are therefore set to be smaller than PM_{2.5}. All assumptions are made by the Norwegian Environment Agency in accordance with the industry. It is assumed that the reported figures include both non-combustion and combustion emissions. Combustion emissions of particles are therefore not calculated.

BC

Emissions have been estimated from a share of PM_{2.5} emissions. Values for glass fiber production are given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 0.06 per cent.

Heavy metals and POPs

Emission figures for Pb, Cd, As and Cr have been reported annually from one of the plants to the Norwegian Environment Agency since 1999. The figures are based on measurements. It is assumed that the reported figures include combustion emissions, and emission calculations from fuel combustion are not done for these heavy metals. Statistics Norway has calculated the emission figures for missing years (1990-1998) based on reported figures in 1999 and production rates for previous years. For the two plants not reporting, Statistics Norway calculates emissions based on derived emission factors from the one plant that reports and production volumes at each plant.

4.2.10.3 Activity data

Production volumes of rock wool are annually reported from the plants to the Norwegian Environment Agency.

4.2.10.4 Emission factors**BC**

Emissions have been estimated from a share of PM_{2.5} emissions. Value given by IIASA in Kupiainen and Klimont (2004) for glass production have been used. As a share of PM₁₀, emission factor is 0.25 per cent.

Heavy metals

A default emission factor is derived for each component (Pb, Cd, As and Cr) based on the annually reported emission figures and production rates from the one plant reporting. The derived emission factors are used to calculate emissions from the two other plants (one of these were closed down in 2003).

*Table 4.4. Emission factors for Pb, Cd, As and Cr from production of rock wool.
g/tonne produced rock wool*

| Component | Emission factors (g/tonne produced rock wool) |
|------------------|--|
| Lead (Pb) | 0.164 |
| Cadmium (Cd) | 0.001 |
| Arsenic (As) | 0.031 |
| Chromium (Cr) | 0.703 |

Source: Statistics Norway/Norwegian Environment Agency

4.2.10.5 Uncertainties**Activity data**

The activity data is assumed to be of good quality since this is production rates reported from each plant to the Norwegian Environment Agency.

Emission factors

Several conditions influence the emission of heavy metals, such as production rates and raw materials, and these factors can vary from one plant to another. To derive emission factors based on one plant's reported emission figures and production volume and use these factors to estimate emissions at other plants is therefore quite uncertain.

4.2.10.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.11 Production of mineral white (plaster)

NFR 2A6

Last update: 11.09.15

4.2.11.1 Description

Two plants producing mineral white in Norway are included in the inventory with their emissions of mercury and particles. The mercury content in the raw materials leads to emission of mercury, and during the production process, particles are emitted.

4.2.11.2 Method**Particles**

Emission figures are reported to the Norwegian Environment Agency. Reported emission figures exist since 1992 and figures for 1990 and 1991 are assumed by Statistics Norway, in accordance with the Norwegian Environment Agency, to be the same as the figures reported in 1992. The particles are purified through a fabric filter, and it is assumed by the Norwegian Environment Agency that the particles emitted after the filter are smaller than PM₁₀.

According to TNO (Institute of environmental and energy technology 2002), PM_{2.5} is 30 per cent of TSP, while PM₁₀ is assumed to be the same as TSP. The Norwegian inventory uses this distribution.

BC

Emissions are estimated from a share of PM_{2.5} emissions. As no share for BC was found in the literature, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

Heavy metals

The plants have reported emission figures to the Norwegian Environment Agency since 2000. For one of the plants, historical emissions are based on reported figures for 2000 and production volumes. For the other plant, emission figures for 1990-1999 are assumed to be the same as the reported figure for 2000, due to lack of production data for previous years. Annual emissions are assumed to be low.

4.2.11.3 Activity data

Production volumes for calculation of historical emissions of mercury for one of the plants are reported to the Norwegian Environment Agency.

4.2.11.4 Emission factors

Emission factors for mercury are derived from historical calculations for one plant, based on reported figures for the first year of reporting and production volumes.

4.2.11.5 Uncertainties

Historical emissions of mercury for both plants are uncertain. For one plant, the emission figures are based on a derived emission factor and production volumes, and do not take into account changes in raw materials and possible cleaning devices. For the other plant, it is assumed, due to lack of historical production data, that the historical emissions are the same as the reported figures for 2000. This is just an estimate and does not consider annual changes in raw materials, production rates, or possible cleaning devices.

The particle size distribution used in the inventory is not specific for the plants. The particles emitted might therefore have another distribution than the one suggested by TNO, which is used in the inventory.

4.2.11.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.12 Construction and repairing of vessels - Sandblasting

NFR 2A6

Last update: 14.09.15

4.2.12.1 Description

Five plants constructing and repairing vessels are included in the inventory with their particle emissions. One of the plants was closed down in 2000. Emission of particles is due to the different processes during construction and repairing of vessels, but most of the particles are emitted from sandblasting.

Emission of particles is often a source of heavy metal emissions since particles often contain heavy metals. Type of metals will however depend on the origin of the particles. Metals might therefore be emitted during sandblasting and repairing/construction of vessels. There are however no data available for calculating emissions of heavy metals.

4.2.12.2 Method

Particles

Emission figures are reported to the Norwegian Environment Agency.

For four of the five plants, there are no information regarding cleaning device, but it is assumed by the Norwegian Environment Agency that they have fabric filter and/or wet washer. For the last one, particle emissions are purified in cyclones, and the size of the particles emitted is larger than PM₁₀.

It is difficult to decide particle size of the particles emitted based on the above information. It is however assumed by the Norwegian Environment Agency that most of the particles are larger than PM₁₀ and therefore no PM_{2.5} and PM₁₀ is considered for this category. Thus, no BC emission has been estimated.

4.2.12.3 Uncertainties

The size of the particles emitted is uncertain and will depend on the cleaning device used at each plant. The different activities during construction and repairing can also result in emission of particles of different sizes.

4.2.12.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3 Chemical Industry

NFR 2B

Last update: 25.02.19

In the Norwegian emission inventory, there are 13 different activities included under chemical industry. Nearly all emission figures from this industry included in the inventory are reported from the plants to the Norwegian Environment Agency (Table 4.5).

Table 4.5. Chemical industry. Components emitted and included in the Norwegian inventory

| Chemical industry | CO | NO _x | NMVOC | SO ₂ | NH ₃ | PM | BC | HM | POP |
|-------------------|----|-----------------|-------|-----------------|-----------------|----|----|----|------|
| Production of: | | | | | | | | | |
| Ammonia | E | IE ¹ | NE | NE | NO | NE | NA | NA | NA |
| Nitric acid | NA | R | NA | NA | R | R | E | NA | NA |
| Silicon carbide | E | NE | R | R | NA | R | NE | R | R |
| Calcium carbide | NA | NE | R | NA | NA | R | E | R | NE |
| Titanium dioxide | NE | NA/R | NE | R | NE | R | E | R | NA/R |
| Methanol | E | R | R | NA | NA | E | E | E | E |
| Sulphuric acid | NA | NA | NA | R | NA | NA | NA | NA | NA |
| Plastic | NA | NA | R | NA | R | R | E | NA | R |
| Explosives | NA | R | NA | NA | NA | NA | NA | NA | NA |
| Chloralkali | NA | NA | NA | NA | NA | NA | NA | R | NA |
| Pigments | NA | NA | NA | NA | NA | NA | NA | R | NA |
| Soap | NA | NA | NA | NA | NA | R | E | NA | NA |
| Paint/varnish | NA | NA | NA | NA | NA | R | E | NA | NA |

E = Figures estimated by Statistics Norway. R = Figures reported by the plant to the Norwegian Environment Agency.

NA = Not Applicable. IE = Included Elsewhere. NE = Not Estimated.

¹ Included in reported figures for nitric acid and other fertilizers.

Source: Statistics Norway/Norwegian Environment Agency

4.3.1 Ammonia Production

NFR 2B1

Last update: 13.02.20

4.3.1.1 Description

In Norway, ammonia is produced by catalytic steam reforming of wet fuel gas (containing ethane, propane and some butane). This is one of the steps during fertiliser production. Hydrogen is needed to produce ammonia, and wet fuel gas is the basis for the production of hydrogen. The plant producing ammonia produces also nitric acid and complete fertilizers (NPK

and calcium nitrate). The reported emissions can not be split and are generally aggregated under 2B2. The exception is for CO from 2B1 where we have estimated emissions. The plant has informed that the process does not result in NH₃ emissions as the NH₃ is absorbed in an argon facility.

4.3.1.2 Method

NO_x

During the production of ammonia there are some non-combustion emissions of NO_x. These emission figures are measured in accordance with standard NS 13284-1 and are included in the reported NO_x emission from nitric acid production and production of other fertilisers.

CO

The emissions are estimated through the production of ammonia as activity data and the emission factor of 0.006 kg CO per tonne ammonia produced (from table 3.7 in the 2019 EMEP Guidebook).

4.3.1.3 Activity data

The ammonia production is reported in the NFR table and in Appendix F.

4.3.1.4 Uncertainties

The uncertainties in the figures reported by the plant are believed to be limited. Uncertainty estimates are given in Appendix C.

4.3.1.5 Source specific QA/QC

The plant reports the emissions of NO_x to the Norwegian Environment Agency. They examine the trends as a quality check.

4.3.2 Production of nitric acid

NFR 2B2

Last update: 13.02.20

4.3.2.1 Description

There are two plants in Norway where nitric acid is produced. Nitric acid is used as a raw material in the manufacture of nitrogenous-based fertiliser. The production of nitric acid (HNO₃) generates NO_x as by-products of high temperature catalytic oxidation of ammonia (NH₃). The production of nitrogenous-based fertiliser also leads to emissions of particles.

4.3.2.2 Method

NO_x

The two plants report the emissions of NO_x to the Norwegian Environment Agency. The reported emissions are based on measurements in accordance with standard NS 13284-1.

NH₃

Emission figures for NH₃ are annually reported to the Norwegian Environment Agency. The reported emissions are based on measurements in accordance with standard NS 13284-1.

Particles

Both plants report emission figures to the Norwegian Environment Agency and have done so since 1990 and 1992. The reported emissions are based on measurements in accordance with standard NS 13284-1. One of the plants has also reported emissions from combustion, but since it is only 1 per cent of the non-combustion emissions, these figures are included in the figures for non-combustion emissions.

In lack of plant specific information regarding particle size distribution of the emitted particles, Statistics Norway uses the distribution given by EEA (2019) for production of ammonium phosphate where PM₁₀ is 80 per cent of TSP and PM_{2.5} is 60 per cent of TSP.

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). BC=1.8 per cent of PM_{2.5}.

4.3.2.3 Activity data

The nitric acid production is reported in the NFR table and in Appendix F.

4.3.2.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

There is uncertainty regarding the size of the particles emitted since there is no plant specific information available. The distribution recommended by EEA (2019) is used in lack of other data.

4.3.2.5 Source specific QA/QC

The plants report the production of HNO₃ to the Norwegian Environment Agency. They compare the trends in the production data with the trend in N₂O emission and use this as a quality check.

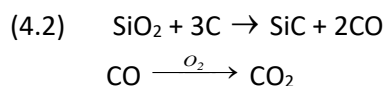
4.3.3 Silicon carbide

NFR 2B5

Last update: 01.03.19

4.3.3.1 Description

Silicon carbide (SiC) is produced by reduction of quartz (SiO₂) with petrol coke as a reducing agent.



In the production of silicon carbide CO are released as by-products from the reaction between quartz and carbon. Sulphur, NMVOCs, particles, heavy metals and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) may also be emitted during the production process. Sulphur originates from the petrol coke.

4.3.3.2 Method

NMVOC

Emission figures are reported to the Norwegian Environment Agency by the plants. The emissions are calculated by multiplying annual production of silicon carbide by an emission factor.

CO

The emissions of CO are calculated by Statistics Norway from the consumption of petrol coke and an emission factor.

SO₂

Emission figures are reported to the Norwegian Environment Agency by the plants. The emissions are calculated from the consumption of petrol coke in dry weight and the sulphur content in the coke. It is assumed that 3 per cent of the sulphur is left in the product or as wastage.

Particles

Emission figures for particles are reported to the Norwegian Environment Agency. Two of the plants have reported since 1990 while the third has reported since 1991. Emission figures for 1990 for this plant are assumed by Statistics Norway and the Norwegian Environment Agency to be the same as the reported figure for 1991. For one of the plants, reported figures have not been used in the inventory for 1990-1993, since the plant means these emission figures are not representative, but a result of different measurement and calculation methods. For this plant, reported emission figures for 1994 have been used for 1990-1993.

There is no detailed information about the particle size distribution for the emissions from silicon carbide production. The Norwegian Environment Agency assumes the emissions have the same particle size distribution as emissions of particles from production of ferroalloys, where all particles are expected to be smaller than PM_{2.5}. This is however an uncertain estimate. This leads to a distribution where TSP=PM₁₀=PM_{2.5}.

Heavy metals

Emission figures have been reported to the Norwegian Environment Agency since 1999/2000. For Pb, Hg and Cd, historical emissions are based on emission factors derived from reported emission figures and production rates for the first year of reporting. Using these emission factors for each plant together with production rates for previous years, historical emissions have been calculated. Cd is reported from one plant for the years after 1992. The calculations for Pb and Cd have been corrected for dust regulations, while emissions of mercury are not affected by these regulations.

Historical emissions of Cu, Cr and As are based on dust emissions for each plant. This has been recommended by the Norwegian Environment Agency, since historical production rate data lack for some years and because changes in emissions will be easier to find when installation of dust control systems reduces the emissions of these metals. Emissions of As are reported to the Norwegian Environment Agency from one plant. Reported figures exist since 1992, and

emissions in 1990 and 1991 are assumed to be the same as reported figures in 1992.

Emission figures for Cu, Cr and Pb are annually reported for all the three plants. In 1999, the plants also reported Hg and Cd due to a heavy metal investigation under the leadership of the Norwegian Environment Agency. After 1999, the plants have not been required to report these metals due to low emissions. Still, one of the plants have reported Cd and Hg figures for all following years, whereas another has reported only Cd; for this plant the 1999 figure for Hg has been used for all later years. For the plant which now has been closed down, the 1999 figures for both Cd and Hg have been used for all later years when the plant still was operating.

POPs

Emission figures for PAH are reported from the plants to the Norwegian Environment Agency. Two of the plants have reported emissions since 1991, while the third one has only reported since 1997. Historical emissions back to 1990 have been calculated based on production rates and an emission factor derived from the first year of reporting and production rate for that year. No PAH profile is available for this source for the years 1990-2015, so the same profile as for aluminium production is used for these years (Table 4.6). After 2015 reported figures for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene from the plants are used in the inventory.

Table 4.6. Distribution of PAH emissions from silicon carbide production. Ratio X^1 /PAH

| Component | Distribution of PAH emissions (ratio) |
|--------------------------|---------------------------------------|
| PAH (Norwegian standard) | 1 |
| PAH-4 (CLRTAP) | 0.15 |

¹ X is either PAH, PAH-6 or PAH-4.

Source: Finstad et al. (2001)

Table 4.7. Distribution of PAH-4 emissions from silicon carbide production. Share of PAH-4. Used 1990-2015

| Component | Distribution of PAH emissions (ratio) |
|------------------------|---------------------------------------|
| benzo(a)pyrene | 0.2 |
| benzo(b)fluoranthene | 0.45 |
| benzo(k)fluoranthene | 0.25 |
| indeno(1,2,3-cd)pyrene | 0.1 |

Source: Norwegian Environment Agency (2016): Expert judgement, Oslo, Norway

4.3.3.3 Activity data

The activity data reported in the NFR table and in Appendix F and used to calculate NMVOC emissions is the annual production of silicon carbide. The activity data used by the plants for the calculation of SO₂ emissions is the consumption of petrol coke in dry weight. The activity data used by Statistics Norway for the calculation of CO emissions is the consumption of petrol coke reported to Statistics Norway. Historical calculations of particle emissions are based on annual production rates and dust emission figures reported to the Norwegian Environment Agency.

4.3.3.4 Emission factors

CO

CO emissions are calculated from the consumption of petrol coke, using a factor of 0.4 tonnes CO/tonnes petrol coke, as recommended by Rosland (1987).

NM VOC

From 2007 and onwards, the emission factors are based on measurements made once a year. The emission factors for the two plants in operation are 10.906 tonne NMVOC/kilotonne Sic and 10.84 tonne NMVOC/kilotonne Sic respectively. For previous years, an average of the emission factors in 2007 and 2008 are applied.

4.3.3.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Heavy metals

The historical calculations for heavy metals are based on derived emission factors for each plant and either production or dust data for previous years, and can only be seen as estimates. The emission figures reported also vary from one year to another, and this is assumed to be, in addition to differences in raw materials, a result of few and uncertain measurements. For the one plant that has not reported emission figures for Hg and Cd since 1999, the same emission figures as those reported in 1999 are used for later years. For the other plant, emissions of Cd have been reported for all years since 1992. Emission figures for Hg have not been reported since 1999. The emission figure for 1999 is used for later years. This is also highly uncertain, but the emission figures are very small and have only marginal impact on the total emissions of these metals.

Particles

The particle size distribution used is not specific for production of silicon carbide, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as the first year of reporting. This is uncertain and a result of lack of better data.

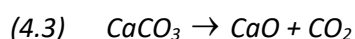
4.3.4 Production of calcium carbide

NFR 2B5

Last update: 16.01.20

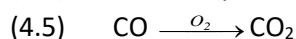
4.3.4.1 Description

One plant in Norway was producing calcium carbide until 2003. The production of calcium carbide generates CO₂ emissions when limestone is heated and when petrol coke is used as a reducing agent. The process can be described through the following equations:



which takes place when limestone (calcium carbonate) is heated.

and



where petrol coke is used as a reducing agent to reduce the CaO to calcium carbide.

Some of the carbon from petrol coke will be sequestered in the product, but not permanently. Thus, this carbon is included in the emission estimate. NMVOC originates from the use of petrol coke in the production process, and NO_x is mainly produced during the high temperature oxidation of nitrogen in the air. Particles are also emitted during the production process. Emission of heavy metals is a result of the heavy metal content in the raw materials.

4.3.4.2 Method

NO_x

Emission figures for NO_x were annually reported to the Norwegian Environment Agency. The reported values are based on calculations.

NMVOC

Reported figures were annually reported to the Norwegian Environment Agency, based on calculations.

Particles

Emission figures for particles were reported from 1992. Figures for 1990 and 1991 are assumed to be the same as for 1992. It does not exist any detailed information about the particle size distribution of the emissions from calcium carbide production. The Norwegian Environment Agency assumes that the emissions are in the same order as emission of particles from production of ferroalloys, where all particles are expected to be smaller than PM_{2.5}. This is however an uncertain estimate. A particle size distribution where PM₁₀ and PM_{2.5} is expected to be the same as TSP, is used in the Norwegian Inventory.

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). BC=1.8 per cent of PM_{2.5}.

Heavy metals and POPs

Emission figures for heavy metals were reported to the Norwegian Environment Agency from 1999. Historical emissions are calculated based on production rates for Pb, Cd and Hg, and based on particle emissions for As, Cu and Cr.

No emission figures for PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) or dioxins are available.

4.3.4.3 Activity data

Particle emissions used in the calculations of As, Cu and Cr have been reported to the Norwegian Environment Agency.

4.3.4.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Heavy metals

Historical emissions are based on a derived emission factor for the first year of reporting (1999) and calculated with production/particle emission figures for previous years. This is uncertain and only an estimate in lack of other data.

Particles

The particle size distribution used is not specific for production of calcium carbide, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order of magnitude as for the first year of reporting. This is uncertain and a result of lack of better data.

4.3.5 Production of titanium dioxide

NFR 2B6

Last update: 16.01.20

4.3.5.1 Description

One plant in Norway produces titanium dioxide. The ore is crushed and pulverized in mills. The crushed raw material is separated in various steps. Ilmenite and the by-product magnetite are cleaned during acid treatment and flotation. The ilmenite concentrate is drained and the water content is reduced to approximately 3.5 per cent. Emissions of SO₂, heavy metals and particles from the plant are included in the inventory. The particle emissions are a result of the crushing of the ore in the mills and from the annealing furnace, while the heavy metal emissions are due to the metal content in the raw material used.

Another plant produces titanium dioxide slag and also pig iron as a by-product. The raw material is the mineral ilmenite, and coal is used as a reducing agent. SO₂ originates from the sulphur in the reducing agent used, while NO_x is produced primarily by the high temperature oxidation of nitrogen in the air. Heavy metal emissions are due to the metallurgical melting process and the content of heavy metals in the raw materials used.

4.3.5.2 Method

SO₂

The emission figures for SO₂ are based on calculations and are reported annually to the Norwegian Environment Agency.

NO_x

The emission figures for NO_x for the plant producing titanium dioxide slag are estimated and reported to the Norwegian Environment Agency.

Particles

Since 1990, emissions of particles have been reported annually to the Norwegian Environment Agency. The particles are assumed to be of a size less than $PM_{2.5}$.

BC

Emissions are estimated from a share of $PM_{2.5}$ emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). $BC=1.8$ per cent of $PM_{2.5}$.

Heavy metals and POPs

Both plants report emission figures to the Norwegian Environment Agency.

One plant reported emission figures for Pb, Cd and Hg for the period 1990 to 1999. After 1999, there has not been any reporting, as a result of very small emission figures. No emissions of persistent organic pollutants are reported or calculated.

The other plant reports emission figures for Pb, Cd, Cr, Cu, As and Hg. Emissions exist from 1990, 1992 or later, depending on type of heavy metal. For dioxins and PAH, reported figures have only been available from 1999. In lack of production rate data for previous years, it has been assumed that yearly emissions are the same as in the first year of reporting. PCB emissions have been measured and reported since 2006. Emissions from 1990 to 2006 are based on reported emissions from 2006. Emission figures for PAH are reported from the plant to the Norwegian Environment Agency. No PAH profile is available for this source for the years 1990-2015. The Norwegian Environment Agency suggests a distribution of the emissions where PAH-4 is 15 per cent of reported PAH emissions. Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are estimated using the same PAH-profile as for aluminium production for the years 1990-2015. After 2015 the plant has reported figures for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene.

4.3.5.3 Uncertainties**Heavy metals and POPs**

Reported emission figures vary from one year to another, partly due to differences in raw materials, but mainly as a result of uncertain measurements. The reported figures are based on a limited number of measurements, and the emissions will vary from minute to minute, since the production of pig-iron is a non-continuous process. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as the first year of reporting. This is uncertain and a result of lack of better data.

Particles

The particle size distribution used is only an assumption, and we can not preclude that the distribution is different from the one used in the inventory.

4.3.5.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.6 Production of methanol

NFR 2B10A

Last update: 25.02.19

4.3.6.1 Description

One plant in Norway produces methanol. Natural gas and oxygen are used in the production of methanol. The conversion from the raw materials to methanol is done in various steps and on different locations at the plant. NMVOC are emitted during the production process. Emissions from flaring of natural gas in connection with production of methanol are reported under 2B10A.

4.3.6.2 Method

The plant reports emission figures for NMVOC and NO_x to the Norwegian Environment Agency. The reported emissions are based on measurements. In addition, emissions from flaring of natural gas are estimated by multiplying the amount of gas flared with the emission factors shown in Table 4.8.

Table 4.8. Emission factors for flare

| Component | Flare natural gas kg/1000 Sm ³ |
|------------------------|--|
| SO ₂ | 0 |
| CO | 1.5 |
| NO _x | ¹ |
| Particles | 0.0018 |
| NMVOC | 0.06 |
| | mg/1000 Sm ³ |
| Pb | 0.25 |
| Cd | 1.7 |
| Hg | 1 |
| Cu | 16 |
| Cr | 21 |
| As | 3.8 |
| Dioxins | 0.00005 |
| Benzo(a)pyrene | 0.02 |
| benzo(b)fluoranthene | 0.04 |
| benzo(k)fluoranthene | 0.02 |
| indeno(1,2,3-cd)pyrene | 0.02 |

¹ Reported to the Norwegian Environment Agency since 2000.

Source: Statistics Norway/Norwegian Environment Agency. PAH: EEA (2016)

BC emissions have been estimated using the same emission factor as for flaring of natural gas 1B2c.

4.3.6.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

4.3.6.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.7 Production of sulphuric acid

NFR 2B10A

Last update: 05.02.15

4.3.7.1 Description

Three plants in Norway produced sulphuric acid until March 2006 when one of them was closed down. The production of sulphuric acid leads to emissions of SO₂. All the three plants report the emissions from the production to the Norwegian Environment Agency, but only one plant have specified that the emissions come from the production of sulphuric acid. For the two other plants, the emissions have been included in the reported emissions from the plants' main production (production of nickel and zinc, respectively).

4.3.7.2 Method

The plant reports annually emission figures for SO₂ to the Norwegian Environment Agency. The reported figures are based on measurements.

4.3.7.3 Uncertainties

No source specific uncertainty is known.

4.3.7.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.8 Production of plastic

NFR 2B10A

Last update: 16.01.20

4.3.8.1 Description

Three plants report emissions to the Norwegian Environment Agency under this source category. One of the plants produces ethylene, one propylene and polyethylene, and the third plant has vinyl chloride production. Two of the reporting plants were merged up to 2001. Various components are emitted during the production of plastic. NMVOC emissions are from leakages in the process.

During the production process of ethylene and vinyl chloride there is an oxide chloride step for production of ethylene chloride, followed by cracking to vinyl chloride monomer and hydrochloric acid. Various chloride components are produced during these processes, including dioxins. However, most of the dioxins end up in the EDC-tar, which is combusted in an own chloride recycling installation. Particles (PVC-dust) are also emitted during the production of vinyl chloride.

Emissions from flaring of fuel gas in connection with production of plastic are now reported under 2B10a.

4.3.8.2 Method

NH₃ and NMVOC

Emission figures are annually reported to the Norwegian Environment Agency. Reported NMVOC emissions are based on measurements. The emissions of NH₃ are regarded as equal to use. As some of the ammonia is stored in the product, the emissions are probably somewhat overestimated.

Particles

Emission figures have been reported to the Norwegian Environment Agency since 1992. Emission figures for 1991 and 1990 are assumed to be the same as reported figures in 1992. The particle emissions have decreased since 1996 as a result of installation of cleaning devices. The emissions are purified in cyclones, but there is no available information regarding particle size. In lack of plant specific information, the distribution TSP=PM₁₀=PM_{2.5}, as in TNO (Institute of environmental and energy technology 2002), is used in the calculation.

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). BC=1.8 per cent of PM_{2.5}.

Dioxins

The plant producing vinyl chloride reports dioxin emission figures. Figures are reported since 1990 except for 1992 and 1994. Emission figures for 1992 and 1994 are based on the reported data for 1991 and 1993.

HCB

The plant producing vinyl chloride reports HCB emission figures since 1996. Emissions from 1990 to 1995 are based on the 1996 reported emissions.

PCB

PCB emissions have been reported since 2010. Emissions from 1990 to 2010 are based on the 2010 reported emissions.

4.3.8.3 Uncertainties

It is difficult to measure leakages of NMVOC and therefore the uncertainty is regarded as being high.

The particle size distribution used is not specific for the plants, and the particles emitted might therefore have another distribution than the one suggested by TNO.

4.3.8.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.9 Production of explosives

NFR 2B10A

Last update: 05.02.15

4.3.9.1 Description

There has been one plant in Norway producing explosives, but the plant was closed down in 2001. Nitric acid was used as a raw material in the manufacture of explosives, and during the production of nitric acid, NO_x was emitted.

Reported particles emission figures to the Norwegian Environment Agency exist only for 1997-1999. Annual emissions were so low that they have not been included in the Norwegian inventory

4.3.9.2 Method

NO_x

Emission figures were annually reported to the Norwegian Environment Agency, and the figures were based on calculations.

4.3.9.3 Uncertainties

No source specific uncertainty is known.

4.3.9.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.10 Chloralkali production

NFR 2B10A

Last update: 05.02.15

4.3.10.1 Description

One plant in Norway produced chloralkali until 2005. Before 1997, mercury was used in the chloralkali production and emitted during the process. In 1997, the plant changed its production process and stopped using mercury, but in the following years there were still some mercury emissions.

4.3.10.2 Method

Hg

Emission figures were reported to the Norwegian Environment Agency.

4.3.10.3 Uncertainties

No source specific uncertainty is known.

4.3.10.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.11 Production of pigments

NFR 2B10A

Last update: 05.02.15

4.3.11.1 Description

Two plants are included in the inventory. One plant produces copper oxide for bottom paint and emits copper to air during the production process. Emissions of Cd and Pb have been reported since 2002. Emissions for 1990-2001 are set to be the same as the reported figure in 2002. Also minor amounts of arsenic and chromium are emitted. The other plant produces zinc chromate, and chromium is emitted.

4.3.11.2 Method

Emission figures are reported to the Norwegian Environment Agency.

4.3.11.3 Uncertainties

Reported emission figures for 1990 and 1991 for the plant producing zinc chromate are not occurring. In the inventory, the same figure as reported for 1992 is used for 1990 and 1991.

4.3.11.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.12 Production of soap

NFR 2B10A

Last update: 10.01.17

4.3.12.1 Method

Two plants producing soap have reported emission figures for particles to the Norwegian Environment Agency. One of the plants has only reported for 1990 and 1991. The plant has after 1991 had a temporary permission without reporting requirements and is therefore not included after 1991 due to lack of data. The other plant reported figures for 1992-1994. Emissions for 1990 and 1991 are assumed to be the same as reported figure in 1992, while emissions for 1995-1997 are assumed to be the same as reported figure in 1994. Annual emission figures are low.

The particles have been purified through filters and scrubbers and the Norwegian Environment Agency assumes the sizes of the particles are smaller than PM_{2.5}. BC emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2016). BC=1,8 per cent of PM_{2.5}.

4.3.12.2 Uncertainties

For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as reported in one of the other years. This is uncertain and a result of lack of better data.

4.3.12.3 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.13 Paint and varnish production

NFR 2B10A

Last update: 10.01.17

4.3.13.1 Method

One plant producing paint has reported emission figures for particles to the Norwegian Environment Agency since 1995, after first getting an emission permit in 1994. Annual emissions are small. It is assumed by the Norwegian Environment Agency that the particles emitted are smaller than PM_{2.5}. BC emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2016). BC=1,8 per cent of PM_{2.5}.

4.3.13.2 Uncertainties

No source specific uncertainty is known.

4.3.13.3 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure QA/QC procedure

4.4 Metal production

NFR 2C

Last update: 10.01.17

Metal production in Norway includes plants producing iron and steel, ferroalloys, aluminium, nickel and zinc and also magnesium until spring 2006. Production of anodes is also included in this chapter. As shown in Table 4.9, most of the figures in the national inventory are from the plants' annual reports to the Norwegian Environment Agency.

Table 4.9. Metal production. Components emitted and included in the Norwegian inventory

| | SO ₂ | NO _x | NH ₃ | NMVOC | CO | PM | BC | HM | POP |
|-------------------------|-----------------|-----------------|-----------------|-------|----|----|----|----|-----|
| Production of: | | | | | | | | | |
| 2C1 Iron and steel | NE | NE | NE | NE | NE | R | E | R | R |
| 2C2 Ferroalloys | R | R | NE | E | NE | R | E | R | R |
| 2C3 Primary aluminium | R | E | NE | NE | E | R | E | R | R/E |
| 2C3 Secondary aluminium | NA | NA | R | NA | NA | R | E | R | R |
| 2C4 Magnesium | R | NA | NA | NA | R | R | NA | R | R |
| 2C6 Zinc | R | NE | NE | NE | NE | R | NA | R | NE |
| 2C7B Nickel | R | R | R | NE | NE | R | NA | R | NE |
| 2C7C Anodes | R | R | NE | NA | NA | R | E | R | R |

E = Figures estimated by Statistics Norway (Activity data * emission factor). R = Figures reported by the plant to the Norwegian Environment Agency. NA = Not Applicable. NE = Not Estimated

Source: Statistics Norway/Norwegian Environment Agency

4.4.1 Production of iron and steel

NFR 2C1

Last update: 13.02.20

4.4.1.1 Description

Several plants are included in the time series for the production of iron and steel, but not all plants are currently in production. Currently, the emissions are from one steel producing plant that uses an electric arc furnace (EAF). The components included in the inventory are particles, black carbon, heavy metals and POPs. One plant producing titanium dioxide slag also produces pig iron as a by-product, but the emissions from this plant are registered under 2B6.

4.4.1.2 Method

Particles

One plant has reported figures since 1990 while the other only has reported since 1998. For this plant, historical emissions in the period 1990-1997 have been assumed to be the same as the reported figure in 1998, since production rate data for previous years are not available.

The Norwegian Environment Agency assumes that the particles emitted in the production of iron and steel are smaller than $PM_{2.5}$. We can, however, not disregard that some of the particles emitted are larger than $PM_{2.5}$.

BC

Emissions have been estimated as a share of $PM_{2.5}$ emissions. Measurements from one plant showed that 0.1 per cent of the dust was carbon and this is used in the inventory. As no information on the share of BC and OC was found in the literature for iron and steel production, BC share has been set to be 50 per cent of $PM_{2.5}$ using the default method described in Aasestad (2013). Indeed, the amount of $PM_{2.5}$ is assumed to be equally shared between BC and organic mass (OM). Hence, BC emissions represents 0.05 per cent of $PM_{2.5}$ emissions.

Heavy metals and POPs

Heavy metal emissions are due to the metallurgical melting process and the content of heavy metals in the raw materials used. One plant reports emission figures to the Norwegian Environment Agency. Reported figures for heavy metals (Pb, Cd, Cr, Cu, As and Hg) exist from 1990, 1992 or later, depending on type of heavy metal. For dioxins and PAH, reported figures have been available since 1997. The reported numbers from 1997 have been used for all years from 1990 to 1996. After 2015, the plant has reported figures for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene.

Diffuse emissions have been included from one plant. In lack of production rate data for previous years, it has been assumed that yearly emissions are the same as in the first year of reporting.

Plants reports HCB emissions to the Norwegian Environment Agency since 2011. Emissions for the previous years have been estimated using the emission factor from the EMEP inventory guidebook (EEA 2013) and the crude steel production.

Plants reported PCB emissions in 2010. Emissions for the other years have been estimated using the data reported in 2010.

4.4.1.3 Activity data

The production of steel is reported in the NFR table and in Appendix F.

4.4.1.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Heavy metals and POPs

Reported emission figures vary from one year to another, partly due to differences in raw materials, but mainly as a result of uncertain measurements. The reported figures are based on a limited number of measurements, and the emissions will vary from minute to minute, since the production of iron and steel is a non-continuous process. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as the first year of reporting. This is uncertain and a result of lack of better data.

Particles

The particle size distribution used is only an assumption, and we can not preclude that the distribution is different from the one used in the inventory. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order of magnitude as for the first year of reporting. This is an uncertain estimate due to lack of better data.

4.4.1.5 Source specific QA/QC

Annually reported emission figures are first controlled by the Norwegian Environment Agency and then by Statistics Norway.

Adjustments and recalculations have been done for years where reported emission figures seem to be unreasonably high or low compared with previous years. This is applicable when the variations in the reported emission figures do not have a natural explanation.

4.4.2 Production of ferroalloys

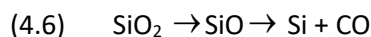
NFR 2C2

Last update: 13.02.20

4.4.2.1 Description

There are 12 plants producing ferroalloys in Norway. One plant closed down in 2001, two plants were closed down during 2003 and two in 2006. One plant was out of production in 2006, but started up again in 2007.

Ferrosilicon, silicon metal, ferromanganese and silicon manganese are now produced in Norway. Ferrochromium was produced until summer in 2001. Ferrosilicon with 65 to 96 per cent Si and silicon metal with 98-99 per cent Si is produced. The raw material for silicon is quartz (SiO_2). SiO_2 is reduced to Si and CO using reducing agents like coal, coke and charcoal.



The waste gas CO and some SiO burns to form CO₂ and SiO₂ (silica dust).

Some of the CO generated from coal is sold for energy use to other industries. The amount of CO gas sold is hence subtracted from the emissions reported under this category and included in energy use in manufacturing industries and construction (NFR 1A2).

In ferroalloy production, raw ore, carbon materials and slag forming materials are mixed and heated to high temperatures for reduction and smelting. The carbon materials used are coal, coke and some biocarbon (charcoal and wood). Electric submerged arc furnaces with graphite electrodes or consumable Soederberg electrodes are used. The heat is produced by the electric arcs and by the resistance in the charge materials. The furnaces used in Norway are open, semi-covered or covered.

The CO stems from the production process. In open or semi-closed furnaces the CO reacts with air and forms CO₂ before it is emitted. This is due to high temperature and access to air in the process. In a closed furnace the CO does not develop to CO₂ as there is no access to air (oxygen) in the process. The waste gas is then led from the furnace and used as an energy source or flared, and is reported under the relevant energy sectors. The technical specification of the furnaces is irrelevant since emissions are calculated using a mass balance or calculated by multiplying the amount of reducing agents in dry weight with country specific emission factors

Several components are emitted from production of ferroalloys. SO₂ originates from the sulphur in the reducing agent used, while NO_x is produced primarily by the high temperature oxidation of nitrogen in the air. NMVOC emissions originate from the use of coal and coke in the production processes by producing ferrosilicon and silicon metal. Heavy metals are emitted from the raw materials (ore) during the metallurgical process, and the particles emitted are mainly silica dust generated during the production process.

4.4.2.2 Method

SO₂

Each plant annually reports emission figures to the Norwegian Environment Agency. Some of the sulphur is trapped in the product. For production of ferromanganese and silicon manganese, 98-99 per cent of the sulphur is trapped, while for other ferroalloys it is assumed that about 5 per cent is trapped. The emissions are calculated from the consumption of reducing agents and electrodes and the content of sulphur in the materials.

NO_x

Emissions of NO_x originate from production of ferrosilicon and silicon metal. Ferromanganese, ferrochrome and silicomanganese do not have significant emissions of NO_x. Emission figures are annually reported by each plant to the Norwegian Environment Agency. The reported emissions are calculated either from the production of metal and metal specific emission factors, see Table 4.12, or on the basis of continuous measurements.

NM VOC

The emissions are estimated by Statistics Norway from the consumption of reducing agents and an emission factor.

Particles

All plants producing ferroalloys report emission figures to the Norwegian Environment Agency. Some have reported since 1990, others since 1992. For plants reported since 1992, emission figures from 1990 and 1991 have been assumed to be the same as reported figures in 1992. According to the ferroalloy industry, particles emitted are smaller than $PM_{2.5}$ (Eikeland, *pers.comm.*⁸). This is, however, an assumption, and we can not preclude that some of the particles might be larger than $PM_{2.5}$. In the inventory, we have decided to use this distribution for all particles emitted from the production of ferroalloys. This means that $TSP=PM_{10}=PM_{2.5}$.

BC

Emissions have been estimated as a share of $PM_{2.5}$ emissions. Measurements of particles composition from several plants were used to estimate the dust carbon content. This value was used to estimate BC. As no information on the share of BC and OC was found in the literature for ferroalloys production, BC share has been set to be 50 per cent of the carbon contents of $PM_{2.5}$ using the default method described in Aasestad (2013). Indeed, the amount of $PM_{2.5}$ is assumed to be equally shared between BC and organic mass (OM). BC emissions represents 3.5 per cent of the $PM_{2.5}$ emissions from ferro-manganese production and 0.23 per cent of the $PM_{2.5}$ from ferro-silicon production (Aasestad 2013).

Heavy metals

The Norwegian Environment Agency imposed in 1999 larger metallurgical plants to map their emissions of heavy metals. Most plants have therefore reported heavy metal emissions to the Norwegian Environment Agency since 1999, but some reported for the first time in 2000 and 2001. An emission factor has been derived for each plant, based on the emissions and production rate for the first year of reporting. These emission factors have been used together with production rates for each year to calculate the emissions back to 1990 for each plant.

Dioxins

All plants producing ferrosilicon report emission figures for dioxins to the Norwegian Environment Agency. It varies, however, when the plants started reporting, so calculations of historical figures back to 1990 have been necessary. An emission factor was derived for each plant based on reported emission data and production rates, and this factor was used to calculate historical emissions based on production rates for each year.

None of the four plants producing ferromanganese and ferrochromium⁹ report emission figures

⁸ Eikeland (2002): Personal information, e-mail dated 29/05 2002. Elkem@elkem.no

⁹ The ferrochromium plant was closed down in 2003.

for dioxins to the Norwegian Environment Agency. The reason is probably that the emissions are so small that they are not measured and therefore not reported (the Norwegian Pollution Control Authority, *pers. comm.*¹⁰). Instead, the emissions are calculated by Statistics Norway based on the general emission factor for combustion of coke and coal in the industry (Table 4.14).

PCB

As for dioxins emissions, PCB emissions are only considered in the ferrosilicon production. Plants reported emissions in 2010 and reported data has been used to estimate emissions for the whole period.

PAHs (Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)

Emissions of PAH from the production of ferroalloys are reported to the Norwegian Environment Agency for plants producing ferrosilicon and silicon metal. All these plants have reported emission figures since 2000. Historical emissions back to 1990 have been calculated based on production rates for each year and an emission factor derived for each plant based on reported figures for 2000, 2001 and 2002. Reported figures and historical calculations are only done for plants producing ferrosilicon and silicon metal. This is based on the assumption that these alloys are produced in open ovens and therefore cause larger emissions of PAH compared to other alloys that are produced in closed ovens, and are assumed to cause no or minor emissions of PAH. No PAH profile is available for this source. The Norwegian Environment Agency suggests a distribution of the emissions where PAH-4 is 15 per cent of reported PAH emissions (Table 4.10).

The PAH emission figures are reported according to Norwegian standard (NS9815), but no PAH profile for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene is available. In lack of other data, the same profile as for aluminium production is used for the years 1990-2015. After 2015 the inventory is based on reported figures for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd) pyrene from eight plants.

Table 4.10. Distribution of PAH emissions from production of ferroalloys

| Component | Distribution of PAH emissions (ratio) |
|--------------------------|---------------------------------------|
| PAH (Norwegian standard) | 1 |
| PAH-4 (CLRTAP) | 0.15 |

Source: Finstad et al. (2001)

¹⁰ Norwegian Pollution Control Authority (2001): Units for dioxins (dioxins.doc). Personal information C. Benestad, 13/03 2001, Oslo: Norwegian Pollution Control Authority.

Table 4.11 Distribution of PAH-4 emissions from production of ferroalloys. Share of PAH-4. 1990-2015

| Component | Distribution of PAH emissions (ratio) |
|------------------------|---------------------------------------|
| benzo(a)pyrene | 0.2 |
| benzo(b)fluoranthene | 0.45 |
| benzo(k)fluoranthene | 0.25 |
| indeno(1,2,3-cd)pyrene | 0.1 |

Source: Norwegian Environment Agency (2016): Expert judgement, Oslo, Norway

4.4.2.3 Activity data

The use of reducing agents are reported in the NFR table and in Appendix F.

4.4.2.4 Emission factors

NO_x

The emission factors used by the ferrosilicon plants in the calculations are based on measurements carried out at three plants.

The emission factors in Table 4.12 are based on several measuring campaigns at four different ferroalloy plants that were carried out from 1995 to 2007. Each measurement period lasted 4 to 8 hours with different operation conditions. Based on this, emission factors for different ferroalloys and operational conditions have been established. The measurements have been carried out by Det norske Veritas, Norsk Energi, SINTEF and TÜV.

The silicon plants have applied a new method. They have used online measurement instruments to measure the emissions of NO_x. The measurements were undertaken in 2010. The instrument applied is NEO laser gas and Testo 350 as a control of the results from the NEO laser gas device. So far there are only two plants where the online measurement devices are installed on a permanent basis. For the other plants the online measurement instruments are used periodically to derive emission factors. One major ferroalloy producing company with four plants use emissions factors (kg NO_x/tonne metal produced) that are: 27 (based on measuring campaigns), 34 and 39 (based on online measurements) and 45 (based on a combination of online measurements and campaigns).

The uncertainties associated with the measurements mainly come from measurement of off-gas flow and measurement of concentration of the NO_x in the off-gas. In addition, the periodical measurement campaigns will not include all variations in the emissions gained over time.

Table 4.12. NO_x emission factors for production of ferrosilicon. Kg NO_x/tonne metal produced

| | Normal operations | Sprinkle - charging | Sprinkle-charging > 750 °C | Source |
|--------------------------|-------------------|---------------------|----------------------------|---|
| Ferrosilicon 75 per cent | 15.3 | 7.0 | 8.3 | Measured in 1995 at Rana Metal and the Thamshavn plant 2005 |
| Ferrosilicon 65 per cent | 6.0 | 4.0 | 5.0 | Estimations ¹ |

¹ Estimations means that this emission factor is not measured but estimated by the plants based on general process experiences.

NM VOC

There is no emission factor for NM VOC in the EMEP 2016 Guidebook, but Statistics Norway uses an emission factor of 1.7 kg NM VOC/tonne coal or coke (EPA 1986) in the calculations.

Dioxins

The emission factors used by the plants in the calculations are given in Table 4.13.

Table 4.13. Emission factors for production of ferroalloys. μg dioxin/tonne metal produced

| | Normal operations | Sprinkle - charging | Sprinkle- charging > 750 °C | Source |
|--------------------------|-------------------|---------------------|-----------------------------|--------------------------------------|
| Silicon metal | 3 | 1.2 | 0.2 | Measured in 1995 at the Fiskaa plant |
| Ferrosilicon 90 per cent | 4 | 1.2 | 0.2 | Estimations ¹ |
| Ferrosilicon 75 per cent | 5 | 1.2 | 0.2 | Measured in 1995 at Rana Metall |
| Ferrosilicon 65 per cent | 5 | 1.2 | 0.2 | Estimations |
| Si96 | 3 | 1.2 | 0.2 | Estimations |

¹ Estimations means that this emission factor is not measured but estimated by the plants based on general process experiences.

For plants that have not reported dioxins emissions to the Norwegian Environment Agency, emissions are estimated through use of an emission factor for combustion of coke and coal in the industry (Table 4.14).

Table 4.14. Emission factor used by Statistics Norway to calculate dioxin emissions from production of ferro manganese/chromium

| Emission factor | |
|-----------------|--------------------------|
| Coal and coke | 1.6 μg /tonne |

Source: Bremmer et al. (1994) and Finstad et al. (2002b)

PAH

The emission factors used by the plants in the calculations are given in Table 4.15.

Table 4.15. Emission factors for production of ferroalloys. g PAH /tonne metal produced

| | Normal operations | Sprinkle - charging | Sprinkle- charging > 750 °C | Source |
|--------------------------|-------------------|---------------------|-----------------------------|--|
| Silicon metal | 3 | 2.6 | 1.6 | Measured in 1995 at the Fiskaa plant |
| Ferrosilicon 90 per cent | 2 | 2 | 1 | Estimations ¹ |
| Ferrosilicon 75 per cent | 1.5 | 1.3 | 0.8 | Measured in 1995 at Rana Metal and the Thamshavn plant |
| Ferrosilicon 65 per cent | 1 | 1.3 | 0.8 | Estimations |
| Si96 | 3 | 2.6 | 1.6 | Estimations |

¹ Estimations means that this emission factor is not measured but estimated by the plants based on general process experiences.

4.4.2.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

The inventory uses a particle size distribution which is an assumption from the ferroalloy industry and not based on measurements. We can therefore not preclude that some of the

particles might be larger than PM_{2.5}.

Heavy metals and POPs

Historical emissions are based on derived emission factors for the first year of reporting, and calculated using production figures for previous years. This is uncertain since the calculation method does not consider quality changes of the raw materials or changes in the production profile at each plant that can have big impact on yearly emissions.

4.4.2.6 Source specific QA/QC

NO_x, NMVOC and CO

The reported emission figures for NO_x, NMVOC and CO are compared with calculations at Statistics Norway.

Emission figures for NMVOC are controlled by multiplying the amount of reducing agents with an emission factor recommended by EPA (1986).

PAH

PAH was first included in the Norwegian Inventory in 2000, and only two plants producing ferrosilicon and silicon metal reported emission figures to the Norwegian Environment Agency for the year 1999. In 2004, a specific emission factor for each plant was derived based on the plants' reported emission figures for 2000, 2001 and 2002 and production volumes,. These factors were then used to recalculate the plants' historical emissions of PAH. A specific emission factor for each plant was considered better to use for historical emissions, instead of using a default emission factor for all plants. The specific emission factors derived for each plant with the new method were lower than those suggested by Benestad (*pers. Comm.*), and this caused approximately 2-12 per cent lower yearly PAH emissions from 1990 to 1999 from this source.

4.4.3 Production of primary aluminium

NFR 2C3

Last update: 13.03.20

4.4.3.1 Description

There are seven plants in Norway producing aluminium. Both prebaked anode and the Soederberg production methods are used.

In the Soederberg technology, the anodes are baked in the electrolysis oven, while in the prebaked technology the anodes are baked in a separate plant. In general, the emissions are larger from the Soederberg technology than from the prebaked technology. There has been a shift from Soederberg to prebaked technology. In 1990, 57 per cent of the aluminium production in Norway was produced with prebaked technology and the share of aluminium production from prebaked increased to 93 per cent in 2017.

Production of aluminium leads to emissions of various components, such as SO₂, NO_x, heavy metals and persistent organic pollutants. The emission of SO₂ are from the sulphur in the reducing agents used. NO_x is primarily produced by the high temperature oxidation of nitrogen

in the air. All plants also report emissions of particles, heavy metals and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene). Emissions of heavy metals are due to the metal content in the raw materials used and the reducing agents.

4.4.3.2 Method

SO₂

The plants report emission figures of SO₂ to the Norwegian Environment Agency. The figures are estimated by each plant based on the amounts of reducing agents used and their sulphur content. All plants have installed flue gas treatment, like, for example, sea water scrubber.

NO_x

NO_x emissions are estimated by Statistics Norway from the level of production and an emission factor derived from measurements at two Norwegian plants. The figure is rather uncertain.

CO

CO emissions are estimated by Statistics Norway from the level of production and the emission factor 120 kg CO/tonn aluminium from EEA 2016.

Particles

Emission figures have been reported to the Norwegian Environment Agency since 1990. The Norwegian Environment Agency assumes that the particles emitted are smaller than PM₁₀. According to EEA (2019), PM₁₀ is 83.3 per cent of TSP, and PM_{2.5} is 67 per cent of TSP for pre-baked cell and 61 per cent of TSP for Soederberg andoes.

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for aluminium production is used, EEA (2016). BC=2.3 per cent of PM_{2.5}.

Heavy metals

The plants report emission figures to the Norwegian Environment Agency. The first requirement for reporting came in 1999, so emission figures before that are insufficient. The concentrations of heavy metals in the air emissions are very low and therefore impossible to measure. Emissions are therefore calculated at each plant, based on the mass flow.

Dioxins

Since the process uses coal and coke as reducing agents, it is assumed that production of primary aluminium gives dioxin emissions. Reported figures for dioxins are not available. The emissions are believed to be so small that reporting is not necessary. Emissions are therefore calculated based on the combustion factor for coal in the industry.

HCB

Emissions are so small that primay aluminium plants do not report them. Hence, emissions have been considered negligible and have not been estimated.

PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno (1,2,3-cd)pyrene)

The reported emission data are assumed to be according to Norwegian standard (NS9815). It is further assumed by the Norwegian Environment Agency that the emissions of PAH-4 accounts for 5 -10 per cent of total PAH emissions reported from production of aluminium (Table 4.16). Historical emission figures have been calculated based on changes in production of aluminium after the Soederberg method.

The PAH profile has been measured at three plants, in addition benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene have been measured at some plants for the last year. Based on these profiles it is believed that PAH-4 accounts for 10 per cent of total PAH emissions from production of aluminium from one plant, 7.5 per cent is used for the other. Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene have been measured in 2016 and in 2001. Based on these measurements a PAH-profile has been made by Hetland, the Norwegian Environment Agency (*pers. comm*)¹¹. The PAH-4 profile used for aluminium production for the years 1990 – 2015 is shown in Table 4.17. After 2015 the inventory is based on reported figures for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene from three plants.

Table 4.16. Distribution of PAH emissions from production of primary aluminium. Ratio

| Component | Distribution of PAH emissions (ratio) |
|--------------------------|---------------------------------------|
| PAH (Norwegian standard) | 1 |
| PAH-4 (CLRTAP) | 0.05-0.1 |

Source: Statistics Norway/Norwegian Environment Agency

Table 4.17. Distribution of PAH-4 emissions from production of primary aluminium. Share of PAH-4

| Component | Distribution of PAH emissions (ratio) |
|------------------------|---------------------------------------|
| benzo(a)pyrene | 0.2 |
| benzo(b)fluoranthene | 0.45 |
| benzo(k)fluoranthene | 0.25 |
| indeno(1,2,3-cd)pyrene | 0.1 |

Source: Hetland (2016)

4.4.3.3 Activity data

The production of aluminium are reported in the NFR table and in Appendix F.

4.4.3.4 Emission factors**NO_x**

Statistics Norway uses the emission factor 0.00071 tonnes NO_x/ tonne produced aluminium in the calculations. This emission factor is assumed by the Norwegian Environment Agency and is based on measurements.

Dioxins

Emissions of dioxins are calculated based on the consumption of coal and an emission factor from Bremmer et al. (1994).

¹¹ Pers. comm, email from Øyvind Hetland, 22. Nov. 2015, Norwegian Environment Agency.

Table 4.18. Emission factor used to calculate dioxin emissions from aluminium production.

| | Emission factor | Source |
|---------------|------------------------|-----------------------|
| Coal and coke | 1.6 µg/tonne | Bremmer et al. (1994) |

4.4.3.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

The particle size distribution are not reported by the plants. Actual emissions are probably somewhat different from those estimated with the size distribution from EEA (2019).

4.4.3.6 Source specific QA/QC

PAH

In 2014, the Norwegian Environment Agency had audits at all aluminium plants. For the four plants that have emisisions of PAH, their systems for monitoring emissions of PAH were checked.

Heavy metals

First requirement for reporting of heavy metals was given in 1999, and the reported figures were based on concentration measurements. The concentration of heavy metals in the air emissions are very low and therefore subject to high degree of uncertainty. The reported emission figures showed large differences from plant to plant, also in the cases where the raw materials came from the same supplier. The Norwegian Environment Agency has had a long discussion with the aluminium industry to find a better method to estimate heavy metals from aluminium production. In 2001 it was decided that reported figures should be based on calculations. Historical reported data were recalculated based on the new calculation method. Recalculation of historical data are normally based on production rate data, but due to very low emissions and relative stable production rates, historical data are set to be the same as the first year of reporting.

4.4.4 Production of secondary aluminium

NFR 2C3

Last update: 17.01.20

4.4.4.1 Description

One open mill in Norway is handling secondary aluminium production. Heavy metals and persistent organic pollutants (dioxins and PAHs , (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)) are emitted in the production of secondary aluminium due to the remelting process. Particles are also emitted during the production process. For earlier years there have also been some emissions of NH₃ and SF₆ from another plant which closed down in 2001.

4.4.4.2 Method

NH₃

For the years 1993-2001, emissions of NH₃ were reported from one plant. This plant closed down in 2001.

Particles

The plant has reported emission figures to the Norwegian Environment Agency from 1993. Emission figures for 1990 to 1992 are in the inventory assumed to be the same as the reported figure in 1993. The following particle size distribution is assumed and used in the Norwegian inventory; PM₁₀ is 70 per cent of TSP and PM_{2.5} is 27.5 per cent of TSP (Institute of environmental and energy technology 2002).

BC

BC has been estimated as fraction of PM_{2.5} emissions. The share of 2.3 per cent of PM_{2.5} from EEA (2019) has been used.

Heavy metals and POPs

The figures are reported annually to the Norwegian Environment Agency. Emission figures exist since 1993, and emissions before 1993 have been supposed to be the same as reported figures in 1993.

The emission figures for heavy metals are based on metal analyses of dust samples. Figures of Pb, Cd and Cr have been reported since 1997. Annual figures can vary a lot from one year to another, and therefore we have used mean values for years when the changes can not be explained by the industry. We have assumed that the emission figures for 1990-1996 are the same as reported figures in 1997, since there are no reported figures of heavy metals before 1997.

PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)

No PAH profile is available for this source. The Norwegian Environment Agency suggests a distribution of the emissions where PAH-4 is 15 per cent of reported PAH emissions. Since no PAH profile is available, emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are assumed to be 25 per cent each of PAH-4. This assumption is used for the years 1990 - 2015. After 2015 one plant has reported figures for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene.

HCB emissions have been reported since 2010. For the period 1990-2009, aluminium production has been used with a emission factor from Japan (Toda 2006) to estimate HCB emissions. Emission factor is 1.7 mg/tonn secondary aluminium.

PCB emissions have been reported from 2006 to 2008 and in 2010. Data reported have been used to build an emission factor and estimate emissions from 1990.

4.4.4.3 Uncertainties

Heavy metals and POPs

The reported figures for heavy metals are estimated based on heavy metal content in the dust samples. The metal content were only analysed for a few dust samples yearly and the reported figures are therefore only a presumption of yearly emission figures. Calculation of emission figures before 1997 are assumed to be the same as reported figures in 1997, and this gives highly uncertain figures since raw materials and production variations may have changed during the period.

The reported emission figures for dioxins and particles vary from one year to another, and it is assumed that this is due to uncertain measurements and process readjustments.

4.4.4.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.4.5 Production of magnesium

NFR 2C4

Last update: 25.02.19

4.4.5.1 Description

There has been one magnesium producing plant in Norway. From the mid-1970s, both the magnesium chloride brine process and the chlorination process were used for magnesium production. Since 1991, only the chlorination process was in use. The plant closed down the production of primary magnesium in 2002, but the production of cast magnesium continued. During 2006, the production of remelting Mg also stopped.

Production of magnesium leads to non-combustion CO emissions. During the calcination of dolomite ($\text{MgCa}(\text{CO}_3)_2$) to magnesium oxide, CO_2 is emitted. During the next step, magnesium oxide is chlorinated to magnesium chloride, and coke is added to bind the oxygen as CO and CO_2 . SO_2 is emitted due to the sulphur in the reducing agent used.

4.4.5.2 Method

CO

Emission figures of CO were reported annually to the Norwegian Environment Agency. These emissions disappeared when the plant closed down the production of primary magnesium in 2002.

SO_2

The SO_2 emissions were estimated from the amounts of reducing agent used (coke) and their sulphur content and reported from the plants to the Norwegian Environment Agency.

Particles

The plant reported emission figures for particles for the first time for the year 1992. Emissions of particles for 1990 and 1991 are assumed to be larger than the reported figure in 1992, since a

cleaning device was installed in 1992. Statistics Norway has no information that can be used to estimate emissions in 1990 and 1991, so the inventory uses the reported emission figure for 1992 also for 1990 and 1991. The Norwegian Environment Agency assumes that reported figures also include emissions from combustion.

No information is found regarding the particle size distribution for particles emitted during magnesium production. In lack of other data, we use the same distribution as for aluminium production, PM₁₀ is 97 per cent of TSP, and PM_{2.5} is 43 per cent of TSP (Institute of environmental and energy technology 2002).

Heavy metals and POPs

Emission of heavy metals is due to the metal content in the reducing agent used. Emission data of Hg, As, Cr and dioxins were reported to the Norwegian Environment Agency. When the plant closed down the production of primary magnesium in 2002, the emissions of As disappeared. Reported figures of heavy metals have only been available since 2000. Emission figures are calculated back to 1990 based on the production rate for each year.

During the chlorination process and the use of coke as a reducing agent, dioxins and HCB are emitted. Emission figures for dioxins were reported to the Norwegian Environment Agency from 1990 while emissions from HCB have been reported from 1992. For 1990 and 1991, 1992 figures have been considered. As no reports were available in 2004 and 2006, emissions have been estimated using 2003 and 2005 figures.

Since the plant is closed, information concerning emission factors has not been prioritised.

4.4.5.3 Activity data

The production of magnesium is reported in the NFR table and in Appendix F.

4.4.5.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

For years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as that of the first year of reporting. This is uncertain and a result of lack of better data. The particle size distribution used is not specific for production of magnesium, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate.

Heavy metals

Historical emissions are based on a derived emission factor for the first year of reporting and calculated with production figures for earlier years. This is uncertain and only an estimate since it does not consider annually changes in raw materials nor possible cleaning devices.

4.4.5.5 Source specific QA/QC

The latest reported emission data from the plant were compared with previous reported data

and the emissions were compared with the production.

4.4.6 Production of zinc

NFR 2C6

Last update: 17.01.20

4.4.6.1 Description

One plant in Norway produces zinc. SO₂, particles and heavy metals are emitted during the process. Emission of SO₂ originates from the sulphur in the reducing agent used.

4.4.6.2 Method

SO₂

The plant reports emission figures to the Norwegian Environment Agency. The SO₂ emissions are estimated from infrequent measurements combined with calculations.

Particles

Emission figures for particles have been reported since 1991. Emissions for 1990 are assumed to be the same as the reported figure for 1991. It is assumed that of the particles emitted, 90 per cent is PM₁₀ and 80 per cent is PM_{2.5} (Institute of environmental and energy technology 2002) and this particle size distribution is used in the Norwegian inventory.

Heavy metals and POPs

The plant reports emission figures for Cd, Pb, Hg, Cu, Cr and As. Reported figures exist since 1992, and emissions in 1990 and 1991 are assumed to be the same as reported figures in 1992.

PCB emissions have been estimated using the emission factor given by EEA (2013). This emissions factor, which amounts to 2.0 µg/tonn of zinc, is used for the whole period.

4.4.6.3 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.4.7 Production of nickel

NFR 2C7B

Last update: 25.02.19

4.4.7.1 Description

One plant in Norway produces nickel. During the production of nickel SO₂, NO_x, NH₃, particles and heavy metals are emitted. CO₂ is emitted in the production of nickel, due to the soda from the production of nickel carbonate and use of coke as a reducing agent, while SO₂ is a result of the sulphur content in the coke used. NO_x is produced primarily by the high temperature oxidation of nitrogen in the air. Emission of heavy metals is due to the metal content in reducing agent used. Particles are also emitted during the production process.

4.4.7.2 Method

SO₂

Emission figures of SO₂ are reported from the plant to the Norwegian Environment Agency based on continuous measurements. Flue gas treatment is installed at the plant.

NO_x

Emission figures of NO_x are annually reported from the plant to the Norwegian Environment Agency. The emission figures are based on calculations.

NH₃

Emission figures based on calculations are annually reported from the plant to the Norwegian Environment Agency.

Particles

Emission figures for particles have been reported to the Norwegian Environment Agency since 1992. Emissions in 1990 and 1991 are assumed to be the same as the reported figure in 1992. The emission permit sets requirements to emissions from the melting furnace, transport, crushing and packing of the raw materials and products. The Norwegian Environment Agency assumes that the particles emitted are smaller than PM_{2.5}. This means that TSP=PM₁₀=PM_{2.5} is used in the inventory.

Heavy metals and POPs

Emission figures for Cu have been reported to the Norwegian Environment Agency since 1990. Reported figures for Cd, Hg and Pb were available from 1990-1994, but because of low emissions the plant stopped reporting these metals.

4.4.7.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

The particle size distribution used is only an assumption and we can not preclude that the distribution might be different than the one suggested. The particle size distribution can therefore only be seen as an estimate.

4.4.7.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.4.8 Manufacture of anodes

NFR 2C7C

Last update: 27.02.19

4.4.8.1 Description

Four plants in Norway produce anodes. Three plants produce prebaked anodes and one plant produced coal electrodes. These are alternatives to the use of coal and coke as reducing agents

in the production process for aluminium and ferroalloys. The anodes and coal electrodes are produced from coal and coke. The production of anodes and coal electrodes leads to emissions of NO_x, SO₂, particles, BC, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and heavy metals.

4.4.8.2 Method

SO₂ and NO_x

Emission figures of SO₂ are based on measurements while NO_x emissions are calculated by the plants and reported to the Norwegian Environment Agency.

Particles

Production of anodes leads to emission of particles. One of the plants has reported emissions since 1990, while the other one has reported since 1992. Emission figures for 1990 and 1991 are assumed to be the same as the reported figure in 1992 for this plant. The Norwegian Environment Agency assumes that the particles emitted are smaller than PM₁₀, but also expects some to be smaller than PM_{2.5}. No information has been found regarding the particle size distribution, so in lack of other data we use the same distribution profile as used for production of aluminium where PM₁₀ is 97 per cent of TSP and PM_{2.5} is 43 per cent of TSP.

BC

Emissions have been estimated as a share of PM_{2.5} emissions. Measurements of the composition of the particulate matter at one plant showed that 8 per cent of the particulate matter was carbon. As no information on the share of BC and OC was found in the literature for anode production, BC share has been set to be 50 per cent of PM_{2.5} using the default method described in Aasestad (2013). Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM). BC emission have therefore been set to 4 per cent of PM_{2.5} emissions.

PAHs (Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)

Emission figures for PAH are based on measurements and reported from both plants to the Norwegian Environment Agency. One plant has developed a new and better method for measuring PAH. This method is used for the period 1992 to 2003. The reported figures of PAH are assumed to be according to the Norwegian standard (NS9815). Measurements from production of Soederberg paste (at three Norwegians plants) and a PAH-profile of baked anodes from EPA are used to derive a PAH-profile to find the emission of PAH-OSPAR and PAH-4. Based on these profiles it is assumed that PAH-4 account for 5 per cent of the total PAH emissions (Table 4.19). Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are estimated using the same PAH-profile as for aluminium production for the years 1990 - 2015, see Table 4.17. After 2015 two plants have reported figures for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene.

Table 4.19. Share of PAH-4 emissions of total PAH emisison from production of anodes.

| Component | Share of PAH emissions |
|----------------|------------------------|
| PAH-4 (CLRTAP) | 0.05 |

Source: Norwegian pollution control authority (SFT 1999a)

Heavy metals

Production of anodes leads to emission of heavy metals due to the metal content in the reducing agents (coke and coal). Emission figures are based on measurements and are reported for arsenic and mercury from one plant since 2001, and for lead since 2004. Emission figures have not been measured or reported before 2001 for As and Hg and before 2004 for Pb and are therefore not available for previous years. Historical emission figures back to 1990 are assumed to be the same as reported figures for 2001 for As and Hg and 2004 for Pb.

4.4.8.3 Uncertainties

Historical calculations of heavy metals from 1990 to 2001 are very uncertain since they are assumed to be the same as reported figures for the first year of reporting (2001). Annual changes in production volumes, coke quality and the amount of heavy metals in the reducing agents are not taken into account, and the historical emissions can only be seen as an estimate in lack of better data.

4.4.8.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5 Solvents and product use

NFR 2D, 2G

Last update: 02.01.18

Within solvents and product use, Norway includes emissions from solvent losses, creosote-treated materials, road paving with asphalt, mercury-containing products, tobacco and use of fireworks.

Use of solvents and products containing solvents result in emissions of non-methane volatile organic compounds (NMVOC). In addition to solvents emitting NMVOC, there are other products that emit other volatile components. Creosote treated materials and tarry jointing paste cause emissions of PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene). Dioxins are also emitted during road paving with asphalt (2D3B). Emissions of mercury from mercury-containing products as well as emissions from combustion of tobacco and the use of fireworks are also included in the Norwegian inventory.

Table 4.20. Solvents and product use. Components emitted and included in the Norwegian inventory.

| Solvents and product use | SO ₂ | NO _x | NH ₃ | NMVO C | CO | Particl es | BC | Heavy metals | Dioxin s | PAH |
|-----------------------------|-----------------|-----------------|-----------------|-----------|----|---------------|----|-----------------|-------------|-----|
| Solvent losses | NA | NA | NA | E | NA | NE | NA | NA | NA | NA |
| Creosote-treated materials | NE | NE | NE | E | NE | NE | NE | NE | NE | E |
| Road paving with asphalt | NE | NE | NA | E | NE | E | E | NA | NE | NE |
| Production of asphalt | NA | NA | NA | E | NE | E | E | NA | E | E |
| Leather preparing | NA | NA | R | E | NA | NA | NA | NA | NA | NA |
| Mercury-containing products | NA | NA | NA | NA | NA | NA | NA | E | NA | NA |
| Tobacco | NE | E | E | E | E | E | E | E | E | E |
| Use of fireworks | E | E | NE | NA | E | E | NA | E | NE | NE |

E = Figures estimated by Statistics Norway (Activity data * emission factor). R = Figures reported by the plant to the Norwegian Environment Agency. NA = Not Applicable.

Source: Statistics Norway/ Norwegian Environment Agency

4.5.1 Solvent losses (NMVOC)

NFR 2D3A, 2D3D, 2D3E, 2D3F, 2D3G, 2D3H, 2D3I.

Last update: 12.02.20

4.5.1.1 Method

The general model represents a mass balance *per substance*, where emissions are calculated by multiplying relevant activity data with an emission factor. For better coverage, point sources reported from industries to the Norwegian Environment Agency and calculated emissions from a side model for cosmetics, are added to the estimates. For a detailed description of method and activity data, see Holmengen and Kittilsen (2009). Activity data are reported in the NFR table and in Appendix F.

It is assumed that all products are used the same year as they are registered, and substances are not assumed to accumulate in long-lived products. In other words, it is assumed that all emissions generated by the use of a given product during its lifetime take place in the same year as the product is declared to our data source, the Norwegian Product Register. In sum, this leads to emission estimates that do not fully reflect the actual emissions taking place in a given year. Emissions that in real life are spread out over several years all appear in the emission estimate for the year of registration. However, this systematic overestimation for a given year probably more or less compensates for emissions due to previously accumulated amounts not being included in the estimate figures.

No official definition of solvents exists, and a list of substances to be included in the inventory on NMVOC emissions was thus created. The substance list used in the Swedish NMVOC inventory (Skårman et al. 2006) was used as a basis. This substance list is based on the definition stated in the UNECE Guidelines¹². The list is supplemented by NMVOC reported in the UK's National Atmospheric Emissions Inventory (NAEI) (AEA 2007). The resulting list comprises 678

¹² "Volatile compound (VOC) shall mean any organic compound having at 293.15 degrees K a vapor pressure of 0.01 kPa or more, or having a corresponding volatility under the particular conditions of use."

substances. Of these, 355 were found in the Norwegian Product Register for one or more years in the period 2005-2007.

Cosmetics

Cosmetics are not subject to the duty of declaration. The side model is based on a study in 2004, when the Norwegian Environment Agency calculated the consumption of pharmaceuticals and cosmetics (SFT 2005). The consumption was calculated for product groups such as shaving products, hair dye, body lotions and antiperspirants. The consumption in tonnes each year is calculated by using the relationship between consumption in Norwegian kroner and in tonnes in 2004. Figures on VOC content and emission factors for each product group were taken for the most part from a study in the Netherlands (IVAM 2005), with some supplements from the previous Norwegian solvent balance (the previous NMVOC emission model).

4.5.1.2 Activity data

The data source is the Norwegian Product Register. Any person placing dangerous chemicals on the Norwegian market for professional or private use has a duty of declaration to the Product Register, and import, export and manufacturing is reported annually. The only exception is when the amount of a given product placed on the market by a given importer/producer is less than 100 kg per year.

The information in the data from the Product Register makes it possible to analyse the activity data on a substance level, distributed over product types (given in UCN codes; (Norwegian Product Register 2007)), industrial sectors (following standard industrial classification (NACE)), including private households (no NACE), or a combination of both. As a consequence, the identification of specific substances, products or industrial sectors that have a major influence on the emissions is greatly facilitated.

Allocation

The data received from the Product register covers produced, imported and exported chemicals. The chemicals are allocated by CAS-nr, UCN-code and industrial sector giving

Cosmetics

The side model for cosmetics is updated each year with data on sales from the Norwegian Association of Cosmetics, Toiletries and Fragrance Suppliers (KLF).

Point sources

Data from nine point sources provided by the Norwegian Environment Agency are added to the emissions estimates. The point sources are reported from the industrial sector "Manufacture of chemicals and chemical products" (NACE 20). In order to avoid double counting, NMVOC used as raw materials in this sector are excluded from the emission estimates from the Product Register data.

4.5.1.3 Emission factors

Emission factors are specific for combinations of product type and industrial sector. Emission factors from the Swedish model for estimating NMVOC emissions from solvent and other product use (Skårman et al. 2006) are used. The emission factors take into account different

application techniques, abating measures and alternative pathways of release (e.g. waste or water). These country-specific emission factors apply to 12 different industries or activities.

It is assumed that the factors developed for Sweden are representative for Norwegian conditions, as we at present have no reason to believe that product types, patterns of use or abatement measures differ significantly between the two countries. Some adjustments in the Swedish emission factors were made when the model was first developed (see Holmengen and Kittilsen (2009)) and several improvements of single emission factors have been made in the following years.

In accordance with the Swedish model, emission factors were set to 0 kg/tonne solvent for a few products that are assumed to be completely converted through combustion processes, such as EP-additives, soldering agents and welding auxiliaries. Quantities that have not been registered to industrial sector or product type are given emission factor of 0.95 kg/tonne solvent. Emission factors may change over time, and such changes may be included in this model. However, all emission factors are at the moment constant for all years.

4.5.1.4 Uncertainties

Uncertainty in emission factors

The emission factors are more detailed in the new NMVOC model than in the previous model, as this model can take into account that emissions are different in different sectors and products, even when the substance is the same. However, for this to be correct, a thorough evaluation of each area of use is desirable, but not possible within a limited time frame. Thus, the emission factor is set with general evaluations, which leads to uncertainty.

The emission factors are taken from several different sources, with different level of accuracy. The uncertainties in emission factors depend on how detailed assessment has been undertaken when the emission factor was established. Some emission factors are assumed to be unbiased, while others are set close to the expected maximum of the range of probable emission factors. This, together with the fact that the parameter range is limited, gives us a non-symmetrical confidence interval around some of the emission factors. For each emission factor we thus have two uncertainties; one negative (n) and one positive (p). These are aggregated separately, and the aggregated uncertainty is thus not necessarily symmetrical.

Uncertainty in activity data

For the activity data, the simplified declarations and the negative figures due to exports lead to known overestimations, for which the uncertainty to a large extent is known. A more elaborate problem in calculations of uncertainty is estimating the level of omissions in declaration for products where the duty of declaration does apply. In addition, while declarations with large, incorrect consumption figures are routinely identified during the QA/QC procedure, faulty declarations with small consumption figures will only occasionally be discovered. There is however no reason to believe that the Product Register data are more uncertain than the data source used in the previous model (statistics on production and external trade), as similar QA/QC routines are used for these statistics.

The errors in activity data are not directly quantifiable. Any under-coverage in the Product Register is not taken into account. Skårman et al. (2006) found that the activity data from the Swedish Product register had an uncertainty of about 15 per cent. The Norwegian Product Register is assumed to be comparable to the Swedish, and thus the uncertainty in the activity data is assumed to be 15 per cent. For some products, simplified declarations give an indication of maximum and minimum possible amounts. In these cases, the maximum amount is used, and the positive uncertainty is set to 15 per cent as for other activity data, while the negative uncertainty is assumed to be the interval between maximum and minimum amount. All activity data are set to zero if negative.

For a detailed description of the uncertainty analysis, see Holmengen and Kittilsen (2009). The variance of total emission was estimated from the variance estimates obtained for emission factors and activity data, using standard formulas for the variance of a sum and the variance of a product of independent random variables. The aggregated uncertainties in level and trend are given in Table 4.21 and Table 4.22.

Table 4.21 Uncertainty estimates for level of NMVOC emissions, 2005-2007. Tonnes and per cent

| Uncertainty in level | Negative (n) | Negative (n) (per cent of total emissions) | Positive (p) | Positive (p) (per cent of total emissions) |
|-----------------------------|---------------------|---|---------------------|---|
| 2005 | 2 288 | 4.58 | 1 437 | 2.88 |
| 2006 | 1 651 | 3.70 | 1 103 | 2.47 |
| 2007 | 1 299 | 2.79 | 1 168 | 2.51 |

Source: Holmengen and Kittilsen (2009)

Table 4.22. Uncertainty estimates for trend in NMVOC emissions, 2005-2007. Tonnes

| Uncertainty in trend | Negative (n) | Positive (p) | 95% confidence interval for change |
|-----------------------------|---------------------|---------------------|---|
| 2005-2006 | 2 135 | 1 067 | (-7 366, -4 164) |
| 2006-2007 | 1 420 | 947 | (407, 2 774) |
| 2005-2007 | 1 882 | 1 076 | (-5 286, -2 328) |

Source: Holmengen and Kittilsen (2009)

4.5.1.5 Source specific QA/QC

Large between-year discrepancies in the time series of substance quantities are routinely identified and investigated, in order to correct errors in consumption figures. Large within-year discrepancies between minimum and maximum quantities in simplified declarations are routinely identified and investigated, in order to prevent overestimation for substances where consumption figures are given in intervals. Large within-year discrepancies between totals for industrial sectors (NACE) and totals for products (UCN) are routinely identified and investigated, in order to detect erroneous or incomplete industrial sectoral and product type distribution.

4.5.2 Road paving with asphalt

NFR 2D3B

Last update: 17.01.20

Road paving with asphalt

4.5.2.1 Method

The emissions from road paving are being calculated in accordance with a Tier 1 approach (EEA 2013) for NMVOC, TSP, PM₁₀ and PM_{2.5}. Emissions of BC and dioxins from production of asphalt are also included.

$$E_{\text{pollutant}} = AR_{\text{production}} * EF_{\text{pollutant}}$$

Where:

$E_{\text{pollutant}}$ = the emission of the specified pollutant

$AR_{\text{production}}$ = the activity rate for the road paving with asphalt

$EF_{\text{pollutant}}$ = the emission factor for this pollutant

Dioxins

Asphalt preparations and asphalt recycling are supposed to be a possible dioxin source, especially in countries using extensive recycling, and that use salt on the roads during winter. A lot of salt is used on Norwegian roads during winter, and when this asphalt is heated during recycling, it is assumed to give emissions of dioxins (Hansen 2000).

4.5.2.2 Activity data

The activity data used is the annual weight of asphalt used for road paving in Norway. EBA, *pers. comm*¹³).

4.5.2.3 Emission factors

Emissions of NMVOC, TSP, PM₁₀ and PM_{2.5} from road paving with asphalt are estimated using Tier 1 emission factors from the 2019 EEA Guidebook.

Table 4.23. Emission factor for road paving with asphalt. g/tonn

| | |
|-------------------|--------|
| NMVOC | 16 |
| TSP | 14 000 |
| PM ₁₀ | 3 000 |
| PM _{2.5} | 400 |

Source: EEA (2019)

BC

Emissions are estimated from a share of PM_{2.5} emissions. As no share for BC was found in the literature, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

¹³ EBA (2014): Expert judgement by Contractors Association - Building and Construction (EBA), Oslo, Norway

Dioxins

Two emission factors are found in the literature. According to SFT (2001), the Oslo and Paris Convention (OSPAR) suggests an emission factor of 0.047 µg/tonne asphalt. This emission factor is however assumed to be very high since it is based on data from a plant only re-circulating old asphalt. Fyns Amt (2000) operates with a much lower emission factor, which probably reflects dioxin emissions from preparation of new asphalt. Since Norway both makes new asphalt and recycles old asphalt, it is assumed that an emission factor in between those suggested from OSPAR and Fyns Amt would be most correct for Norwegian conditions (Table 4.24).

Table 4.24. Dioxin emission factor for asphalt production. µg I-TEQ/tonne produced asphalt

| Source | Emission factor |
|------------------------|-----------------|
| SFT (2001) | 0.047 |
| Fyns Amt (2000) | 0.0022 |
| Emission factor chosen | 0.025 |

4.5.2.4 Uncertainties

The activity data used are uncertain. The emission factors used are also uncertain. The annual emissions are low however, and will not have any impact on the total level of these emissions.

4.5.2.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.3 Leather preparing

NFR 2D3G

Last update: 10.02.20

4.5.3.1 Method

NH₃

NH₃ is used to adjust the pH level in the fattening and colouring process in leather preparing. This means that NH₃ is dissolved in an aqueous solution to feed fatty substances to leather. One plant reports emission figures for NH₃ to the Norwegian Environment Agency. Emission figures are available from 1994. Emissions for the years 1990-1993 are assumed by Statistics Norway and the Norwegian Environment Agency to be the same as the reported figure for 1994. The emission of NH₃ reported by the plant is equal to the consumption of NH₃.

4.5.3.2 Uncertainties

It is not clear if it is correct to assume that all NH₃ consumed is emitted to air. This assumption may have to be revised.

4.5.3.3 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.4 Creosote-treated materials

NFR 2D3I

Last update: 10.02.20

4.5.4.1 Description

Creosote is mainly used in quay materials and conductor poles, but also in fence poles and roof boards. In Norway there is a requirement that all creosote in use should contain less than 50 mg/kg benzo(a)pyrene (Ministry of the Environment 2004). PAH-components will evaporate from the creosote-treated materials in hot weather. In addition, PAH-components will evaporate during impregnation. The smallest PAH-components, like naphthalene, are most volatile, but several components used in wood treatment will not evaporate.

4.5.4.2 Method

Emissions of PAHs benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are calculated based on the import of creosote oil and emission factors. For simplicity, it is assumed that all PAH is emitted the same year as the materials are produced.

4.5.4.3 Activity data

Data on imported amounts of creosote oil are taken from Statistics Norway's statistics on external trade.¹⁴

4.5.4.4 Emission factors

The emission factor used is taken from Finstad et al. (2001). It is assumed that imported creosot oil contains on average 55 per cent PAH and that one per cent will evaporate during the lifetime of the creosot-treated materials. It is assumed that PAH-4 accounts for 0,018 per cent of the total PAH emissions. Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are estimated using the a PAH-profile for creosot oil, Finstad at al (2001), see Table 4.25.

Table 4.25. Distribution of PAH-4 emissions from creosote oil. Share of PAH-4

| Component | Distribution of PAH emissions (ratio) |
|------------------------|---------------------------------------|
| benzo(a)pyrene | .. |
| benzo(b)fluoranthene | 0.50 |
| benzo(k)fluoranthene | 0.50 |
| indeno(1,2,3-cd)pyrene | .. |

Source: Finstad et al (2001)

4.5.4.5 Uncertainties

In the inventory it is assumed that all PAH is emitted the same year as the materials are used. This is however not the case, since PAH will be emitted as long as the creosote-treated materials are in use. However, most of it is likely to be emitted during the first years.

¹⁴ <https://www.ssb.no/en/utenriksokonomi/statistikker/muh/aar>

4.5.4.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.5 Mercury-containing products

NFR 2G

Last update: 05.02.15

4.5.5.1 Method

Breakage of mercury-containing thermometers, fluorescent tubes, economy bulbs, and various measuring and analytical instruments lead to emissions of mercury. The emission estimates are based on an annual report from the Norwegian Environment Agency ("Miljøgifter i produkter"). The sale of mercury-containing thermometers and fluorescent tubes has decreased strongly since the mid-1990s, and the mercury content in these products has been reduced. A prohibition against the production, import and export of mercury-containing products entered into force in 1998, except for some thermometers for professional use, which were prohibited in 2001. Since these products have long operating life times, there will be emissions from these products for many years. In the calculations, however, it is assumed that the emissions occur the same year as the product is sold.

For thermometers, it is assumed that all mercury is emitted in hospitals, despite some breakage of mercury-containing thermometers that occur in households. For fluorescent tubes and economy bulbs, all emissions are placed in households, although emissions occur in all sectors. For measuring and analytical instruments, all emissions are placed under research and development work.

4.5.5.2 Uncertainties

The emissions are assumed to be emitted the same year as the products are sold. This is not accurate, since most of these products have long operating life times. It is however impossible to predict the annual breakage and the mercury content in each of them.

4.5.5.3 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.6 Tobacco

NFR 2G

Last update: 17.01.20

4.5.6.1 Method

NO_x, NMVOC, CO, NH₃, particles, BC, heavy metals and POPs

The emission components included from the combustion of tobacco are NO_x, NMVOC, CO, NH₃, particles, heavy metals and POPs (Persistent organic pollutants). Emission figures have been calculated by multiplying the annual consumption of tobacco with emission factors for each pollutant.

4.5.6.2 Activity data

The total consumption of tobacco in Norway is given by the net import of tobacco from Statistics Norway's external trade statistics and are reported in Appendix F. Tobacco bought tax free abroad and tobacco smuggled are not included in the inventory.

Emission factors

Table 4.26 gives emission factors used for tobacco combustion. For NO_x, NMVOC and CO the emission factors are calculated by Statistics Norway, based on values given in Norwegian Directorate of Health (1990).

Table 4.26. Emission factors used for tobacco combustion

| | Tobacco (unit/kg tobacco) | Source |
|----------------------------|----------------------------------|---|
| NO _x (kg) | 0.0034652 | Statistics Norway, Norwegian Directorate of Health (1990) |
| NMVOC (kg) | 0.0048374 | Statistics Norway, Norwegian Directorate of Health (1990) |
| CO (kg) | 0.1215475 | Statistics Norway, Norwegian Directorate of Health (1990) |
| NH ₃ (kg) | 0.00415 | EEA (2013) |
| TSP (kg) | 0.04 | Institute of environmental and energy technology (2002) |
| PM ₁₀ (kg) | 0.04 | Institute of environmental and energy technology (2002) |
| PM _{2.5} (kg) | 0.04 | Institute of environmental and energy technology (2002) |
| BC | 0.5% of PM _{2.5} | IIASA (Kupiainen & Klimont 2004) |
| Pb (g) | 0.00005 | Finstad et al. (2001) |
| Cd (g) | 0.0001 | Finstad et al. (2001) |
| Hg (g) | 0.00001 | Finstad et al. (2001) |
| As (g) | 0.000159 | Finstad and Rypdal (2003) |
| Cr (g) | 0.000125 | Finstad and Rypdal (2003) |
| Cu (g) | 0.000354 | Finstad and Rypdal (2003) |
| Benzo(a)pyrene (g) | 0.000111 | EEA (2019) |
| Benzo(b)fluoranthene (g) | 0.000045 | EEA (2016) |
| Indeno(1,2,3-cd)pyrene (g) | 0.000045 | EEA (2016) |
| Indeno(1,2,3-cd)pyrene (g) | 0.000045 | EEA (2016) |
| Dioxins (µg) | 0.0013 | Finstad et al. (2002b) |

4.5.6.3 Uncertainties

The emissions are assumed to be emitted the same year as the products are imported.

4.5.6.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.7 Use of fireworks

NFR 2G

Last update: 17.01.20

4.5.7.1 Method

The emission components included from the use of fireworks are SO₂, CO, NO_x, particles and heavy metals. Emission figures have been calculated by multiplying the annual use of fireworks with emission factors for each pollutant.

4.5.7.2 Activity data

The total use of fireworks in Norway is given by the net import of fireworks from Statistics Norway's external trade statistics. The use of fireworks is reported in Appendix F.

4.5.7.3 Emission factors

Table 4.27 gives emission factors used for the use of fireworks.

Table 4.27. Emission factors used for the use of fireworks

| | Value (g/t fireworks) |
|-------------------|-----------------------|
| SO ₂ | 3020 |
| CO | 7150 |
| NO _x | 260 |
| TSP | 109 830 |
| PM ₁₀ | 99 920 |
| PM _{2.5} | 51 940 |
| As | 1.33 |
| Cd | 1.48 |
| Cr | 15.6 |
| Cu | 444 |
| Hg | 0.057 ¹ |
| Pb | 784 ² |

¹) Emissions of Hg from fireworks assumed banned in 2002 like in Denmark.

²) Emissions of Pb from fireworks assumed banned in 2006 like in Denmark.

Source: EEA (2019)

4.5.7.4 Uncertainties

The emissions are assumed to be emitted the same year as the products are imported.

4.5.7.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.6 Other production

NFR 2H

Within other production, Norway includes emissions from pulp and paper, food and beverages industry and ore mines.

Table 4.28. Other production. Components emitted and included in the Norwegian inventory.

| Other production | SO ₂ | NO _x | NH ₃ | NMVOC | CO | Particles | BC | Heavy metals | Dioxins |
|-----------------------------|-----------------|-----------------|-----------------|-------|----|-----------|----|--------------|---------|
| Pulp and paper | R | E | NE | E | E | R | E | NA | NA |
| Food and beverages industry | NA | NA | NA | E | NA | NE | NE | NA | NA |
| Ore mines | R | NA | NA | NA | NA | R | NA | NA | R |

E = Figures estimated by Statistics Norway (Activity data * emission factor). R = Figures reported by the plant to the Norwegian Environment Agency. NA = Not Applicable.

Source: Statistics Norway/ Norwegian Environment Agency

4.6.1 Pulp and paper

NFR 2H1

Last update: 17.01.20

4.6.1.1 Description

Pulp and paper production has three major processing steps; pulping, bleaching and paper production. Kraft (sulphate) pulping is the most widely used pulping process and is generally used to produce strong paper products. The Kraft pulping process includes bleaching, chemical recovery and by-products recovery. The sulphite pulping is another chemical pulping process. It produces a weaker paper than some other types of pulping, but the pulp is less coloured, making it more suitable for printing, often with little bleaching.

In Norway, SO₂ and particles are reported emitted from production of chemical pulp and paper. In the Kraft pulping process, sodium sulphide and sodium hydroxide are used to chemically dissolve the lignin that binds the cellulose fibres, and in the acid sulphite pulping process, sulphurous acid solution is used. SO₂ is emitted in these processes. Emissions of NO_x, NMVOC and CO are estimated.

4.6.1.2 Method

SO₂

Emission figures are reported from producers of chemical pulp to the Norwegian Environment Agency. SO₂ is measured continuously and emission estimates are made from these measurements.

Particles

Four plants producing pulp and paper report non-combustion emissions of particles to the Norwegian Environment Agency. Two of these plants have not reported emission figures from combustion and it is assumed that the reported non-combustion emission figures include emissions from combustion. Some plants lack data for the earliest years, and emissions for those years are assumed to be the same as in the first year of reporting.

Two of the plants state that they clean the emissions by electric filter and wet scrubbers, and it is assumed by the Norwegian Environment Agency that the particles emitted are smaller than PM_{2.5}. The other two clean their emissions using only wet scrubbers, and it is assumed the particles are smaller than PM₁₀. According to EEA (2019), PM_{2.5} is 60 per cent of TSP and PM₁₀ is 80 per cent of TSP.

NO_x, NMVOC and CO

Emissions of NO_x, NMVOC and CO are estimated based on annual production levels and emission factors.

BC

BC emissions have been estimated using a share of PM_{2.5} from EEA (2019) as emission factor. BC=2.6 per cent of PM_{2.5}.

4.6.1.3 Activity data

For the estimates of NO_x, NMVOC and CO, production levels of pulp by different processing steps as reported by the plants are used.

4.6.1.4 Emission factors

For the estimates of NO_x, NMVOC and CO, emission factors from the 2019 Guidebook are used.

Table 4.29. Emission factors for pulp and paper. kg/Mg air dried pulp

| | |
|-----------------|--|
| NO _x | 1 (Kraft), 2 (Acid sulphite) |
| NMVOC | 2 (Kraft), 0.2 (Acid sulphite), 0.05 (neutral sulphite semi) |
| CO | 5.5 (Kraft) |

Source: EEA (2019)

Shares given by IIASA (Kupiainen & Klimont 2004) have been used to estimate BC emissions.

4.6.1.5 Uncertainties

The particle size distribution used is not plant specific and might therefore be different from the one suggested by TNO.

4.6.1.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.6.2 Food and beverages industry

NFR 2H2

Last update: 17.01.20

4.6.2.1 Description

This source category includes NMVOC emissions from production of bread and beer. Emissions of NMVOC from spirit manufacture are considered insignificant and are not included in the inventory.

4.6.2.2 Method

NMVOC

Production of bread and beer (and other similar yeast products) involves fermentation processes that lead to emission of NMVOC (ethanol). Emissions are calculated based on production volumes and emission factors.

4.6.2.3 Activity data

Production volumes of bread and beverages are annually reported to Statistics Norway and these are reported in Appendix F.

4.6.2.4 Emission factors

The emission factors are taken from EEA (2019).

Table 4.30. NMVOC emission factors from production of bread and beverage

| | Emission factor | Unit |
|------------------------|-----------------|------------------------|
| Production of bread | 0.0045 | tonnes/tonnes produced |
| Production of beverage | 0.35 | kg/1000 litres |

Source: EEA (2019)

4.6.2.5 Uncertainties

The emission factors used are recommended by EEA (2019) and are not specific for Norwegian conditions.

4.6.2.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.6.3 Ore mines

NFR 2H3

Last update: 29.02.19

4.6.3.1 Description

Three ore mines are included in the Norwegian Inventory, but one of the mines was closed down in 1996. Emission figures of SO₂, particles and dioxins are included. The treatment of ore generates emissions of SO₂, and particles are also emitted. Dioxin emissions are due to the thermal process during the pellet production. The ore mine which closed down in 1996, had large dioxin emissions due to the thermal process during the pellet production.

SO₂ emissions are only included in the inventory for the ore mine that was closed down in 1996. The SO₂ emissions from the two other ore mines are not included in the inventory.

4.6.3.2 Method

SO₂

The ore mine which was closed down in 1996, reported emission figures for SO₂ to the Norwegian Environment Agency. None of the two other ore mines report any non-combustion SO₂ emissions.

Particles

All the three ore mines report emission figures for particles to the Norwegian Environment Agency. Emissions for the two existing ore mines are reported from respectively 1994 and 1996. Emissions for earlier years are assumed to be the same as in the first year of reporting.

The size distribution used in the Norwegian inventory is according to TNO (Institute of environmental and energy technology 2002) (Table 4.31).

Table 4.31. Particle size distribution for particles emitted from ore mining. Ratio X^1/TSP

| Component | Particle size distribution (ratio) |
|-------------------|------------------------------------|
| TSP | 1 |
| PM ₁₀ | 0.49 |
| PM _{2.5} | 0.07 |

¹ X is either PM_{2.5}, PM₁₀ or TSP.

Source: TNO (Institute of environmental and energy technology 2002).

Dioxins

Emissions of dioxins are only included for the ore mine which was closed down in 1996. Emission figures were first reported to the Norwegian Environment Agency in 1994 and emissions for previous years have been assumed to be the same as the reported figure in 1994.

4.6.3.3 Uncertainties

The size of the particles emitted from ore mining will also depend on the type of ore and production process. The particle size distribution used in the inventory does not consider these differences.

4.6.3.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.7 Wood processing

NFR 2I

Table 4.32. Wood production. Components emitted and included in the Norwegian inventory

| Other production | SO ₂ | NO _x | NH ₃ | NMVOC | CO | Particles | TSP |
|------------------|-----------------|-----------------|-----------------|-------|----|-----------|-----|
| Wood production | NE | NE | NE | NE | NE | NE | E |

E = Figures estimated by Norwegian Environment Agency (Activity data * emission factor). NE = Not Estimated

Source: Statistics Norway/ Norwegian Environment Agency

4.7.1 Wood processing

NFR 2I

Last update: 13.02.20

4.7.1.1 Description

This source category includes TSP emissions from four plants from wood processing.

4.7.1.2 Method

The emissions are calculated based on production volumes and emission factors.

4.7.1.3 Activity data

The production volumes of wood processing products are annually reported to the Norwegian Environment Agency. These are reported in the NFR table and in Appendix F.

4.7.1.4 Emission factors

The emission factor is taken from 2019 EEA Guidebook.

Table 4.33. TSP emission factor for wood processing

| | Emission factor | Unit |
|-----------------|-----------------|--------------------|
| Wood processing | 1 | Kg/Mg wood product |

Source: EEA (2019)

4.7.1.5 Uncertainties

The emission factor is not specific for Norwegian conditions.

4.7.1.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

5 AGRICULTURE (NFR sector 3)

NFR 3

5.1 Overview

Agriculture is an important contributor to NH_3 emissions. Animal manure management, grazing animals and the use of fertilizer (manure, synthetic fertilizer, sewage sludge and other organic fertilizers applied to soils) generate emissions of NH_3 . Another source of NH_3 is treatment of straw using NH_3 as a chemical.

Animal manure management and the use of fertilizer (manure, synthetic fertilizer, sewage sludge and other organic fertilizers applied to soils) also generates emissions of NO_x .

Emissions of NMVOC from manure management and from cultivated crops are also included in the inventory.

Non-combustion emissions of particles from manure management and agricultural soils are also calculated. Additionally, there are long-range transboundary air emissions arising from the burning of agricultural residues. Figure 5.1 shows the trends for most of the long-range transboundary air pollutants from the agricultural sources.

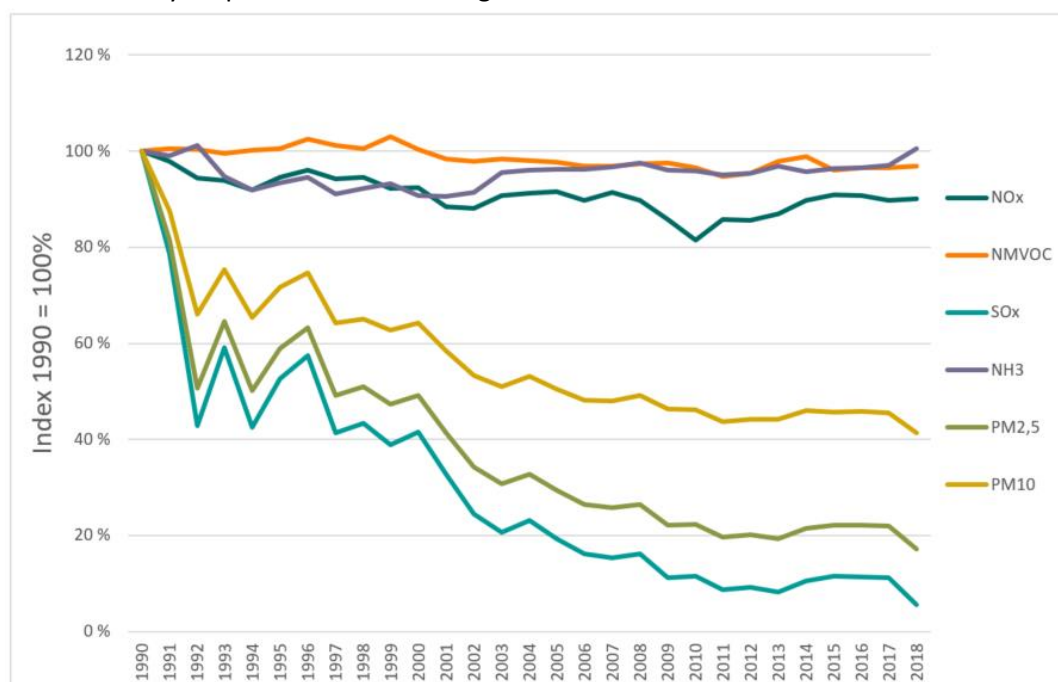


Figure 5.1. Trends for the emissions for most of the air pollutants from the agricultural sector, relative to 1990

While the trends for NH_3 , NMVOC and NO_x have been quite stable since 1990, the emissions from PM₁₀, PM_{2,5} and SO_x have had a considerable decline. The main reason for this decline is primarily due to reduced amount of straw burned since 1990.

The total emissions of NH_3 from agriculture have been relatively stable and are in 2018 about 3% higher than in 1990. Figure 5.2 and Figure 5.3 shows the NH_3 trends for the different agriculture sources.



Figure 5.2. Trends for the NH_3 emissions for agricultural sources. 1000 tonnes NH_3 . 1990-2018

Source: Statistics Norway/ Norwegian Environment Agency

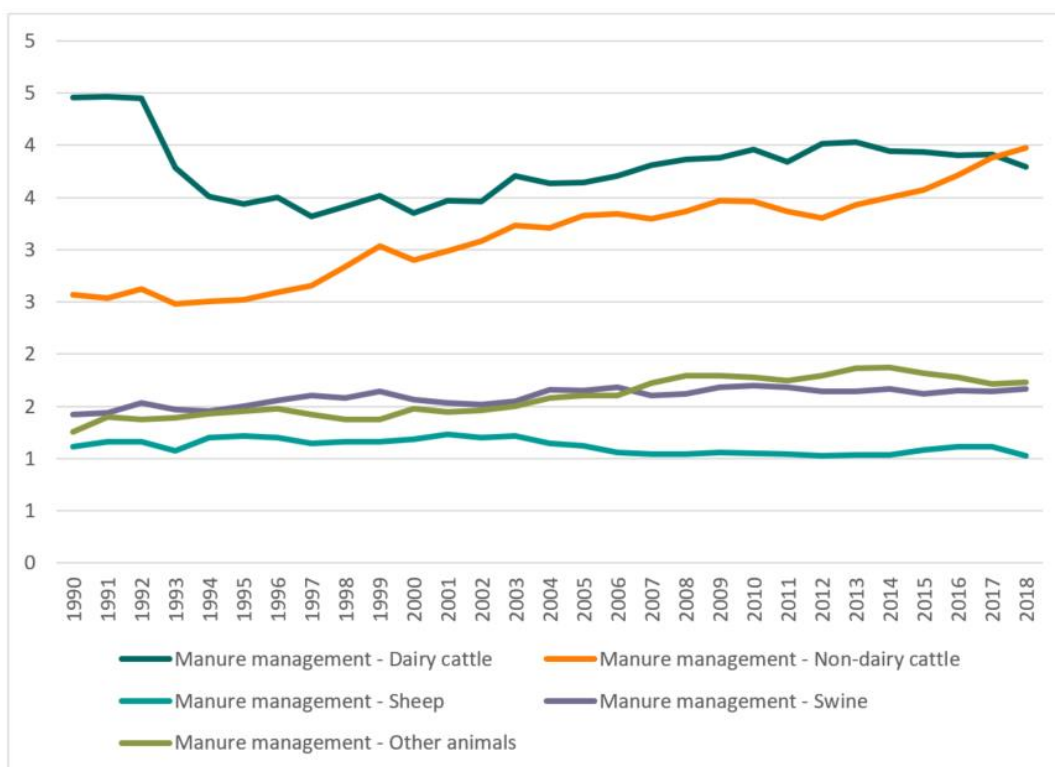


Figure 5.3: Trends for the NH_3 emissions from manure management. 1000 tonnes NH_3 . 1990-2018

Source: Statistics Norway/ Norwegian Environment Agency

The increase of NH₃ emissions from manure management for cattle since second half of 1990 is due to a increase in the number of animals. The strong increase in other agricultural emissions in 2018 can be explained by a increase in NH₃ treatment of straw because of a unusual hot and dry summer. The irregular emission trend from inorganic fertilizers around 2009 is due to a price change and is explained in more detail in chapter 5.5.2.2.

See chapter 2 for more information about trends.

5.2 Livestock population characterisation

Last update: 06.02.20

5.2.1 Data sources

The main sources of the livestock statistics are the register of production subsidies (sheep for breeding, goats, breeding pigs, poultry for egg production and beef cows), statistics of approved carcasses (animals for slaughter) and the Cow Recording System at TINE BA¹⁵ (heifers for breeding and dairy cows). The animal numbers from production subsidies are corrected with the estimated coverage of animal populations before Statistics Norway receive the data. This means that the figures used in the calculations represent the total population. The number of dairy cows and heifers for breeding derive from the Cow Recording Systems (TINE BA Annually). Between 98 and 99 per cent of all dairy cows are registered here, and in addition, the number used in the inventory is adjusted for this missing part. The adjustment is based on the percentage of herds controlled by the cow recording system. The coverage in the data sources used is shown in Table 5.1.

¹⁵ TINE BA is the sales and marketing organisation for Norway's dairy cooperative and covers most of the milk production and the meat production induced by milk production.

Table 5.1. Estimated coverage (%) of animal populations in the data sources used. 2018

| Animal categories | Statistics Norway, production subsidies | Statistics Norway, statistics of approved carcasses ⁴ | TINE | Other |
|-----------------------------|---|--|-------------------|---------------------|
| Dairy cows | | 100 ² | 96.4 ¹ | |
| Heifers for breeding | | | 96.4 ¹ | |
| Young cattle for slaughter | | 100 | 100 ³ | |
| Beef cows | 99.8 | 100 ² | | |
| Sheep | 99.6 | 100 | | |
| Goats | 100 | | | |
| Laying hens | 99.9 | | | |
| Chicks for breeding | 100 | | | |
| Chicken for slaughter | | 100 | | |
| Other poultry for breeding | 96.5 | | | |
| Other poultry for slaughter | | 100 | | |
| Sows | 97.8 | | | |
| Young pigs for breeding | 100 | | | |
| Pigs for slaughter | | 100 | | |
| Horses | Unkown ⁵ | | | Unkown ⁵ |
| Fur-bearing animals | 100 | | | |
| Deer | 100 | | | |
| Reindeer | | | | 100 ⁶ |

¹ Share of livestock herds.² Data source only for slaughter weight³ Data source only for slaughter age⁴ Figure refers to share of slaughtered animals, excluding home slaughter. Animals dead from other causes also excluded⁵ Total number of horses used in the inventory is based on data from productions subsidies (roughly 50 per cent of total number) and an additional estimation of number of horses outside agriculture by NIBIO.⁶ Norwegian Agriculture Agency

Source: Estimations by Statistics Norway and the Cow Recording System (dairy cows and heifers).

The statistics of approved carcasses covers close to 100 per cent of all slaughtered animals. Home slaughter is not included, but the extent of home slaughter is very low due to legal restrictions. Even animals consumed by producers are in most cases registered at the slaughterhouses. The number of dairy cows and heifers for breeding derive from the Cow Recording Systems. Between 98 and 99 per cent of all dairy cows are assumed to be registered here.

The registers are updated annually. In addition to the animals included in these registers, an estimate of the number of other horses is obtained from the Norwegian Institute of Bioeconomy Research (NIBIO)¹⁶. The number of reindeer is obtained from the Norwegian Reindeer Husbandry Administration.

For the categories of animals living shorter than a year or two, generally animals for slaughter, lifetime is taken into account to get a yearly average for the number of animals.

5.2.2 Method for estimating number of cattle

For dairy cows, additional information from the Cow Recording System concerning annual milk production and proportion of concentrate in the diet is used (TINE BA Annually). The Cow Recording System also supplies annual information about slaughter age for heifers and bulls and data for estimating live weight of dairy cows and heifers for breeding, and also the age of young cows at their first calving.

For heifers and bulls for slaughter, animal numbers are based on data from statistics of approved carcasses which provide data on numbers slaughtered and slaughter weights. Combined with slaughter age from the Cow Recording System (TINE BA Annually), this gives a precise estimation of animal life time for each animal slaughtered. One principal draw-back of this method for estimating animal population is that emissions in all stages of these animals' lives will be accounted for in the year of slaughter, even though the emissions in the early stages of the lives of these animals to a large extent took place in the previous year. In a stable population of animals, this error is automatically adjusted for. Since animal populations are relatively stable, this error is considered much smaller compared to errors related to estimating animal year based on animal populations in the register of production subsidies which was previously used. The data sources used also ensure a better coherence between animal numbers, life time and weight. Estimated animal years for cattle are given in Table 5.2.

The number of milk cows calving their first time (=heifers for replacement) and their average age at time of calving is reported by the Cow Recording System (TINE BA Annually) on request from Statistics Norway. These data date back to 2004. For the years 1990-2003, average fraction (number of heifers)/(number of milk cows) for the years 2004-2011 is used to estimate number of heifers based on number of milk cows. Number of heifers for replacement in beef production is collected from annual reports from Animalia (Norwegian Meat and Poultry Research Center (www.animalia.no)). Figures exist from 2007. For previous years, the number is estimated with the same method as for heifers for milk production.

¹⁶ Former named the Agricultural Economics Research Institute (NILF).

Table 5.2 Estimated animal years for cattle

| | Heifers for replacement | Heifers for slaughter | Bulls for slaughter | Beef cows ¹ | Dairy cows |
|------|-------------------------|-----------------------|---------------------|------------------------|------------|
| 1990 | 311 279 | 47 020 | 289 945 | 8 193 | 325 896 |
| 1995 | 299 284 | 47 103 | 284 237 | 20 334 | 310 346 |
| 2000 | 280 121 | 63 512 | 285 349 | 42 324 | 284 880 |
| 2005 | 255 862 | 57 619 | 263 170 | 54 841 | 255 663 |
| 2006 | 246 711 | 58 446 | 255 963 | 55 706 | 250 903 |
| 2007 | 235 282 | 56 607 | 247 578 | 57 609 | 246 624 |
| 2008 | 240 399 | 54 831 | 238 111 | 60 401 | 238 550 |
| 2009 | 247 902 | 53 397 | 235 689 | 63 803 | 235 480 |
| 2010 | 239 839 | 53 410 | 230 872 | 67 110 | 232 294 |
| 2011 | 239 007 | 48 778 | 223 536 | 68 539 | 224 721 |
| 2012 | 235 891 | 42 863 | 217 050 | 71 834 | 229 767 |
| 2013 | 239 386 | 47 294 | 220 401 | 70 969 | 225 163 |
| 2014 | 244 601 | 67 624 | 208 979 | 73 894 | 221 032 |
| 2015 | 238 485 | 64 814 | 206 328 | 77 408 | 217 576 |
| 2016 | 241 173 | 64 361 | 217 885 | 84 372 | 215 015 |
| 2017 | 249 229 | 43 501 | 250 630 | 88 332 | 217 318 |
| 2018 | 245 428 | 52 356 | 260 129 | 92 304 | 211 523 |

¹ Counted animals

Source: Cow Recording System at TINE BA (dairy cows), slaughter statistics and estimations by Statistics Norway

5.2.3 Method for estimating number of sheep

In the estimations, the sheep population is divided between sheep > one year and sheep < one year. Data from both the register of production subsidies and slaughter statistics is used in estimating the number of animals. In 2017 two changes appeared in the register data: The counting date changed from 1. of January to 1. of March, and the two categories sheep > one year and sheep < one year were merged into one category for adult sheep. To solve this, figures are split into the two categories sheep > and < one year based on data from the Norwegian Meat and Poultry Center, Animalia.

Sheep over one year is estimated as the number of sheep registered 1. of March deducted for the number of sheep slaughtered before May. The sheep slaughtered later in the year are counted as living the whole year.

Sheep under one year is estimated as number of sheep under one year registered 1. of March + number of lambs slaughtered June-December *143/365. Lambs slaughtered before June are assumed to be registered as sheep under one year the 1. of March. Practically all lambs slaughtered after June are born in the spring. An expert judgment suggests an average lifetime

of 143 days for slaughtered lambs born in the spring (UMB, *pers. comm*¹⁷).

5.2.4 Deviations from FAO statistics

There are some differences between the number of animals used in these calculations and the FAO statistics. The general reason that animal statistics used in the emission inventory differ from the statistics delivered to FAO is that the statistics are used for different purposes. Animal statistics used in the inventory has to be categorized so that the categories fit the recommended methodology and the various emission factors used in the emission estimations. The figures reported to the FAO are provided by the Norwegian Institute of Bioeconomy Research (NIBIO)¹⁸. NIBIO makes an overall estimation for the agricultural sector, which is the basis for the annual negotiations for the economic support to the sector. This estimate includes a grouping of all agricultural activities, comprising area, number of animals and production data. Differences include:

- Different emphasis on the dates for counting.
- NIBIO does not register pigs under 8 weeks, whilst Statistics Norway (SN) does. For the number of animals for slaughter, SN uses the statistics of approved carcasses and estimates animal years (average population through the year) on basis of this, while NIBIO uses figures for registered animals at specific dates.
- For the number of dairy cows and heifers for replacement, Statistics Norway uses statistics from the Cow Recording System (TINE BA Annually)

5.2.5 Uncertainties

Activity data

The uncertainty in the data is considered to be within ± 5 per cent. There is also an uncertainty related to the fact that some animals are only alive part of the year and how long this part is.

5.2.6 Source specific QA/QC

In 2001, a project was initiated to improve the estimate of the number of animals. This was completed in 2002. In 2012, a further revision of the numbers of bulls and heifers was implemented. In 2016, the method for estimating the number of sheep was revised. The revised data on animal populations form the basis for the emission calculations for all years.

5.3 Nitrogen in animal manure as basis for emission estimates

Access to nitrogen is vital for all plant growth; hence nitrogen is added to the soil from i.a. animal manure. This causes emissions to air of compounds containing nitrogen at various points. Of the nitrogen compounds emitted to air from animal manure, N_2O , NO_x and NH_3 are estimated.

¹⁷ UMB (2001): Expert judgement by Department of Animal Science, Ås: Norwegian University of Life Sciences.

¹⁸ Former named the Agricultural Economics Research Institute (NILF).

According to the IPPC and LRTAP guidelines, process emissions of nitrogen compounds from use of animal manure are calculated from the following sources:

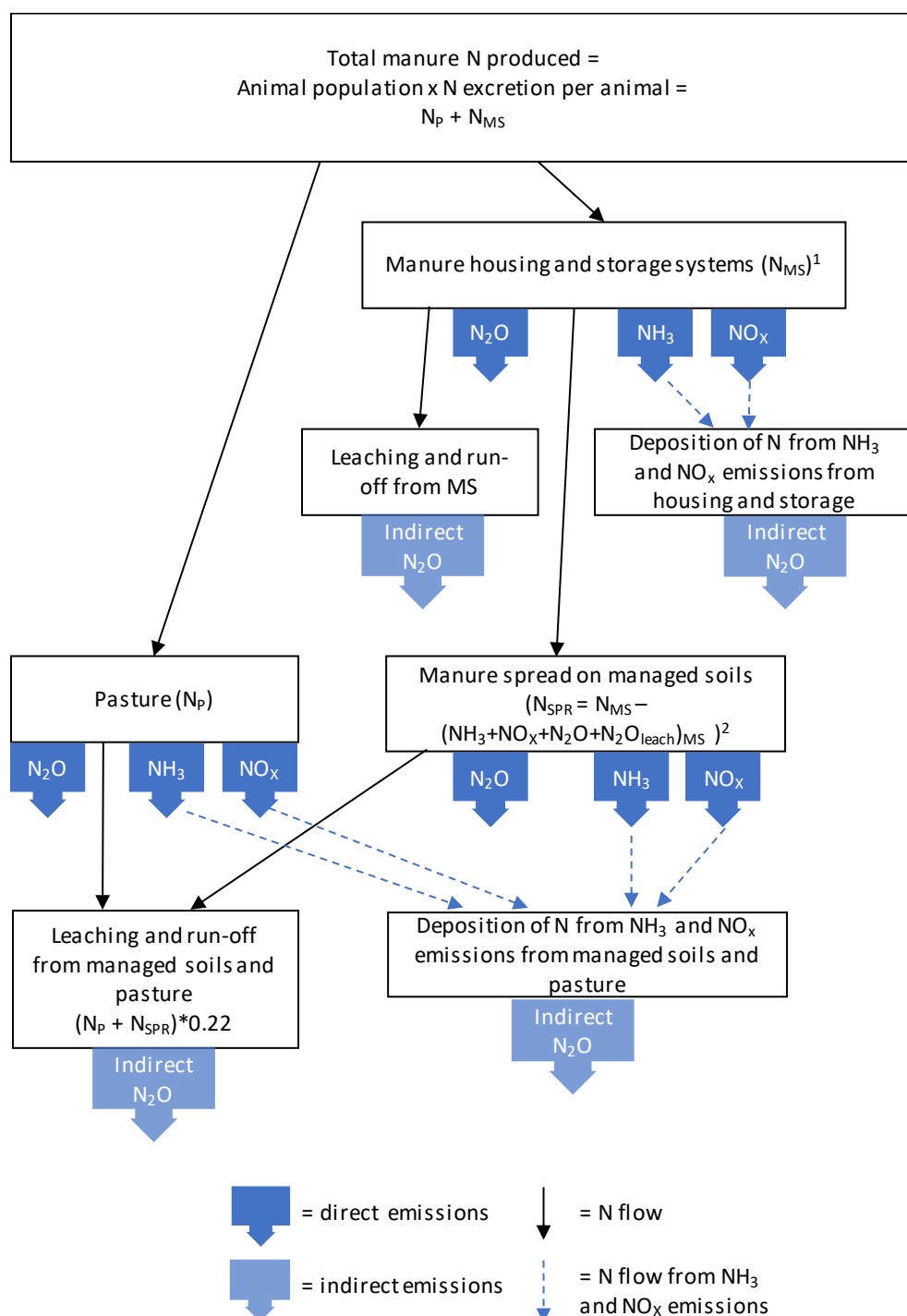
1. Manure management systems (N_2O , NO_x and NH_3)
2. Application of manure on soil (N_2O , NO_x and NH_3)
3. Droppings from animals on pastures (N_2O , NO_x and NH_3)
4. Leakage of nitrogen through manure management systems and soils (N_2O)
5. Deposition of nitrogen from emissions of NH_3 and NO_x (N_2O)

The nitrogen flow is continuously dependent on its surroundings (soil characteristics, temperature, moisture etc.) and the preceding supplies and losses of N. The Norwegian model for calculating agricultural nitrogen emissions to atmosphere is described in Carbon Limits (2018). The model is designed based on the EMEP/EEA 2016 Tier 2 technology-specific approach, which uses a mass-flow approach based on the flow of both total ammoniacal nitrogen (TAN) and total nitrogen through the manure management system until application to land (or deposition during grazing). The emission estimates of each of the sources therefore take into account emissions and losses of N species from the preceding sources. Figure 5. gives an overview of the manure nitrogen flows in the Norwegian greenhouse gas inventory.

The following decides the amounts of N that are used as the basis for the respective emission calculations:

- The amount of N in manure systems is calculated as total N in manure adjusted for the N that is dropped on pastures.
- N_2O emitted during spreading is calculated from the amounts of N in manure spread to land. This means that N lost through leaching in manure storage and as N_2O , NH_3 and NO_x in manure housing and storage is deducted. However, N lost as N_2O , NH_3 and NO_x during spreading, as well as indirect emissions of N_2O due to atmospheric deposition, are not deducted.
- NH_3 emitted during and after spreading of manure is based on the amounts of TAN in manure spread to land minus N lost through leaching in manure storage and as N_2O , NH_3 and NO_x in manure housing and storage. NO_x emitted during and after spreading of manure also has the same basis. For NH_3 emissions, N lost as N_2O and NO_x during spreading, as well as indirect emissions of N_2O due to atmospheric deposition is not deducted. Similarly, for NO_x emissions, N lost as N_2O and NH_3 during spreading, as well as indirect emissions of N_2O due to atmospheric deposition is not deducted.
- Emissions of N_2O , NH_3 and NO_x from pasture are calculated independently of each other, and are based on the amounts of N (or TAN for NH_3 emissions) estimated in manure dropped during grazing.
- N_2O lost through leaching due to spreading is based on total N in manure spread to land minus N lost through leaching in manure storage and as N_2O , NH_3 and NO_x in manure housing and storage. N_2O lost through leaching due to grazing is based on total N excreted on pastures. N_2O lost through leaching during storage of manure is based on the amounts of N estimated for the particular management systems that are susceptible to leaching. N lost through emissions of NH_3 from housing is not deducted.

- The nitrogen in NH_3 and NO_x volatilised during housing, storage, pasture and spreading of manure is the basis for the calculation of N_2O emissions from atmospheric deposition. How the amounts of N are estimated in the various emission estimates, is described in more details in the respective chapters below.



¹ For estimation of NH_3 and NO_x emissions from manure storage systems, emissions of NH_3 from housing are deducted from N excreted in housing. N_2O emissions (direct and indirect) are estimated directly from N excreted in housing.

² Emissions of N_2O , NH_3 and NO_x that have occurred prior to spreading of manure on managed soils (during housing and storage) are deducted before emissions of N_2O , NH_3 and NO_x from application to soils are estimated.

Figure 5.4 Overview of the manure nitrogen flows in the Norwegian greenhouse gas inventory

5.4 Emissions from manure management

NFR 3B

Last update: 04.02.20

5.4.1 Description

Manure management in Norway is a source of emissions to air of NH_3 , NO_x , NMVOC and PM.

5.4.2 NH_3 emissions from manure management

5.4.2.1 Description

The dominating pollutant emitted from manure management is NH_3 (NFR 3B). Emissions from cattle are most important in Norway. Emissions of NH_3 from manure depend on several factors, e.g. type of animal, nitrogen content in fodder, manure management, climate, time of spreading of manure, cultivation practices and characteristics of the soil.

5.4.2.2 Method

In Norway, all animal excreta that are not deposited during grazing are managed as manure. The estimations are made in accordance with the IPCC tier 2 method (IPCC 2006), using Norwegian values for N in excreta from different animals according to Table 5.3. The rationale for the Norwegian values for N in excreta is given in Karlengen et al. (2012) and for beef cow in Aspeholen Åby et al. (2018). The N-excretion factors for cattle, poultry and pigs have been scientifically investigated, while the remaining categories have been given by expert judgements (Karlengen et al. 2012). Based on typical Norwegian feedstock ratios, the excretion of nitrogen (N) were calculated by subtracting N in growth and products from assimilated N and P. Comparisons have also been made with emission factors used in other Nordic countries and IPCC default factors.

The factors for cattle are based on equations using animal weight, production (milking cows), life time (young cattle) and protein content in the fodder as activity data.

The Nordic feed evaluation system (NorFor) was used to develop the nitrogen factors for cattle. Excretions of N in the manure were calculated as the difference between their intake, and the sum of what is excreted in milk, fetus and deposited in the animal itself. The procedure used for calculating the excretion of feces and N consisted of two steps:

1. Simulations in "NorFor" were conducted to gain values for the feces/manure characteristics covering a wide variation of feed characteristics (N content) and production intensities (milk yield/meat production).
2. The results from the simulations were used to develop regression equations between feces/manure characteristics and parameters related to the diet (N content) and animal characteristics (milk yield, weight, age etc.).

Calculations of N-factors based on these equations have been made back to 1990 for cattle. For poultry and pigs, N-factors have been estimated for 2011 in Karlengen et al. (2012). The factors

used until this update were estimated in 1988 (Sundstøl & Mroz 1988), and are regarded as still valid for 1990. A linear interpolation has been used for the years between 1990 and 2011. For the remaining animal categories the N in excreta are considered constant throughout the time series. The factors are shown in Table 5.3. The factors for total N are used in the estimations of N₂O emissions, and ammonium N are used in the estimations of NH₃ and NO_x emissions. Norwegian values are also used for the fraction of total excretion per animal categories for each management system (MS) and for pasture. The fractions are updated every year.

Table 5.3. N in excreta from different animals¹. 2018. kg/animal/year unless otherwise informed in footnote

| | Total N | Ammonium N |
|--|----------------|-------------------|
| Dairy cattle | 128.3 | 73.3 |
| Suckling cows | 93.0 | 52.6 |
| Replacement heifers ² | 87.5 | 48.2 |
| Heifers for slaughter ² | 61.8 | 37.4 |
| Bull for slaughter ² | 69.0 | 41.8 |
| Sows | 24.4 | 15.3 |
| Boars | 24.4 | 15.3 |
| Piglets | 1.4 | 0.9 |
| Fattening pigs ³ | 3.2 | 2.1 |
| Young pigs for breeding | 9.7 | 6.5 |
| Laying hens | 0.7 | 0.3 |
| Chickens reared for laying ³ | 0.05 | 0.02 |
| Broilers ³ | 0.03 | 0.01 |
| Turkeys for slaughter ³ | 0.5 | 0.2 |
| Ducks and geese for slaughter ³ | 0.1 | 0.03 |
| Turkeys, ducks and geese reared for laying | 2.0 | 0.8 |
| Horses | 50.0 | 25.0 |
| Dairy goats | 16.9 | 10.1 |
| Other goats | 8.5 | 5.1 |
| Sheep over 1 year old | 11.6 | 6.4 |
| Sheep under 1 year old | 7.7 | 4.3 |
| Mink | 4.3 | 1.7 |
| Foxes | 9.0 | 3.6 |
| Deer | 12.0 | 5.4 |
| Reindeer | 6.0 | 2.7 |

¹ Includes pasture.

² Factors for excreted nitrogen apply for the whole life time of animals, and nitrogen is calculated when animals are slaughtered/replaced.

³ Per animal. For these categories, life time is less than a year. This means that the number of animals bred in a year is higher than the number of stalls (pens).

Source: Karlengen et al. (2012), Aspeholen Åby et al (2018), and estimations by Statistics Norway 2018.

A model based on the stepwise approach proposed by the EEA (2016) is used for calculating the emissions of ammonia from manure management. The principle of the model is illustrated in Figure 5.2.

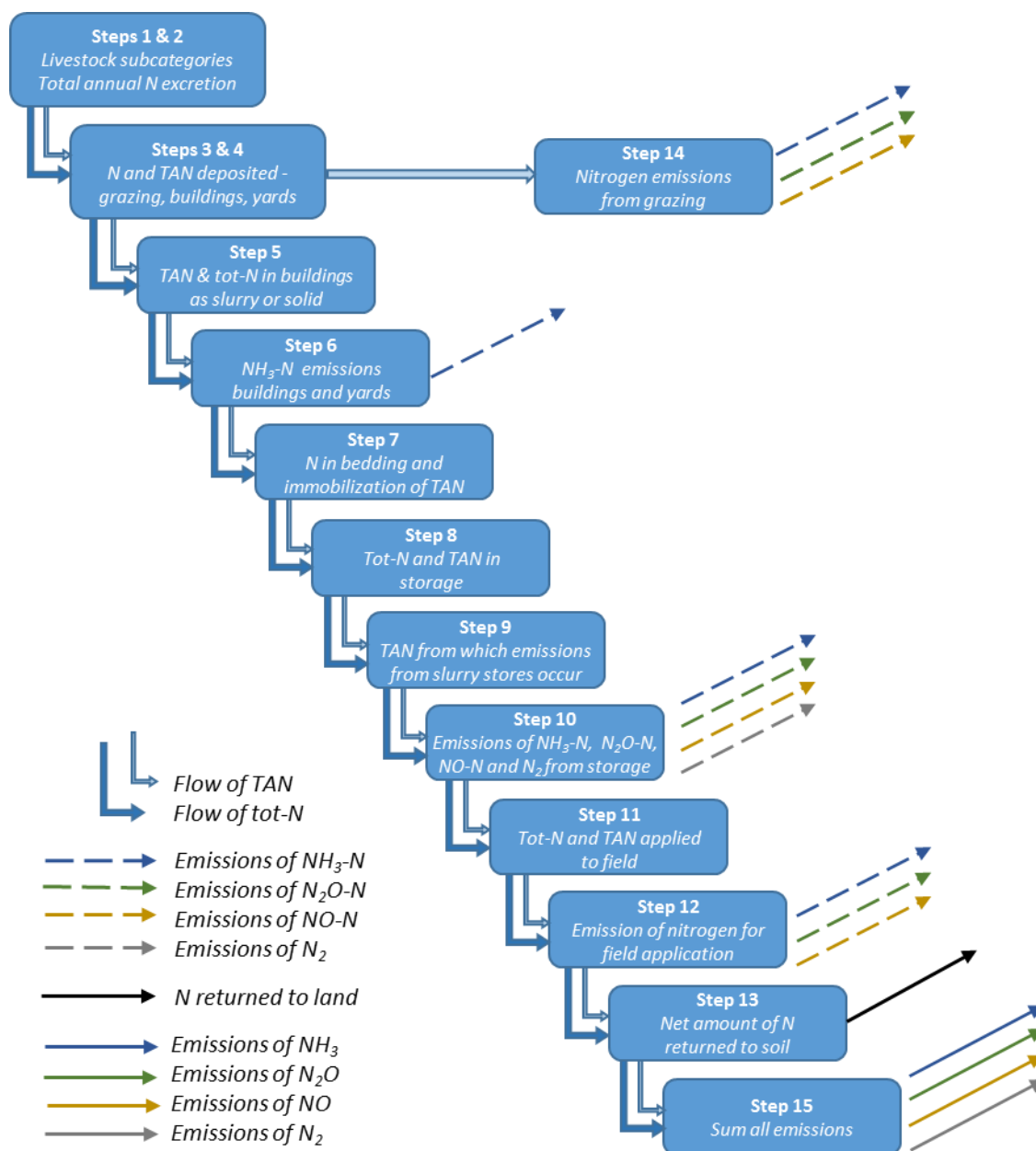


Figure 5.2 The principle of the Norwegian nitrogen model

Emissions of NH_3 are determined from buildings and yards and from manure storage systems. Total NH_3 emissions from manure management (housing and storage) are estimated by multiplying the amount of manure nitrogen (TAN) by the different emission factors for the housing and storage systems, taking into account the effect of any abatement measures and improved practices. The amount of ammonium nitrogen in the manure is estimated by the number of animals and ammonium nitrogen excretion factors for each type of animal (see Table 5.3).

5.4.2.3 Activity data

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

Surveys for assessing use of manure management systems (MMS) have been carried out in 2000, 2003, 2013 and 2018. The surveys aim to determine the fraction of manure from each animal category that is deposited in pastures during grazing, which is summarized in Table 5.4, in addition to collecting data on the MMS used for manure deposited in buildings and yards (Table 5.5).

Table 5.4 Percent of total excretion per species processed by a MMS (i.e. deposited in housing) and deposited on pasture. 2018

| | % manure to pasture | % manure to MMS |
|-------------------|------------------------|--------------------|
| Dairy cattle | 16 % | 84 % |
| Suckling cows | 31 % | 69 % |
| Young beef cattle | 31 % | 69 % |
| Swine | 0 % | 100 % |
| Laying hens | 0 % | 100 % |
| Broilers | 0 % | 100 % |
| Turkeys | 0 % | 100 % |
| Other poultry | 0 % | 100 % |
| Horses | 26 % | 74 % |
| Goats | 37 % | 63 % |
| Sheep | 62 % | 38 % |
| Fur animals | 0 % | 100 % |
| Deer | 100 % | 0 % |
| Reindeer | 100 % | 0 % |

Source: Data for storage systems from Statistics Norway (Gundersen & Heldal 2015), data for pasture times from (TINE BA Annually) (Dairy cattle, goat), Statistics Norway's Sample Survey 2001 (Statistics Norway 2002a)

Data on storage systems for other years are not available. Separate estimations of the effects on emissions of the assumed changes in storage systems since 1990 show that these assumed changes do not impact significantly. For the intermediate years 2004-2012 between the surveys of 2003 and 2013, the distribution of management system has been estimated using a linear interpolation of changes between 2003 and 2013 and between 2014 and 2018 for each system. Past surveys on management systems did include pasture for cattle, sheep. Goats and horses, but the results was not finish in time for implementation in the calculations. Currently, data for pasture times for dairy cattle and dairy goat are however annually updated in the Cow Recording System, while for the other animals, data from Sample survey of agriculture and forestry for 2001 at Statistics Norway is used. The data source for pasture times for dairy cattle in 2015 has however not been updated since 2013.

Table 5.5 Fraction of total excretion per animal category for each management system used in the estimations of NH₃ and NO_x. 2018.

| | In-house slurry pit [pit storage below animal confinements] | Tank without cover [Liquid/ slurry] | Tank with cover [Liquid/ slurry] | Heaps [solid storage] | In-house deep litter [Cattle and swine deep bedding] | Dry lot |
|--------------------------|---|---|--|-----------------------------|--|---------|
| Dairy cattle | 0.68 | 0.01 | 0.31 | 0.00 | 0.00 | 0.00 |
| Suckling cows | 0.50 | 0.01 | 0.19 | 0.12 | 0.13 | 0.04 |
| Young beef cattle | 0.66 | 0.01 | 0.24 | 0.04 | 0.04 | 0.01 |
| Swine | 0.54 | 0.31 | 0.11 | 0.02 | 0.01 | 0.01 |
| Laying hens | 0.49 | 0.00 | 0.00 | 0.51 | 0.00 | 0.00 |
| Broilers | 0.49 | 0.00 | 0.00 | 0.51 | 0.00 | 0.00 |
| Turkeys | 0.49 | 0.00 | 0.00 | 0.51 | 0.00 | 0.00 |
| Other poultry | 0.49 | 0.00 | 0.00 | 0.51 | 0.00 | 0.00 |
| Horses | 0.46 | 0.01 | 0.00 | 0.36 | 0.10 | 0.07 |
| Goats | 0.46 | 0.01 | 0.00 | 0.36 | 0.10 | 0.07 |
| Sheep | 0.70 | 0.03 | 0.01 | 0.08 | 0.14 | 0.05 |
| Fur animals ¹ | 0.00 | 0.00 | 0.00 | 1.00 | 0.00 | 0.00 |
| Deer | NA | NA | NA | NA | NA | NA |
| Reindeer | NA | NA | NA | NA | NA | NA |

¹ Based in expert opinion NIBIO

Source: Data for storage systems from Statistics Norway (2020)

In the manure surveys of 2000, 2013 and 2018, the manure of each management system is distributed by all combinations of the following productions¹⁹:

- Cattle
- Pigs
- Sheep
- Goats and horses
- Poultry

5.4.2.4 Emission factors

Emission factors vary with production and storage system; in the model there is no variation between regions for the manure management systems. The factors used are shown in Table 5.6. All emission factors in Table 5.6 are sourced from EEA (2016), since measurements of NH₃ losses in animal housing and manure storage have so far not been carried out in Norway.

¹⁹ The grouping of animals are different in the surveys. Cattle is one category in the 2000 survey and three categories in the 2013 and 2018 survey. Goats are grouped with sheep in the 2000 survey, but with horses in the 2013 and 2018 survey. Horses are grouped with other animals in the 2000 survey. Fur bearing animals are not included in the 2013 and 2018 survey, but added to the horse/goat category. All manure from fur bearing animals are considered to be stored in heaps.

Table 5.6 NH_3 emissions factors for various storage systems and productions. Per cent losses of N of ammonium N.

| | Housing | | Storage | |
|-------------------|---------|--------------|---------|-------|
| | Slurry | Solid manure | Slurry | Solid |
| Dairy cattle | 20 % | 19 % | 20 % | 27% |
| Suckling cows | 20 % | 19 % | 20 % | 27% |
| Young beef cattle | 20 % | 19 % | 20 % | 27% |
| Swine | 28 % | 27 % | 14 % | 45% |
| Laying hens | 41 % | 41 % | 14 % | 14 % |
| Broilers | 28 % | 28 % | 17 % | 17 % |
| Turkeys | 35 % | 35 % | 24 % | 24 % |
| Other poultry | 57 % | 57 % | 24 % | 24 % |
| Horses | 22 % | 22 % | 35 % | 35 % |
| Goats | 22 % | 22 % | 28 % | 28 % |
| Sheep | 22 % | 22 % | 28 % | 28 % |
| Fur animals | 27 % | 27 % | 9 % | 9 % |
| Deer | 20 % | 20 % | 20 % | 27% |
| Reindeer | 20 % | 20 % | 20 % | 27% |

The factors are combined with activity data from the Statistics Norway survey of different storage systems (as described in the previous section), and emission factors for NH_3 emissions from storage of manure and animal housing are calculated. To estimate losses, these emission factors are in turn multiplied with the amount of manure (based on number of animals and N-factors per animal, Table 5.3).

5.4.2.5 Uncertainties

Uncertainty estimates are provided in Appendix C.

5.4.2.5.1 Activity data

Emissions are estimated from the animal population. The data for the number of animals are considered to be known within ± 5 per cent.

For the emissions of NH_3 from manure management, Norwegian data for N in excreta are used. The nitrogen excretion factors are uncertain, but the range is considered to be within ± 15 per cent (Rypdal 1999). The uncertainty has not been estimated for the revised nitrogen excretion factors from Karlengen et al. (2012), and in the key category analysis the uncertainty estimate for the country specific nitrogen excretion factors from 1999 is still used as the best available estimate. This can be considered as a conservative estimate of the uncertainty since it is expected that the new nitrogen excretion factors have a lower uncertainty. The uncertainty is connected to differences in excretion between farms in different parts of the country, the fact that the survey farms may not have been representative, general measurement uncertainty and the fact that fodder and fodder practices have changed since the factors were determined.

There is also an uncertainty connected to the division between different storage systems for manure, which is considered to be within ± 10 per cent, and the division between storage and pasture, which is considered to be within ± 15 per cent.

5.4.2.5.2 *Emission factors*

Ammonia emissions from agriculture are estimated based on national conditions. There are uncertainties in several parameters as fraction of manure left on pastures, amount of manure, conditions of storage, conditions of spreading and climate conditions. All emission factors for NH_3 which have been used for both housing and storage are sourced from EMEP/EEA 2019. As stated in EMEP/EEA 2019, uncertainties with regard to NH_3 EFs vary considerably. EMEP/EEA 2019 concludes that the overall uncertainty for the United Kingdom NH_3 emissions inventory, as calculated using a Tier 3 approach, was ± 21 % (Webb and Misselbrook, 2004), while that for the Netherlands, also calculated using a Tier 3 approach, was ± 25 % (Wever et al., 2018, cited in Bruggen et al., 2018).

5.4.3 NO_x emissions from manure management

5.4.3.1 **Description**

Emissions of NO_x from manure management for the different animal groups in Norway is included in the inventory .

5.4.3.2 **Method**

In Norway, all animal excreta that are not deposited during grazing are managed as manure. Norwegian values for N in excreta from different animals according to Table 5.3 are used. How these are estimated is described in section 5.4.2.2. Norwegian values are also used for the fraction of total excretion per animal categories for each management system (MS) and for pasture. The fractions are updated every year.

NO_x volatilised from manure storage is part of the estimations of indirect N_2O emissions from atmospheric deposition.

5.4.3.3 **Activity data**

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

Surveys for assessing use of management systems have been carried out in 2000, 2003, 2013 and 2018. The distribution of manure systems used in the latest inventory is given in Table 5.5.

5.4.3.4 **Emission factors**

The emission factors used for NO_x emissions in manure management systems are shown in Table 5.7.

Table 5.7 NO_x and N₂ emission factors for manure management per manure management system.

| | kg of N in NO or N ₂ (kg TAN) ⁻¹ |
|--------------------------------|--|
| NO _x Slurry storage | 0.0001 |
| N ₂ Slurry storage | 0.0030 |
| NO _x Solid storage | 0.01 |
| N ₂ Solid storage | 0.30 |

Source: EEA (2016)

N excretions is estimated as ammonia-N (or TAN), which is the same N excretion factor that is used in the estimations of NH₃ from manure management systems.

For estimating N₂ from storage the amount is estimated as ammonia-N (TAN) and the default values for the EFs given in EMEP/EEA 2019 (Table 5.7) are used.

5.4.4 NMVOC emissions from manure management

5.4.4.1 Description

Livestock production is a source of emissions of NMVOC during feeding with silage and manure management.

5.4.4.2 Method

The emissions have been estimated using Tier 1 methodology from EEA (2016), where animal population numbers are multiplied with a default emission factor.

5.4.4.3 Activity data

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

5.4.4.4 Emission factors

Emission factors are taken from EEA (2016), see Table 5.8.

Table 5.8 Default Tier 1 EF for NMVOC. **kg AAP**(annual average population)⁻¹. **a**⁻¹

| Livestock (a) | EF, with silage feeding | EF, without silage feeding |
|---|---|----------------------------|
| | NMVOC, kg AAP ⁻¹ . a ⁻¹ | |
| Dairy cattle | 17.937 | 8.047 |
| Non-dairy cattle ¹ | 8.902 | 3.602 |
| Sheep | 0.279 | 0.169 |
| 'Swine' (Finishing pigs ²) | - | 0.551 |
| 'Swine' (Sows) | - | 1.704 |
| Goats | 0.624 | 0.542 |
| Horses | 6.028 ⁵ | 4.275 |
| Laying hens (laying hens and parents) | - | 0.165 |
| Broilers (broilers and parents) | - | 0.108 |
| Turkeys ³ | - | 0.489 |
| Other poultry (ducks, geese) ³ | - | 0.489 |
| Other animals (Fur animals) | - | 1.941 |
| Other animals (Reindeer ⁴) | - | 0.045 |

(1) Includes all other cattle

(2) Includes pigs from 8 kg to slaughtering

(3) Based on data for turkeys

(4) Assume 100% grazing

(5) EMEP/EEA factor is 7.781, but has been reduced to 6.028 which is mean value between the factors with and without silage because silage for horses normally is dryer than other silage

Source: EEA (2016) and expert judgement by Dag Austbø (see note 5)

The share of silage in fodder intake is registered for dairy cows in the Cow recording system, see Table 5.9.

Table 5.9 Silage fodder intake, dairy cows, and estimated emission factor, dairy cows.

| Year | Silage fodder, average share of daily intake | Estimated NMVOC, kg AAP ⁻¹ . a ⁻¹ |
|------|--|---|
| 1990 | 38.1 | 11.8 |
| 1995 | 39.6 | 12.0 |
| 2000 | 41.6 | 12.2 |
| 2005 | 43.6 | 12.4 |
| 2008 | 44.4 | 12.4 |
| 2009 | 44.8 | 12.5 |
| 2010 | 45.6 | 12.6 |
| 2011 | 45.3 | 12.5 |
| 2012 | 45.1 | 12.5 |
| 2013 | 45.3 | 12.5 |
| 2014 | 45.3 | 12.5 |
| 2015 | 45.3 | 12.5 |
| 2016 | 45.3 | 12.5 |
| 2017 | 45.3 | 12.5 |
| 2018 | 45.3 | 12.5 |

Source: Cow recording system (TINE BA Annually)

Amounts of silage fodder is not registered systematically for the other ruminants, and the share of silage in fodder intake is therefore based on expert judgments for these groups of animals.

Table 5.10 Silage as share of total feed intake and estimated emission factor for NMVOC emission per animal, other ruminants

| Livestock | Silage fodder, average share of daily intake | Estimated average EF, NMVOC, kg AAP ⁻¹ , a ⁻¹ |
|--|--|---|
| Growing cattle and mature non dairy cattle | 50 ¹ | 6.25 |
| Sheep | 33 ² | 0.21 |
| Goats | 40 ³ | 0.57 |
| Horses | 33 ⁴ | 4.85 |

¹ Expert judgement Harald Volden, Norwegian University of Life Sciences. November 2016

² Expert judgement Finn Avdem, Nortura. July 2016

³ Judgement by Statistics Norway. November 2016.

⁴ Expert judgement Dag Austbø, Norwegian University of Life Sciences. June 2016

Source: EEA (2016)

5.4.5 PM emissions from manure management

5.4.5.1 Method

Tier 1 methodology from EEA (2016) is used.

5.4.5.2 Activity data

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

5.4.5.3 Emission factors

Default Tier 1 emission factors are used, see Table 5.11.

Table 5.11 Default Tier 1 estimates of EF for particle emissions from animal husbandry (housing)

| Livestock | EF,TSP kg AAP ⁻¹ , a ⁻¹ | EF, PM10 kg AAP ⁻¹ , a ⁻¹ | EF, PM2.5 kg AAP ⁻¹ , a ⁻¹ | Source |
|---------------------------------------|--|--|---|------------|
| Dairy cattle | 1.38 | 0.63 | 0.41 | EEA (2016) |
| Other cattle | 0.59 | 0.27 | 0.18 | EEA (2016) |
| Finishing pigs | 1.05 | 0.14 | 0.006 | EEA (2016) |
| Weaners | 0.27 | 0.05 | 0.002 | EEA (2015) |
| Sows | 0.62 | 0.17 | 0.01 | EEA (2015) |
| Sheep | 0.14 | 0.06 | 0.02 | EEA (2016) |
| Goats | 0.14 | 0.06 | 0.02 | EEA (2016) |
| Horses | 0.48 | 0.22 | 0.14 | EEA (2016) |
| Laying hens (laying hens and parents) | 0.19 | 0.04 | 0.003 | EEA (2016) |
| Broilers (broilers and parents) | 0.04 | 0.02 | 0.002 | EEA (2016) |
| Other poultry (ducks, geese, turkeys) | 0.11 | 0.11 | 0.02 | EEA (2015) |
| Fur animals | 0.018 | 0.008 | 0.004 | EEA (2016) |

Source: EEA (2019).

5.4.6 Source specific QA/QC

Statistics Norway, in cooperation with the Norwegian University of Life Sciences (NMBU), made improvements in 2003 in the calculation model for NH_3 emissions from the agricultural sector. Data sources used for the recalculations in the revised NH_3 model are coefficients from the Norwegian University of Life Sciences, and three surveys from Statistics Norway; two manure surveys (Gundersen & Rognstad 2001) and the sample survey of agriculture and forestry 2001 (Statistics Norway 2002b).

Statistics Norway's detailed manure survey gave more extended activity data which are better related to emission source categories, for manure management and spreading. New loss factors for different manure management categories are also used in the revised NH_3 -model. These factors are closer connected to specific activities.

In 2014, a new manure survey for 2013 was carried out by Statistics Norway (Statistics Norway 2015) and in 2018 (Statistics Norway 2019). The results are implemented in the estimations of CH_4 and N_2O emissions from manure. Statistics Norway's detailed manure survey gave more extended activity data which are better related to emission source categories, for manure management and spreading. New loss factors for different manure management categories are also used in the revised NH_3 -model. These factors are closer connected to specific activities.

5.5 Crop production and agricultural soils

NFR 3D

Last update: 06.02.20

5.5.1 Description

The use of synthetic fertilizers, animal excreta nitrogen, sewage sludge and other organic fertilizers applied to soils, and droppings on pastures result in emissions of NH_3 . Agricultural activities are also a source of NO_x , NMVOC from crop plants and non-combustion emissions of particles.

5.5.2 NH_3 emissions from agricultural soils

5.5.2.1 Method

5.5.2.1.1 *NH_3 emissions from use of inorganic N-fertilizers*

NFR 3Da1

The calculations of NH_3 emissions from the use of synthetic fertilizer are based on the amounts of nitrogen supplied and emission factors for the percentage of nitrogen emitted as NH_3 during spreading, $\text{frac}_{\text{GASF}}$. The amount of nitrogen supplied is estimated based on data for total annual amount of fertilizer sold in Norway and its nitrogen content. Before 2013 these figures contained synthetic fertilizer applied in forest, so the figures were corrected for this amount in order to calculate the amount applied on agricultural fields.

5.5.2.1.2 *NH₃ emissions from animal manure applied to soils*

NFR 3Da2a

In Norway, all animal excreta that are not deposited during grazing are used as manure and applied to soils. NH₃ emissions from spreading of manure depend on several factors, e.g. climate and time of spreading of manure, type of cultivation and cultivation practices and characteristics of the soil.

Emissions of ammonia are calculated for spreading of manure on cultivated fields and meadow. The total amount of manure nitrogen that is spread is estimated by the number of animals and nitrogen excretion factors for each type of animal, and is thereafter distributed on different spreading methods based on national data. The nitrogen basis for the estimated amounts of nitrogen that volatilises as NH₃ during spreading takes into account the amount of nitrogen in the NH₃, NO_x, and N₂ that volatilises during housing and storage, as well as the N lost as N₂O and leaching during storage. Total emissions from spreading are estimated by emission factors for each different spreading method used multiplied by the amount of manure nitrogen spread with the respective method.

5.5.2.1.3 *NH₃ emissions from sewage sludge applied to soils*

NFR 3Da2b

To calculate NH₃ emissions from sewage sludge used as fertilizer, the fraction of N in manure lost as NH₃ during spreading is used ($\text{frac}_{\text{GASM}}$). The loss equals to total N in sewage sludge multiplied by $\text{frac}_{\text{GASM}}$.

5.5.2.1.4 *NH₃ emissions from other organic fertilizers applied to soils*

NFR 3Da2c

Emissions of NH₃ from other organic fertilizers applied to soils are estimated by multiplying estimated amounts of N in organic fertilizers with the $\text{frac}_{\text{GASM}}$ -factor. The annual amount of nitrogen in other organic fertilizers applied in agriculture during the period 1990-2013 was assessed in 2014 (Aquateam COWI AS 2014). Other organic fertilizers comprise three main categories; biomanure and other biological residues from biogas plants, compost from composting plants and other commercial organic fertilizer products sold. This was a practically non-existent source of nitrogen before 2000. Since then, the emissions have varied over the years, but it is a minor emission source for all years. One reason for the inter-annual variations is changes in regulations for the usage of meat and bone meal as fertilizer on agriculture land.

5.5.2.1.5 *NH₃ emissions from urine and dung deposited by grazing animals*

NFR 3Da3

Animal population data, data for pasture times, and factors for the nitrogen amount in excreta for different animal categories give the nitrogen amounts for the animal categories on pastures. The amount of animal manure dropped on pastures is given by estimations of total N in manure excreted from animals and data for pasture times (Table 5.4). It is assumed that the share of time the animals spend on pastures corresponds to the share of total N produced that is dropped during grazing. The emissions are calculated by the estimated amount of N deposited during grazing multiplied with specific emission factors by animal category (see Table 5.15).

5.5.2.2 Activity data

5.5.2.2.1 *NH₃ emissions from use of synthetic fertilizer*

The Norwegian Food Safety Authority calculates a total value for annual consumption of synthetic fertilizers in Norway based on sales figures. These data are corrected for the amount of fertilizer used in forests, which is provided by the Norwegian Institute of Bioeconomy Research.

For the calculation of the emission of NH₃ we need a specification of the use of different types of synthetic fertilizer since the NH₃ emission factor vary between different types. This is given by the Norwegian Food Safety Authority for the years from 2000. Due to lack of data for the years before 2000, we have to assume that the percentual distribution between the usage of different fertiliser types is the same as in 1994 for these years.

The calculation of emissions from use of nitrogen fertiliser is based on sales figures for each year. A strong price increase for nitrogen fertiliser caused a stock building in 2008 and corresponding lower sales in 2009. In addition, new fertilisation standards may have brought about a reduction of the use of fertilisers. To correct for this, a transfer of fertiliser use has been made from 2008 to 2009.

5.5.2.2.2 *Animal manure applied to soil and pasture*

There are several sources of activity data on spreading of manure. The main sources are the manure survey in 2000, 2013 and in 2018 by Statistics Norway and Statistics Norway et al. (2015) various sample surveys of agriculture and forestry 1990-2007 and the animal population.

Animal population is updated annually. The animal population estimation methodology is described in section 5.2. Data from the manure survey do only exist for 2000, 2013 and 2018, while the data from the sample surveys have been updated for several, but not all, years. The manner of spreading the manure also affects the NH₃ emission estimates.

Data for time on pasture and share of animals on pasture are collected from the Sample Survey in Statistics Norway 2001 and from TINE BA (TINE BA is the sales and marketing organisation for Norway's dairy cooperative and covers most of the milk production). The data from TINE BA comprises pasture data for goats and milking cows and are updated annually. All other pasture data are from the Statistics Norway Sample survey 2001. The parameters used in the calculations and their sources are shown in Table 5.12.

Table 5.12. Parameters included in the estimation of NH_3 emissions from manure

| Parameters (input) | Sources |
|---|---|
| Number of animals | Statistics Norway (applications for productions subsidies, no. and weight of approved carcasses), the Cow Recording System at TINE BA |
| Nitrogen factors for manure | Karlengen et al. (2012), Aspeholen Åby et al. (2018), various sources, compiled by Statistics Norway |
| Area where manure is spread, split on cultivated field and meadow. | Statistics Norway (Sample Surveys of Agriculture, various years), Gundersen and Rognstad (2001), Statistics Norway et al. (2015) and Statistic Norway et al. (2019) |
| Area and amount where manure is spread, split on spring and autumn. | Gundersen and Rognstad (2001), Statistics Norway et al. (2015) and Statistic Norway et al. (2019) |
| Amount of manure is spread, split on spring and autumn. | Statistics Norway (Sample Surveys of Agriculture, various years) |
| Addition of water to manure | Gundersen and Rognstad (2001) and Statistics Norway et al. (2015), expert judgements, Statistics Norway's Sample Survey 2007 |
| Spreading techniques | Gundersen and Rognstad (2001), Statistics Norway et al. (2015), expert judgements and Statistic Norway et al. (2019) |
| Usage and time of harrowing and ploughing. | Statistics Norway (Sample Surveys of Agriculture), Gundersen and Rognstad (2001), Statistics Norway et al. (2015) Statistic Norway et al. (2019) and expert judgements. |
| Pasture times for different animal categories | TINE BA (Annually) (Dairy cattle, goats), Statistics Norway's Sample Survey 2001 (Statistics Norway 2002b) (non-dairy cattle, sheep), expert judgements. |

5.5.2.3 Emission factors

5.5.2.3.1 Synthetic fertilizer

Different types of synthetic fertilizers are being used, resulting in different emissions of NH_3 . Their respective share is based on sales statistics provided annually by the Norwegian Food Safety Authority for the years from 2000. For earlier years the distribution is based on data from 1994. The NH_3 emission factors for the different types of fertilisers are shown in Table 5.13.

Table 5.13. Emission factors for $\text{NH}_3\text{-N}$ for different fertilizers. kg $\text{NH}_3\text{-N}$ per kg N

| Fertiliser | Emission factor (kg $\text{NH}_3\text{-N}$ per kg N) |
|--|--|
| Urea | 0.155 |
| Ammonium sulphate | 0.09 |
| Ammonium nitrate | 0.015 |
| Liquid ammonia | 0.019 |
| Calcium and boron calcium nitrate | 0.05 |
| Calcium ammonium nitrate and other nitrate types | 0.08 |
| NPK (Nitrogen, phosphorus, potassium) | 0.05 |
| Magnesium fertiliser | 0.05 |
| Diammonphosphate | 0.05 |
| Other NP fertiliser | 0.05 |
| NK fertiliser | 0.015 |

Source: EEA (2019)

5.5.2.3.2 Animal manure applied to soil and pasture

Emission factors for spreading of stored manure vary with spreading method (Gundersen & Rognstad 2001; Gundersen & Heldal 2015; Kolle & Oguz-Alper 2020), water contents (Statistics Norway 2007), type and time of treatment of soil (Gundersen & Rognstad 2001; Gundersen & Heldal 2015; Kolle & Oguz-Alper 2020), time of year of spreading (Gundersen & Rognstad 2001; Gundersen & Heldal 2015; Kolle & Oguz-Alper 2020; Statistics Norway 2007), cultivation and region. The basic factors used are shown in Table 5.14.

Table 5.14. Emission factors for $\text{NH}_3\text{-N}$ for various methods of spreading of manure. Per cent of ammonium N

| Meadow | | | Spring | Summer | Autumn |
|------------------------------|-------------|--------------------|--------|--------|--------|
| kg NH ₃ -N/kg TAN | | | | | |
| Spreading method | Added water | | | | |
| Broadcast spreading | < 100% | | 0.4 | 0.7 | 0.7 |
| | > 100% | | 0.24 | 0.35 | 0.35 |
| Trailing hose | < 100% | | 0.3 | 0.5 | 0.4 |
| | > 100% | | 0.18 | 0.25 | 0.2 |
| Injection | | | 0.15 | 0.30 | 0.05 |
| Dry manure | | | 0.7 | 0.9 | 0.7 |
| Arable land | | Incorporation time | Spring | Summer | Autumn |
| kg NH ₃ -N/kg TAN | | | | | |
| Spreading method | Added water | Hours | | | |
| Broadcast spreading | < 100% | 0-1 | 0.08 | 0.08 | 0.12 |
| | | 1-4 | 0.20 | 0.20 | 0.30 |
| | | 4-12 | 0.33 | 0.33 | 0.45 |
| | | 12+ | 0.50 | 0.50 | 0.45 |
| | > 100% | 0-1 | 0.04 | 0.04 | 0.06 |
| | | 1-4 | 0.10 | 0.10 | 0.15 |
| | | 4-12 | 0.17 | 0.17 | 0.28 |
| | | 12+ | 0.25 | 0.25 | 0.28 |
| Trailing hose | < 100% | 0-1 | 0.03 | 0.03 | 0.05 |
| | | 1-4 | 0.12 | 0.12 | 0.17 |
| | | 4-12 | 0.23 | 0.23 | 0.35 |
| | | 12+ | 0.50 | 0.50 | 0.45 |
| | > 100% | 0-1 | 0.02 | 0.02 | 0.02 |
| | | 1-4 | 0.06 | 0.06 | 0.09 |
| | | 4-12 | 0.12 | 0.12 | 0.22 |
| | | 12+ | 0.25 | 0.25 | 0.28 |
| Dry manure | | | 0.70 | 0.70 | 0.70 |

The factors in Table 5.14 are combined with data from the Sample survey of agriculture and forestry 2006 (Statistics Norway 2007) and a time series on mixture of water in manure. Emission factors for NH_3 emissions from spreading of manure are connected to activity data that is updated for the whole time series when new information is available, i.e. number of animals (amount of manure), time of spreading and type of cultivation of the areas where the manure is spread.

The emission factors used for the calculation of the NH_3 emissions from grazing animals are shown in Table 5.15. These are the same as the emission factors recommended in EEA (2016).

Table 5.15 Ammonia emission factors from droppings from grazing animals on pasture. Per cent of TAN.

| | NH_3 loss, % of TAN |
|-------------------|--|
| Dairy cattle | 10 % |
| Suckling cows | 10 % |
| Young beef cattle | 6 % |
| Swine | 25 % |
| Horses | 35 % |
| Goats | 9 % |
| Sheep | 9 % |
| Fur animals | 9 % |
| Deer | 10 % |
| Reindeer | 10 % |

Source: EEA (2016). For deer and reindeer EF for dairy cattle is used

5.5.3 NO_x emissions from agricultural soil

NFR 3Da1

5.5.3.1 Method

The sum of all nitrogen applied to soil has been multiplied with the default tier 1 emission factor to estimate the nitric oxide emission from crop production. Thereafter the amount of NO is translated to amount of NO_2 .

5.5.3.2 Activity data

Total N from the following sources are included:

- Synthetic fertilizers
- Animal manure spread
- Urine and dung deposited by grazing animals
- Sewage sludge
- Other organic fertilizers

5.5.3.2.1 Synthetic fertilizer

The Norwegian Food Safety Authority calculates a total value for annual consumption of synthetic fertilizers in Norway based on sales figures. These data are corrected for the amount of fertilizer used in forests, which is provided by the Norwegian Institute of Bioeconomy Research.

5.5.3.2.2 *Animal manure spread*

In Norway, all animal excreta that are not deposited during grazing are used as manure and applied to soils. The total amount of N in manure used as fertilizer is equivalent to total N excreted from the animals deducted for the amount dropped during grazing and for the amount emitted during housing and storage. How the amount of nitrogen in animal manure are calculated is described further in section 5.4.2.3.

5.5.3.2.3 *Urine and dung deposited by grazing animals*

Animal population data, data for pasture times, and factors for the nitrogen amount in excreta for different animal categories give the nitrogen amounts for the animal categories on pastures. The emissions are calculated by the estimated amount of N deposited during grazing multiplied with specific emission factors by animal category. Emissions factor used is showed in table 5.16.

5.5.3.2.4 *Sewage sludge applied to soils*

Statistics Norway (waste water statistics) annually gives values for the amount of sewage sludge, and the fraction of the sewage sludge that are applied on fields. The N-content in the sludge is given in Statistics Norway (2001), and the same value of 2.82 per cent is used for all years.

5.5.3.2.5 *Other organic fertilizers applied to soils*

How the amount of nitrogen in other organic fertilizer are estimated is described further in section 5.5.2.1.4.

5.5.3.3 Emission factors

Tier 1 default emission factor for NO_x emissions from agricultural soils has been used, see Table 5.16.

Table 5.16. Tier 1 default emission factor for NO_x emissions from agricultural soils

| Pollutant | Value | Unit |
|-----------------|-------|--|
| NO ₂ | 0.04 | kg NO ₂ kg ⁻¹ fertilizer-N applied |

Source: EEA (2016)

5.5.4 NMVOC emissions from cultivated crops

NFR 3De

5.5.4.1 Method

The tier 1 methodology has been used, multiplying cultivated area in Norway with the default emission factor from EEA.

5.5.4.2 Activity data

The activity data used are fully cultivated area given by Statistics Norway.

5.5.4.3 Emission factors

The recommend average emission factor of 0.86 kg NMVOC per ha from EEA (2016) is used. There are great variations in NMVOC emissions, dependent on crop, temperatures, yield etc. The average factor is based on a 50-50 distribution between grass and cropland. In Norway, about two thirds of the agricultural land is grassland. This may indicate an underestimation, but

lower average temperatures compared to the average for the whole EMEP area has the opposite effect.

5.5.5 Particle emissions from farm-level agricultural operations

NFR 3Dc

Agriculture is responsible for various types of non-combustion emissions of particles. This is for example dust from crops that are harvested, soil dust from work with agricultural machines, wood particles from felling of trees etc.

5.5.5.1 Method

The tier 1 methodology described in EEA (2016) is used. The area of crop land in Norway is multiplied with tier 1 emission factors, which gives emissions per area unit.

5.5.5.2 Activity data

The area of crop land given by Statistics Norway (open fields and gardens) is used since these emissions are mainly from combine harvesting and soil cultivation. The emissions may therefore be slightly overestimated since parts of the cropland is not plowed or harrowed every year.

5.5.5.3 Emission factors

Table 5.17. Tier 1 emission factors for emissions of particles from farm-level agricultural operations. kg/ha

| Pollutant | Value (kg ha ⁻¹) |
|-----------|------------------------------|
| PM10 | 1.56 |
| PM2.5 | 0.06 |
| TSP | 1.56 |

Source: EEA (2016)

5.5.6 Use of pesticides

NFR 3Df

Hexachlorobenzene (HCB) was earlier used as a pesticide, but is now forbidden. The use of this substance is not known in products in Norway today, but it can arise unintentionally and constitute a contamination in some products, among them pesticides. Pesticides can contain among other things pentachlorophenol, atrazine, simazine, picloram, pentachloronitrobenzene (PCNB, quintozone), chlorothalonil, endosulfan and chlopyralid (SYKE 2013). Emissions from the use of pesticides that can include a contamination of HCB are part of the emission inventory estimations. Information about the concentration of HCB in some of the above mentioned pesticides are shown in

Table 5.18. This information is collected from Finland (SYKE 2013) and in the estimations it is supposed that half of the HCB remnants in the pesticides are emitted to air.

5.5.6.1 Activity data

The amounts sold of the substances that can contain contaminants of HCB have been given by the Norwegian Food Safety Authority and the Product Register in the Norwegian Environment Agency. The amount of the effective substance sold in Norway have been used as activity data for the period 1996 to 2008. Since 2008, no substances containing HCB have been sold in Norway. For the years 1990-1995 the value for 1996 is used due to lack of data.

5.5.6.2 Emission factors

Table 5.18 HCB-contamination in pesticides. mg/kg

| Pesticide | |
|---------------|-----|
| Clorothalonil | 10 |
| Clopyralid | 2.5 |
| Endosulfan | 0.1 |
| Simazine | 1 |

Source: SYKE (2013)

5.5.7 Uncertainties

5.5.7.1 Activity data

There are several types of activity data entering the calculation scheme:

Sales of nitrogen fertilizer: The data are based on sales figures during one year (The Norwegian Food Safety Authority). The uncertainty in the sales figures is within ± 5 per cent (Rypdal & Zhang 2000). In addition, there is a possible additional error due to the fact that sales do not necessarily equal consumption in a particular year, due to storage.

Amount of nitrogen in manure: The figures are generated for each animal type, by multiplying the number of animals with a nitrogen excretion factor. The nitrogen excretion factors are uncertain. The range is considered to be within ± 15 per cent (Rypdal 1999). The uncertainty is connected to differences in excreted N between farms in different parts of the country, that the survey farms may not have been representative, general measurement uncertainty and the fact that fodder and feeding practices have changed since the factors were determined. This uncertainty was substantially reduced in 2013 when the nitrogen factors were assessed in a research project (Karlengen et al. 2012).

The uncertainty connected to the estimate of the amount of manure is higher than for the amount of synthetic fertiliser used.

Fate of manure: There is significant uncertainty connected to the allocation of manure between what is used as fertilizer and droppings on pastures.

Atmospheric deposition of agricultural NH_3 emissions: The data are based on national figures for NH_3 emission from agriculture. These are within ± 30 per cent (Rypdal 1999).

5.5.7.2 Emission factors

NH_3

The uncertainty in the estimate of NH_3 emissions from use of fertilizer is assessed to be about ± 20 per cent (Rypdal & Zhang 2001). The uncertainty is higher for animal manure (± 30 per cent (Rypdal & Zhang 2001)). This is due to uncertainties in several parameters (fraction of manure left on pastures, amount of manure, conditions of storage, conditions of spreading and climate conditions) (Rypdal & Zhang 2001). Other factors that could lead to uncertainty are variation in storage periods, variation in house types and climate, and variation in manure properties.

NO_x, NMVOC and PM

Default Tier 1 emission factors from EEA (2016) are used for estimation of emissions of NO_x, NMVOC and PM from crop production and agricultural soils in the Norwegian inventory. The uncertainty is given in Table 5.19.

Table 5.19. Uncertainty estimates for Tier 1 default EFs.

| Pollutant | Value (kg ha ⁻¹) | Unit | 95% confidence interval | |
|-----------------|------------------------------|--|-------------------------|-------|
| | | | Lower | Upper |
| NO ₂ | 0.04 | kg NO ₂ kg ⁻¹ fertilizer-N applied | 0.005 | 0.104 |
| NMVOC | 0.86 | kg ha ⁻¹ | 0.22 | 3.44 |
| PM10 | 1.56 | kg ha ⁻¹ | 0.78 | 7.8 |
| PM2.5 | 0.06 | kg ha ⁻¹ | 0.03 | 0.3 |
| TSP | 1.56 | kg ha ⁻¹ | 0.78 | 7.8 |

Source: EEA (2016)

5.5.8 Source specific QA/QC

Statistics Norway, in cooperation with the Norwegian University of Life Sciences (NMBU), made improvements in 2003 in the calculation model for NH₃ emissions from the agricultural sector. Data sources used for the recalculations in the revised NH₃ model are coefficients from the Norwegian University of Life Sciences, and two surveys from Statistics Norway; a manure survey (Gundersen & Rognstad 2001) and the sample survey of agriculture and forestry (2001). New factors for nitrogen excretion from animals and a revision of animal statistics has been made in 2012, to better reflect the actual nitrogen excretion from each animal category and to have a more correct linkage between the nitrogen excretion factors used and the different animal categories. Data from the manure survey of 2013 was implemented in the estimations of NH₃ emissions from manure in the 2016 submission (Statistics Norway et al. 2015).

5.6 Field burning of agricultural wastes

NFR 3F

Last update: 15.03.18

5.6.1 Description

Burning of agricultural residues gives emissions of a large range of standard combustion products. Emissions of NO_x, CO, NH₃, NMVOC, SO₂, particles and the heavy metals Pb, Cd, Hg, As, Cu and Cr, and benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4) and dioxins are included in the inventory.

5.6.2 Method

The emissions from the burning of crop residues are being calculated in accordance with a Tier 1 approach (EEA 2009):

$$E_{\text{Pollutant}} = AR_{\text{residue_burnt}} * EF_{\text{Pollutant}}$$

Where:

$E_{\text{Pollutant}}$ = emission (E) of pollutant

$AR_{\text{residue_burnt}}$ = activity rate (AR), mass of residue burnt (dry matter)

$EF_{\text{Pollutant}}$ = emission factor (EF) for pollutant

5.6.3 Activity data

The annual amount of crop residue burned on the fields is calculated based on crop production data for cereals and rapeseed from Statistics Norway, and estimates of the fraction burned made by the Norwegian Crop Research Institute and Statistics Norway. The fraction of crop residue burned on field was updated in 2012 by the Norwegian Agricultural Authorities²⁰. This reduced the fraction for 2011 from 7.5 to 4 per cent. For cereals a water content of 15 per cent is used (Statistics Norway). The activity data are consistent with the data used in the estimations of N₂O from crop residues.

5.6.4 Emission factors

Table 5.20. Emission factors for agricultural residue burning.

| Components | Emission factors | Unit | Source |
|------------------------|------------------|--|--------------------------------------|
| Precursors | | | |
| NO _x | 2.3 | kg/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| CO | 66.7 | kg/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| SO ₂ | 0.5 | kg/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| NMVO | 0.5 | kg/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| NH ₃ | 2.4 | kg /tonnes crop residue (d.m.) burned | (EEA 2016) |
| Heavy metals | | | |
| Pb | 0.11 | g/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| Hg | 0.14 | g/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| Cd | 0.88 | g/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| As | 0.0064 | g/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| Cr | 0.08 | g/tonnes crop residue (d.m.) burned | (EEA 2016) |
| Cu | 0.073 | g/ tonnes crop residue (d.m.) burned | (EEA 2013; EPA 2002) |
| Particles | | | |
| TSP | 5.8 | kg/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| PM ₁₀ | 5.7 | kg/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| PM _{2.5} | 5.4 | kg/ tonnes crop residue (d.m.) burned | (EEA 2016) |
| BC | 13 | % of PM _{2.5} | GAINS model (IIASA) |
| benzo(a)pyrene | 0,39266 | g/ tonnes crop residue (d.m.) burned | (Jenkins et al. 1996) |
| benzo(b)fluoranthene | 1,09678 | g/ tonnes crop residue (d.m.) burned | (Jenkins et al. 1996) |
| benzo(k)fluoranthene | 0,46806 | g/ tonnes crop residue (d.m.) burned | (Jenkins et al. 1996) |
| indeno(1,2,3_cd)pyrene | 0,33582 | g/ tonnes crop residue (d.m.) burned | (Jenkins et al. 1996) |
| Dioxins | 0.5 | µg I-TEQ/tonnes crop residue (d.m.) burned | (EEA 2016) |
| PCB | 2.7 | µg/tonnes crop residue (d.m.) burned | (Nielsen et al. 2015), (EEA 2013) |

²⁰ Johan Kollerud, Norwegian Agricultural Agency, unpublished material 2012.

Heavy metals and POPs

For heavy metals default emission factors from the EEA emission inventory guidebook are used (EEA (2016)). The emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4) are calculated based on emission factors from Jenkins et al. (1996).

5.6.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

5.6.6 Source specific QA/QC

In 2002, the emissions of NO_x, CO, Pb, Hg, Cd, and dioxins from agricultural residual burning were included in the Norwegian inventory, and in 2003, the emissions of particles, As, Cr and Cu were added. In 2011, also emissions of SO₂, NMVOC and NH₃ were included in the inventory. In 2016, a project to split PAH-4 emissions on individual PAHs; benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene has been performed. The time series were included but it should be noted that the figures for the earlier years have a higher uncertainty than the more recent years.

5.7 Other agricultural emission sources

NFR 3I

Last update: 15.03.18

5.7.1 Description

Straw treated with NH₃ to be utilised as fodder is a source of NH₃ emissions in Norway.

5.7.2 NH₃ emissions from treatment of straw

5.7.2.1 Method

Emissions of NH₃ from treatment of straw depend in the estimations only on the amount of NH₃ used. The amount of straw treated are influenced by the availability of grass as fodder. The total amount of NH₃ used for treatment of straw in Norway is multiplied with the share of the NH₃ that is not integrated in the straw.

5.7.2.2 Activity data

The amount of NH₃ used per year is obtained from the Budget Committee for Agriculture²¹. The area of cultivated fields is annually updated from Statistics Norway's agriculture statistics.

5.7.2.3 Emission factor

It is estimated that 54 per cent of the NH₃ applied is not integrated with the straw, and is

²¹ NILF (2017): Totalkalkylen for jordbruket.

http://www.nilf.no/statistikk/totalkalkylen/2017/BMposter/Totalkalkylen-Post220B-Halmbeh_middel_Ammoniakk

therefore emitted after the treatment (EEA 2019).

5.7.2.4 Uncertainties

Uncertainty in the estimate of emissions from NH₃ treatment of straw is rather low (± 5 per cent) (Rypdal & Zhang 2001).

5.7.2.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

6 WASTE (NFR sector 5)

NFR 5

6.1 Overview

This sector includes solid waste disposal on land (5A), other biological treatment of waste (5B), waste incineration (5C), waste water handling (5D), and other waste (5E).

Emissions from waste incineration included in sector 5C are emissions from flaring, except flaring from energy sectors (included in NFR 1 energy), and emissions from cremation and hospital waste (until 2005). The main emissions from Waste Incineration are included in the energy sector (1A) since most of incineration of municipal, industrial and medical waste in Norway is now done with energy recovery. The source sector 5E Other Waste covers emissions from municipal sewage sludge applied to parks etc., emissions from accidental car fires, building fires, and emissions from recovering processes in the waste trade.

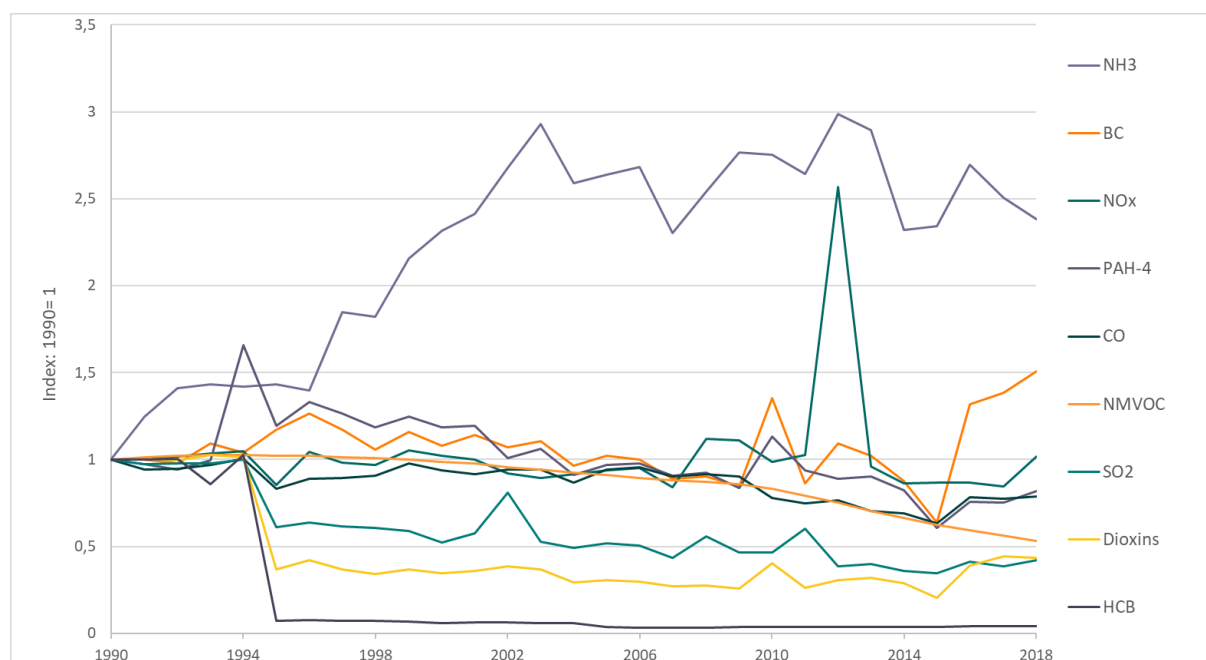


Figure 6.1. Trends for the emissions for most of long-range transboundary air pollutants from waste, relative to 1990

Source: Statistics Norway/ Norwegian Environment Agency

Figure 6.1 shows the emission trends for most of the long-range transboundary air pollutants from waste, relative to 1990. With the exception of NH₃, BC and NO_x, the emissions of all pollutants have decreased since 1990.

6.2 Solid waste disposal on land

NFR 5A

Last update: 10.01.17

6.2.1 Description

This category is mainly a source of greenhouse gas emissions. Emissions of NMVOC and particulate matter are included in this inventory. Small quantities of CO and NH₃ may be released as well but are considered insignificant and are not estimated in this inventory.

6.2.2 Method

Emissions of NMVOC and particulate matter from solid waste disposal are being calculated in accordance with a Tier 1 approach (EEA 2016) using equation:

$$E_{\text{Pollutant}} = AR_{\text{production}} * EF_{\text{Pollutant}}$$

Where:

$E_{\text{Pollutant}}$ = emission (E) of pollutant
 $AR_{\text{production}}$ = activity rate (AR), mass of landfilled waste
 $EF_{\text{Pollutant}}$ = emission factor (EF) for pollutant

Emission factors for TSP, PM₁₀ and PM_{2.5}, extracted from (EEA 2016), are shown in the Table 6.1.

Table 6.1. Emission factors for biological treatment of waste. kg/tonnes

| | TSP | PM₁₀ | PM_{2.5} |
|--|------------|------------------------|-------------------------|
| | 0.463 | 0.219 | 0.033 |

Source: EEA (2019)

NMVOC

Small quantities of NMVOC are also emitted. US Environmental Protection Agency (US EPA) evaluates that 98.7 % of the landfill gas is methane and 1.3 % are other VOCs such as perchlorethylene, pentane, butane, etc. (EEA 2016). NMVOC have therefore been estimated assuming being 1.3 % of the landfill gas. Landfill gas is estimated using IPCC methodology (IPCC 2006).

6.2.3 Activity data

The activity data are gathered from two different sources. The annual amount of waste deposited is taken from Statistics Norway's waste accounts, and the amount of methane formed by decomposition of biological waste in landfills is taken from Statistics Norway's estimation of methane at MSWD.

As the activity data consists of different sources they are not included in the NFR table. An overview of the activity data are given in Table 6.2.

Table 6.2. Activity data included in the calculation of emission in 5A. Tonnes.

| | NMVOC: CH4 from MSWD | Particulate matter: Total amount of waste deposited |
|------|-------------------------|--|
| 1990 | 82470 | 2194000 |
| 1991 | 81912 | 2194000 |
| 1992 | 80310 | 2194000 |
| 1993 | 79796 | 2194000 |
| 1994 | 79441 | 2194000 |
| 1995 | 77596 | 2194000 |
| 1996 | 75925 | 2160000 |
| 1997 | 74038 | 2210000 |
| 1998 | 68857 | 2300000 |
| 1999 | 63318 | 2255000 |
| 2000 | 65376 | 2168000 |
| 2001 | 62623 | 2037000 |
| 2002 | 59630 | 1998000 |
| 2003 | 58825 | 2002000 |
| 2004 | 58327 | 2008000 |
| 2005 | 54745 | 1996000 |
| 2006 | 55499 | 1962000 |
| 2007 | 54336 | 1998000 |
| 2008 | 51942 | 1902000 |
| 2009 | 52703 | 1375000 |
| 2010 | 51645 | 3048000 |
| 2011 | 51167 | 3033000 |
| 2012 | 49398 | 2564000 |
| 2013 | 47980 | 2931000 |
| 2014 | 46822 | 2976000 |
| 2015 | 43973 | 3310000 |
| 2016 | 41621 | 4151000 |
| 2017 | 39105 | 4731000 |
| 2018 | 38489 | 4731000 |

Source: Statistics Norway

6.3 Compost production

NFR 5B

Last update: 19.02.16

6.3.1 Description

This category covers emissions from the biological treatment of waste: composting. Emissions of NH₃ and CO from home composting and emissions of NH₃ from industrial composting are included in the inventory. This source category is not considered to be significant in Norway in

terms of long-range transboundary air pollutants. It can also be a source of NMVOC emissions which are not estimated in the Norwegian inventory.

6.3.1.1 Methodological issues

Emissions of NH₃ from composting of municipal waste have been calculated according to the Tier 2 default methodological guidance given by the 2019 Guidebook EEA (2019).

6.3.1.2 Activity data

All Norwegian waste treatment plants are obligated to statutory registration and reporting of all waste entering and leaving the plants. All waste streams are weighed, categorized with a waste type and a type of treatment. Data is available for all years since 1995 and for the year 1992. Activity data for the year 1992 and since 1995 are collected from Statistics Norway's waste statistics. For the years 1990 and 1991, activity data for 1992 are used, while AD for 1993 and 1994 are estimated by linear interpolation of activity data from 1992 and 1995.

Home composting

Emissions from home composting of garden waste and vegetable waste are also included in this inventory. The activity data for this category is available from Statistics Norway for the years 2009-2012. The amount of organic waste from households composted in the period 1990-2008 has been estimated assuming that 3 per cent of all households compost their garden and vegetable food waste (Lystad 2005). The average value of the period 2009-2012, 2.6 per cent, has been used for the following period.

6.3.1.3 Emission factors

Emissions from composting, will depend on both the composition of waste composted, amount and type of supporting material (such as wood chips and peat) used, temperature, moisture content and aeration during the process.

Table 6.2 gives default factors for CO and NH₃ emissions from biological treatment for Tier 2 method used for the estimation of Norwegian emissions.

Table 6.2. Composting emission factors. kg/tonnes

| | CO | NH ₃ |
|--------------------|------|-----------------|
| Compost production | NE | 0.24 |
| Home composting | 0.56 | 0.66 |

Source: EEA (2019)

6.4 Waste incineration

NFR 1A1a, 1A2d and 5C

Last update: 20.02.19

6.4.1 Description

In this chapter, the focus will be on waste from flaring, except flaring from energy sectors, and emissions from cremation and hospital waste until 2005. Since 2006, hospital wastes are

incinerated in incinerators for municipal wastes and are included in the energy sector.

Emissions from waste incineration are also included in this chapter, though waste incineration in district heating plants are reported under energy (NFR 1A1a), and therefore described in section 3.2.2. In 2018, there were 21 waste incineration plants where household waste was incinerated. In addition, some incineration plants burn waste other than household waste, mainly wooden waste, paper, pasteboard and cardboard. With one exception, these emissions are reported and described under energy. Waste, other than household waste, is also used as energy source in some manufacturing industries. These emissions are reported and described in the relevant subsectors under 1A2. Flaring off-shore and in refineries is included under sector 1B2c, Flaring in chemical industry is included under sector 2B5.

In Norway, the open burning of private yard waste is under different restrictions according to the respective municipality. These restrictions involve what can be burned, but also the quantity, how, when and where. In some municipalities, a complete ban is imposed. There is no registration of private waste burning and activity data on this subject are difficult to estimate. Citizens are generally encouraged to compost their yard waste or to dispose of it through one of the many waste disposal/recycling sites. Emissions from open burning of waste are considered to be insignificant and have therefore not been included in the inventory.

PCB containing material are sent abroad, mostly to Sweden, to be destroyed. There is no incineration of PCB in Norway.

6.4.2 Method

Emissions from flaring of landfill gas are estimated by multiplying the amount of gas flared with the emission factors shown in Table 6.3. Emissions from flaring of biogas from industrial waste water treatment plants are estimated. Emissions have been estimated by multiplying the amount of gas flared with the emission factors shown in Table 6.3.

A description of the method used for estimation of emissions from incineration of municipal waste is given in section 3.2.2.2.

Emissions from cremation and hospital waste are estimated by emission factors multiplied with activity data. For hospital waste, the emissions of lead, cadmium and mercury used in the model are reported to the Norwegian Environment Agency. Emissions of arsenic, chromium and copper have only been reported by two hospitals to the Norwegian Environment Agency for the year 1999. Country specific emission factors have been estimated for each component. This factor is based on the ratio between reported emission figures for 1999 and the quantities of waste burned in 1999. This factor is then multiplied with the amount of waste burned at other hospitals for the years 1995 to 2005. Around 1995, more control device systems were installed at waste incineration plants as a result of stricter emission requirements. It is assumed that this also applied to incineration of hospital waste. For the years before 1995, it is assumed that the emissions were higher. The emission standards for particulate matter from waste incineration changed from 100 to 30 mg/Nm³. It was assumed that emissions of lead, cadmium, copper and

chromium followed the same pattern as particulate matter. It is believed however, that arsenic and mercury has similar properties and it has thus been assumed that emissions of arsenic have been reduced in the same way as mercury. Emissions of mercury were regulated from 0.1 to 0.05 mg/Nm³ from 1994 to 1995. It is therefore assumed that emissions of arsenic before 1995 were twice as large as after 1995. Emissions of particulate matter were reported for all hospitals for the period 1990-1999. Since 2000, emissions from hospitals incinerators have been estimated based on EF and the amount of waste incinerated. Since 2006, all hospital waste has been incinerated at waste incineration plants.

6.4.3 Activity data

Municipal waste

Most of the waste incinerated in Norway is used for district heating and is accordingly reported under energy. The amount of waste incinerated at facilities without utilisation of energy are reported under 5C1A. Both the amount of waste and emissions are reported directly to the Norwegian Environment Agency.

Landfill gas

Information on the amount flared is given by the operators of landfills to the Norwegian Environment Agency. Emissions from landfill gas flared is included in 5C. Emissions from landfill gas used for district heating and used in other sectors are reported in the relevant subsectors under 1A1 and 1A4.

Biogas

The amount of biogas flared at some industrial waste water treatment plants are reported to the Norwegian Environment Agency for all years since 1991.

Natural gas

The amount of natural gas flared by the production of methanol is reported under 2B5.

Hospital waste

The amount of hospital waste was reported to Statistics Norway by some hospital incinerators. The hospital incinerators have gradually been closed down, mainly due to new emission limits. Since 2006, no hospital incinerators have been in operation. Nowadays, hospital waste is incinerated in incinerators for municipal waste and emissions are included under 1A1a.

Cremation

Incineration of human bodies is a common practice that is performed on an increasing part of the deceased. The number of cremated bodies is gathered by the Ministry of Culture and published in Statistics Norway's Statistical Yearbook.

6.4.4 Emission factors

Table 6.3 presents emissions factors for the waste incineration sector.

Table 6.3. Emission factors for flare of landfill gas, cremation and hospital waste incineration

| Component | Flare landfill gas and biogas kg/tonnes | Cremation Tonnes/body | Cremation Tonnes/body CS EF ¹ 2007-> | Hospital waste Tonnes/tonnes |
|----------------------------|--|---------------------------|---|---------------------------------|
| SO ₂ | 0.02 | 0.000113 | | 0.0014 |
| CO | 0.04 | 0.00014 | | 0.0028 |
| NO _x | 0.17 | 0.000825 | | 0.0014 |
| Particles PM ₁₀ | 0.14 | 0.0000347 | 3.15504E-06 | 0.0005 |
| TSP | | 0.00003856 | 3.506E-06 | 0.0005 |
| PM _{2.5} | | 0.000031 | 2.81862E-06 | 0.0005 |
| BC | 7% of PM _{2.5} | 50 % of PM _{2.5} | 50 % of PM _{2.5} | 18% of TSP |
| OC | | 36 % of PM _{2.5} | 36 % of PM _{2.5} | |
| NMVOC | 0 | 0.000013 | | 0.0007 |
| | g/tonne | kg/body | | mg/tonne |
| Pb | NA | 0.00003003 | 2.73042E-06 | Plant-specific emission factors |
| Cd | NA | 0.00000503 | 4.57344E-07 | Plant-specific emission factors |
| Hg | NA | 0.00149 | 5.59943E-05 | Reported |
| Cu | NA | 0.00001243 | 1.13018E-06 | 2594.6* |
| Cr | NA | 0.00001356 | 1.23292E-06 | 1272.4* |
| As | NA | 0.00001361 | 1.23747E-06 | 4705.6 |
| Dioxins | NA | 2.7E-11** | | 0.29685*** |
| PCB | NA | 4.1E-07 | | 0.39* |
| HCB | NA | 1.5E-07 | | 2.6* |
| benzo(a)pyrene | NA | 1.32E-08 | | 0.004179 |
| benzo(b)fluoranthene | | 7.21E-09 | | 0.035821 |
| benzo(k)fluoranthene | NA | 6.44E-09 | | .. |
| indeno(1,2,3-cd)pyrene | NA | 6.99E-09 | | .. |

NA=Not Applicable.

¹ Country specific emission factor based on measurements of Hg and TSP for the years 2013-2015. EFs for all HM are reduced as much as TSP (91 %). The new emission factors are used for all years since 2007.

* Country specific emission factor used for the years after 1995. Emission factors for the years 1990 to 1994 can be given on request.

** Emissions factor is given in kg I-TEQ/body

*** Emissions factor is given in mg I-TEQ/tonne

Source: EEA (2016), Kupiainen and Klimont (2004) and Danish IIR (Aarhus University, 2016)

BC emissions have been estimated using shares of PM_{2.5} as emission factors. Shares given by IIASA (Kupiainen & Klimont 2004) have been used. For cremation, as no share for BC was found in the literature, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

6.4.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

6.4.5.1 Activity data

No data on amounts of hospital waste have been reported since 1999. The amount of hospital waste for the subsequent years may vary from the data reported in 1998 and 1999. Since 2006, no hospital incinerators have been in operation.

6.4.5.2 Emission factors

The composition of the hospital waste could be different from the waste the emission factors are based on. In that case, the calculated emissions will be incorrect. Combustion engineering and processes also influence the emissions. These uncertainties have not been calculated.

6.4.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

6.5 Waste water handling

NFR 5D

Last update: 11.01.17

6.5.1 Description

This category covers emissions from the biological treatment of waste water and latrines. This source category is not considered to be significant on the Norwegian level in terms of long-range transboundary air pollutants.

Emissions of NMVOC are estimated in the Norwegian inventory. Emission factors for all other pollutants are not available and may be assumed negligible in most cases.

6.5.2 Method

The emissions of NMVOC from waste water treatment are being calculated in accordance with a Tier 1 approach (EEA 2016) using equation:

$$E_{\text{Pollutant}} = AR_{\text{production}} * EF_{\text{Pollutant}}$$

Where:

| | |
|--------------------------|---|
| $E_{\text{Pollutant}}$ | = emission (E) of pollutant |
| $AR_{\text{production}}$ | = activity rate (AR), amount of waste water |
| $EF_{\text{Pollutant}}$ | = emission factor (EF) for pollutant |

The emission factors for NMVOC is given in EEA (2016). The emission factors used is 15 mg NMVOC/m³ waste water.

6.5.3 Activity data

6.5.3.1 Domestic waste water

Total amount of waste water handled by all waste water treatment plants in the country is taken from Statistics Norway's municipal water supply for the years after 2009. For the years from 1990 to 2008, the amount of waste water is estimated based on the part of the population connected to treatment plants, using equation:

$$\text{Waste water} = \text{Population} \times \text{NR}_{\text{PEOPLE}} \times \text{EF}$$

Where:

$\text{NR}_{\text{PEOPLE}}$ = share of people connected to treatment plants

EF = emission factor (average household consumption per person per year)

Norwegian population data are extracted from Statistics Norway's population statistics. Data for the number of people in Norway connected to waste water treatment plants are extracted from Statistics Norway's waste water statistics. Data for the average household consumption per person per year (2002-2008) are extracted from Statistics Norway's statistics on municipal water supply. Varies between 70-76 m³ water/inhabitant/year. The number for 2001 have been used for all years 1990-2001.

6.5.3.2 Industrial waste water

The amount water released into recipient is reported by industries to the Norwegian Environment Agency (pulp and paper industry, chemical industry and food processing industries).

6.6 Other emission sources from the waste sector

NFR 5E

Last update: 18.02.16

6.6.1 Description

This category is a catch all for the waste sector. In the Norwegian inventory, emissions from sewage sludge applied on fields other than agricultural soils, accidental car fires, house fires and emissions from recovering processes in the waste trade are included in this category.

6.6.2 Method

6.6.2.1 Sewage sludge applied on fields

NH_3

Emissions of NH_3 are calculated for sewage sludge applied on fields other than agricultural soils. To calculate NH_3 emissions from sewage sludge, the fraction of N in manure lost as NH_3 is used ($\text{fra}_{\text{CGASM}}$). The loss equals to total N in sewage sludge multiplied by $\text{fra}_{\text{CGASM}}$. See section 5.5.2.1.3.

6.6.2.2 Car and house fires

Particles, heavy metals and POPs

Emissions of particles, heavy metals, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins are calculated for car fires and house fires. In addition, SO_2 , NO_x , NMVOC and CO are calculated for car fires. Emissions are estimated by multiplying the annual number of car and house fires with emission factors. Four types of buildings are separated with different emission factors: detached houses, undetached

houses, apartment buildings and industrial buildings.

6.6.2.3 Waste trade

NH₃, particles, heavy metals and POPs

Emissions from recovering processes in the waste trade include emissions of NH₃, particles, heavy metals (As, Cd, Cr, Cu, Hg, Pb), and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene). The emission figures are reported annually by the actual plants to the Norwegian Environment Agency.

6.6.3 Activity data

6.6.3.1 Sewage sludge applied on fields

Statistics Norway's waste water statistics annually gives values for the amount of sewage sludge and the fraction of the sewage sludge that is applied on fields.

6.6.3.2 Car and house fires

Data on the number of car and house fires are provided annually by the Directorate for Civil Protection and Emergency Planning. These figures only include fires reported to the fire service.

6.6.4 Emission factors

6.6.4.1 Sewage sludge applied on fields

The N-content in the sludge is given in Statistics Norway (2001), and the same value of 2.82 per cent is used for all years.

6.6.4.2 Car fires

The emission factor for particles is given by EPA (2002). EPA recommends the factor of 0.9 kg/car for combustion of wrecked cars without car tyres, and a factor for combustion of car tyres of 1.4 kg/car. This results in an overall emission factor of 2.3 kg/car. The emission factor for dioxins emissions from car fires is found in Hansen (2000). Emissions factors for heavy metal and PAHs from car fires is found in the Danish IIR (Aarhus University 2016). Emission factors for mercury from car fires is found in the French IIR (CITEPA 2016). No data are available for HCB and PCBs. NH₃ is assumed not to be emitted. It is difficult to estimate the amount of material burned in a car fire. It is assumed that the average weight of a car is 1 383 kg, average weight loss is assumed to be 18.2 per cent or 252 kg (CITEPA 2016). Emissions from vehicle fires are calculated by multiplying the mass of vehicle fires with selected emission factors. Emission factors are not available for different vehicle types, whereas it is assumed that all the different vehicle types leads to similar emissions.

6.6.4.3 House fires

It is difficult to estimate the amount of material burned in a house fire. In Finstad *et al.* (2002b) a calculation was made that has been used to scale the chosen emission factors, to reflect how much of the building that is lost in a fire. This scaling calculation is based on the amount of damage estimated in monetary value, and value on how much of the building and the furniture that is burned. The emission factors used for particles in the inventory are given by scaling the emission factors used for combustion of fuelwood in the households (Haakonsen and Kvingedal

2001). The emission factors for heavy metals are given by scaling the emission factors for combustion of wood waste in the industry (EPA 2002). For dioxins, OSPAR (Norwegian pollution control authority 2001) gives the emission factor of 170 µg I-TEQ per tonne burned material. Emissions factors for PAHs is found in Danish IIR (Aarhus University 2016). The scaled emission factors used for the different building types are given in Table 6.4.

Table 6.4. Emission factors used for car fires and house fires, unit/fire

| | Car | Detached house | Undetached house | Apartment building | Industrial building |
|-----------------------------|------------|--------------------------|--------------------------|---------------------------|----------------------------|
| SO ₂ (tonnes) | 0.0013 | NE | NE | NE | NE |
| NO _x (tonnes) | 0.0005 | NE | NE | NE | NE |
| NM VOC (tonnes) | 0.0021 | NE | NE | NE | NE |
| CO (tonnes) | 0.016 | NE | NE | NE | NE |
| TSP (tonnes) | 0.0023 | 0.14382 | 0.06162 | 0.04378 | 0.02723 |
| PM ₁₀ (tonnes) | 0.0023 | 0.14382 | 0.06162 | 0.04378 | 0.02723 |
| PM _{2.5} (tonnes) | 0.0023 | 0.14382 | 0.06162 | 0.04378 | 0.02723 |
| BC | NE | 9 % of PM _{2.5} | 9 % of PM _{2.5} | 9 % of PM _{2.5} | 9 % of PM _{2.5} |
| Pb (kg) | 0.206 | 0.00042 | 0.00018 | 0.00013 | 8E-05 |
| Cd (kg) | 0.0004 | 0.00085 | 0.00036 | 0.00026 | 0.00016 |
| Hg (kg) | 0.0001 | 0.000087 | 0.000037 | 0.000026 | 0.000016 |
| As (kg) | 6.5E-05 | 0.00135 | 0.00058 | 0.00041 | 0.00025 |
| Cr (kg) | 0.00096 | 0.00129 | 0.00055 | 0.00039 | 0.00024 |
| Cu (kg) | 0.0067 | 0.00299 | 0.00128 | 0.00091 | 0.00057 |
| benzo(a)pyrene (kg) | 0.0037 | 0.008 | 0.0064 | 0.0037 | 0.0096 |
| benzo(b)fluoranthene (kg) | 0.0041 | 0.0126 | 0.0101 | 0.0059 | 0.0152 |
| benzo(k)fluoranthene (kg) | 0.0041 | 0.0044 | 0.0036 | 0.0021 | 0.0054 |
| indeno(1,2,3-cd)pyrene (kg) | 0.0059 | 0.0086 | 0.0069 | 0.004 | 0.0104 |
| Dioxins (mg) | 0.048 | 1.43817 | 0.61621 | 0.43779 | 0.27234 |

Source: Statistics Norway, Danish IIR (Aarhus University 2016) and French IIR (CITEPA 2019)

6.6.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

6.6.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

7 Other and Natural emissions

There is no long-range transboundary air pollution reported for Norway as Other or Natural emissions.

8 Recalculations and Improvements

8.1 Recalculations

8.1.1 Overall description of the recalculations for the long-range transboundary air pollutants

As part of the continual process of improving the emission estimates, the Norwegian emission inventory has been recalculated. The process involves correcting discovered errors and utilising new or improved information where this has become available. The entire time series, at present 1990-2017, are recalculated when the method for a certain source category is revised. The figures in the inventory are therefore, as far as possible, consistent through the whole time series.

The most important recalculations in the 2020 submission are:

- Emission of NH₃ and NO_x from animal manure has been revised as a result of updated activity data from Statistics Norway 2018 survey on manure management systems.
- There are recalculations for most components in the year 1997 due to an error in the energy balance in the previous submission. Activity data from the statistics “Energy use in the manufacturing sector” for the year 1998 was wrongly used in both 1997 and 1998.
- The emission factors for particles for 2A1 Cement production, 2A2 Lime production, 2A5a Quarrying and mining of minerals other than coal, 2C3 Primary aluminium production, 2C3 Secondary aluminium production and 2H1 Pulp and paper industry has been updated according to EEA (2019) guidebook.

In combination with some minor changes from other sources, the recalculations have caused several changes in the emission figures, see Table 8.4, Table 8.5 and Table 8.6.

8.1.2 Specific description of the recalculations

8.1.2.1 Energy

There will always be some changes in the energy figures for the last year in the previous submission, e.g. some figures on energy use in manufacturing industries will be adjusted, which will lead to adjustments in other sectors, as total use of oil products in the energy accounts sum up to national sales of petroleum products. Now the final figures for energy use are available and are used in the emission calculations. Changes in emission figures caused by such revisions will not be commented on specifically under each NFR code.

1A1A Public electricity and heat production

- Revised activity data in 2017 due to updated information on waste burned for district heating. The recalculation results in a decline of all components between 1 and 4 per cent.

1A1C Manufacture of solid fuels and other energy industries

- Correction of error in SO₂ emissions for 2005 that was found in last years review.
- Recalculation of SO₂ emissions for the years 2008-2017. This resulted in an increase in SO₂ emissions for most of the years, but a decrease in 2008 and 2010.

1A2b Non-ferrous metals-1A2f Non-metallic minerals

- Reallocation of activity data between to plants in the years 2002 to 2017. The recalculations affect all components.

1A3dii National navigation and 1A4ciii Fishing

- Revised activity data in the years 2010 to 2017 due to an error in the energy balance. The recalculations affect all components except NH₃, but is most important for NO_x-emissions.

1A4ai Commercial/institutional: Stationary

- Correction of error in HG emissions for 2008 that was found in last years review.

1A5a Other, Stationary

- Correction of allocation. Activity data for stationary military was misplaced under 1A4ai in the previous submission for the years 1990 to 1995, resulting in a reallocation of emissions for these years.

1A5b Other, Mobile

- Correction of error. Distincted emission factors for NO_x, NMVOC and CO for mobile military were lost in the previous submission, resulting in an increase in the emissions from NMVOC,CO and NO_x emissions in the period 1990 – 2017.
- Reallocation of activity data. Activity data for military aviation has been split in LTO and cruise phase, resulting in decreased NO_x emissions.

IB2ai Fugitive emissions oil: Exploration, production, transport

- Recalculation of NMVOC emissions for the years 2016 and 2017.

IB2av Distribution of oil products

- Revised data from 1990 to 2017. A revised emission factor from one plant reduced the emissions of NMVOC of averagely 179 tonnes a year.

IB2C Venting and flaring (oil, gas, combined oil and gas)

- New method for estimating fugitive emissions resulted in corrections in NMVOC emissions for the years 1990 to 2017.

8.1.2.2 Industrial processes and product use

2A3 Glass production

- Correction of error. A previous erroneous figure on NO_x emissions from one plant in 2017 has been corrected and resulted in an increase of 21 tonn NO_x.

2A5A Quarrying and mining of minerals other than coal.

- Correction of error. A previous erroneous figure on TSP emissions from one plant in 1996 has been corrected and resulted in a 1232 tonn emission reduction. This correction also reduces the emissions of PM₁₀ and PM_{2,5}.

2A6 Other mineral product

- Reallocation of emissions. Emissions from production of asphalt are moved from 2A6 to 2D3B, causing a decrease in emissions of TSP, PM₁₀, PM_{2.5} and BC in 2A6.

2C2 Ferroalloys production

- Correction of error. The correction of a previous erroneous figure on SO₂ emissions from one plant in 2017, has caused a 57 tonn emission reduction.

2C7C Other metal production

- Correction of error. Previous erroneous figures on NO_x emissions has been corrected in the period 1990 to 2002, increasing the NO_x emissions with an average of 73 tonnes a year.

2D3 Solvent losses

In connection to the calculation of emissions in 2D3 there has been three reallocations:

- Emissions from production of asphalt are moved from 2A6 to 2D3B. This has caused increased emissions of TSP, PM₁₀, PM_{2.5} and BC in 2D3B.
- Reallocation of emissions. Emissions of NMVOC from creosot treated materials are moved from 2D3g to 2D3i.
- Reallocation of emissions. Emissions of NH₃ from leather preparing are moved from 2G to 2D3g in the years 1990-2017.

In addition to these reallocations there has been used revised activitydata in the years 2016 and 2017. A more detailed description of the effect on the emissions of NMVOC from the revised activity data are given in Table 8.1.

Table 8.1 Revised activity data in 2D3 in 2020 submission compared to the 2019 submission. Emission of NMVOC.

| | 2016 | | 2017 | |
|------|--------|----------|--------|----------|
| | Tonnes | per cent | tonnes | per cent |
| 2D3A | 1 896 | 13.0 | 3 568 | 24 |
| 2D3D | 2 | 0.04 | 1 147 | 20 |
| 2D3E | 0 | 0 | 0.02 | 0.00 |
| 2D3F | 0 | 0 | 12 | 15 |
| 2D3G | 0 | 0 | -14 | -12 |
| 2D3H | 0 | 0 | -15 | -6 |
| 2D3I | 2 875 | 11.4 | 2 482 | 10 |

2G Other mineral products.

- Reallocation of emissions. Emissions of NH₃ from leather preparing are moved from 2G to 2D3g in the years 1990-2017, reducing the emissions of NH₃ with a third early in the periode falling to no change in the end of the timeseries.

2H2 Food and beverages

- Correction of error. Correction of erranous activity data for food and beverages in 2017, causing a 2 kt decrease of NMVOC.

8.1.2.3 Agriculture

3B NH₃ from manure management

- Improvement: The estimations in the N-model were revised due to updated activity data from Statistics Norway 2018 survey on manure management systems. New data on both housing and storage practice for all animal types gave changes in the emissions of NH₃, and the whole time series back to 1990 has been revised. The 2018 survey also gave data on housing practice for solid and slatted floors, and open connection to storage which used to be based on expert opinion. The changes are summarized in Table 8.2.

Table 8.2 Differences between NH₃ emissions from animal manure using old and new activity data for MMS, tonne NH₃.

| MMS-data 2018 survey | | | MMS-data 2013 survey | | | Per cent change |
|----------------------|-------------------|--------------------------------|----------------------|--------------------------------|-----|-----------------|
| Year | Manure management | Animal manure applied to soils | Manure management | Animal manure applied to soils | | |
| | 3 B | 3 D a 2 a | 3 B | 3 D a 2 a | 3 B | 3 D a 2 a |
| 1990 | 10 813 | 13 426 | 10 704 | 13 483 | 1.0 | -0.4 |
| 1995 | 10 127 | 11 728 | 9 983 | 11 793 | 1.4 | -0.5 |
| 2000 | 10 473 | 11 810 | 10 301 | 11 878 | 1.7 | -0.6 |
| 2005 | 11 354 | 12 846 | 11 156 | 12 892 | 1.8 | -0.4 |
| 2012 | 11 790 | 13 158 | 11 546 | 13 150 | 2.1 | 0.1 |
| 2013 | 12 002 | 13 338 | 11 751 | 13 322 | 2.1 | 0.1 |
| 2014 | 12 012 | 13 207 | 11 769 | 13 308 | 2.1 | -0.8 |
| 2015 | 12 027 | 13 139 | 11 801 | 13 355 | 1.9 | -1.6 |
| 2016 | 12 171 | 13 196 | 11 951 | 13 537 | 1.8 | -2.5 |
| 2017 | 12 262 | 13 245 | 12 039 | 13 693 | 1.9 | -3.3 |

3Da2a NH₃ from animal manure applied to soils:

- Improvement: New activity data from 2018 manure survey gave new information about spreading practice. The whole time series back to 1990 is revised.
- The new data on grazing practice from the 2018 manure survey were not completed in time for calculation of the 2018 emission inventory. That explains why there are no recalculations in N-components for 3Da3 NH₃ from urine and dung deposited by grazing animals.

3Da2a NO_x Animal manure applied to soils:

- Correction of error: In the 2019 IIR recalculation chapter an error was reported for NO_x and that NO_x figures were calculated in tonnes of N instead of NO_x. This discovery was however not true, and the figures were originally reported correctly. The NO_x figures for animal manure applied to soils were wrongly multiplied with 46/14 in the 2019 submission. For the 2020 submission corrections are made, and this cause 69 per cent lower emissions of NO_x.

3Da3 NO_x from urine and dung deposited by grazing animals

- Improvement: In earlier reports NO_x emissions from grazing have not been reported. The N-model gives us the possibilities to calculate this, and the timeseries back to 1990 is reported in the 2020 submission. Animal population data, data for pasture times and factors for the nitrogen amount in excreta for different animal categories give the

nitrogen amounts for the animal categories on pastures. The NO_x emissions are calculated by the estimated amount of N deposited during grazing multiplied with specific emission factors by animal category.

3Da1 NH₃ and NO_x emissions from use of inorganic N-fertilizers

- Correction of error: Emissions of inorganic fertilizer are calculated based on sales figures of fertilizer. These figures use to be corrected for the amount applied in forest. Since the last report we learned that back to 2013 the sales figures for fertilizer sold in Norway was exclusive synthetic fertilizer for forest. The timeseries is recalculated back to 2013. Because of this the emissions increased by 0.7 per cent for NO_x in the period 2013-2017 and 0.6 per cent for NH₃.

3Da2b NH₃ from sewage sludge applied to soils

- Revision of factor: The timeseries for NH₃ from sewage sludge has been revised due to new fraction values. The fraction of kg NH₃-N is dependent on the values from the N-model and since the N-model was revised in 2019 the fraction changed from 0.187 to 0.156 for 2017.

8.1.2.4 Waste

5A Solid waste disposal on land

- Revised activity data. Activity data was updated for 2017 and led to an increase of emissions by 14 per cent.

5B1 Biological treatment of waste - Composting

- Revised activity data. Activity data was updated for 2017 and led to a decrease of emissions by 2 per cent CO and 27 per cent NH₃.

5C1A Municipal waste incineration

- Revised activity data. Activity data was updated for 2016 and 2017. This led to a decrease in the emissions of particulate matter by approximately 38 per cent in 2016 and 43 per cent in 2017, while there were only minor changes in the emissions of heavy metals.

5E Other waste

- Revised activity data. The activitydata are updated in the period og 1990-2017 due to changes in the N model in agriculture. The change has led to an increase in the emmissions of NH₃ in the whole timeseries. The effect of the revised activity data are given in Table 8.3.

Table 8.3 Recalculations for 5E in 2020 submission compared to the 2019 submission. Change in NH3.

| | Tonnes | Per cent |
|-------------|---------------|-----------------|
| 1990 | 23 | 23.7 |
| 1991 | 28 | 22.7 |
| 1992 | 29 | 20.2 |
| 1993 | 24 | 16.2 |
| 1994 | 21 | 14.4 |
| 1995 | 19 | 13.0 |
| 1996 | 16 | 11.3 |
| 1997 | 20 | 10.1 |
| 1998 | 15 | 7.9 |
| 1999 | 17 | 7.9 |
| 2000 | 13 | 5.8 |
| 2001 | 13 | 5.6 |
| 2002 | 12 | 4.6 |
| 2003 | 13 | 4.4 |
| 2004 | 10 | 4.1 |
| 2005 | 15 | 6.1 |
| 2006 | 18 | 7.3 |
| 2007 | 15 | 8.4 |
| 2008 | 20 | 9.4 |
| 2009 | 24 | 9.7 |
| 2010 | 21 | 8.7 |
| 2011 | 22 | 9.2 |
| 2012 | 24 | 9.0 |
| 2013 | 27 | 11.2 |
| 2014 | 18 | 9.2 |
| 2015 | 17 | 8.2 |
| 2016 | 17 | 7.1 |
| 2017 | 16 | 6.5 |

Source: Statistics Norway

8.1.3 Implications of the recalculations for long-range transboundary air pollutants

8.1.3.1 Implications for emissions levels

Table 8.4 shows the effects of recalculations on the emission figures for the main pollutants 1990-2017 and Table 8.6 the effects on the POP and heavy metal emission figures.

Table 8.4 Recalculations in 2020 submission compared to the 2019 submission. Main pollutants

| | SO2 tonnes | NOx tonnes | NMVOC tonnes | CO tonnes | NH3 tonnes |
|------|---------------|---------------|-----------------|--------------|---------------|
| 1990 | 94 | -4892 | 15927 | -47093 | 149 |
| 1991 | 69 | -4990 | 14769 | -42511 | 171 |
| 1992 | 74 | -4991 | 14700 | -41689 | 190 |
| 1993 | 75 | -4829 | 16778 | -47610 | 192 |
| 1994 | 86 | -4873 | 18190 | -50745 | 205 |
| 1995 | 55 | -5069 | 18157 | -49725 | 219 |
| 1996 | 48 | -5553 | 19742 | -53018 | 236 |
| 1997 | 114 | -5791 | 20637 | -55042 | 257 |
| 1998 | 64 | -5937 | 19802 | -51750 | 266 |
| 1999 | 80 | -6538 | 20128 | -51014 | 291 |
| 2000 | 81 | -6288 | 19611 | -49061 | 293 |
| 2001 | 63 | -6320 | 19405 | -46427 | 323 |
| 2002 | 129 | -5898 | 22003 | -44844 | 350 |
| 2003 | 129 | -6248 | 18521 | -34608 | 351 |
| 2004 | 121 | -6219 | 15389 | -21119 | 388 |
| 2005 | -413 | -6318 | 16859 | -24295 | 439 |
| 2006 | 127 | -6377 | 16860 | -20446 | 458 |
| 2007 | 117 | -6560 | 14535 | -18354 | 455 |
| 2008 | 113 | -6902 | 14826 | -15738 | 477 |
| 2009 | 99 | -7020 | 15009 | -13224 | 490 |
| 2010 | 128 | -6863 | 16825 | -12294 | 509 |
| 2011 | 103 | -7031 | 14335 | -9615 | 520 |
| 2012 | 117 | -7251 | 15592 | -7394 | 522 |
| 2013 | 266 | -7175 | 11256 | -3229 | 542 |
| 2014 | 324 | -7482 | 5245 | -3497 | -85 |
| 2015 | 342 | -7369 | 7678 | -1352 | 302 |
| 2016 | 403 | -6909 | 12098 | -816 | 194 |
| 2017 | 314 | -7256 | 13133 | -1588 | 136 |

Source: Statistics Norway

Table 8.5 Recalculations in 2020 submission compared to the 2019 submission.

Particulate matter

| | TSP tonnes | PM ₁₀ tonnes | PM _{2.5} Tonnes | BC Tonnes |
|------|---------------|----------------------------|-----------------------------|--------------|
| 1990 | 1639 | 698 | 189 | 99 |
| 1991 | 1475 | 580 | 190 | 92 |
| 1992 | 1448 | 595 | 84 | 89 |
| 1993 | 1651 | 734 | 25 | 100 |
| 1994 | 1751 | 834 | 16 | 106 |
| 1995 | 1712 | 822 | -75 | 101 |
| 1996 | 585 | 306 | -204 | 106 |
| 1997 | 1864 | 878 | -91 | 106 |
| 1998 | 1712 | 816 | -164 | 103 |
| 1999 | 1581 | 643 | -285 | 100 |
| 2000 | 1379 | 419 | -390 | 91 |
| 2001 | 1033 | 26 | -584 | 79 |
| 2002 | 576 | -441 | -1026 | 59 |
| 2003 | -233 | -1053 | -1721 | 19 |
| 2004 | -1123 | -1853 | -2175 | -20 |
| 2005 | -704 | -1496 | -1804 | -5 |
| 2006 | -1025 | -1715 | -2114 | -21 |
| 2007 | -1113 | -1815 | -2044 | -23 |
| 2008 | -1353 | -1933 | -2390 | -36 |
| 2009 | -1626 | -2099 | -2719 | -53 |
| 2010 | -1969 | -2451 | -3237 | -68 |
| 2011 | -1917 | -2349 | -2959 | -68 |
| 2012 | -2194 | -2618 | -3226 | -76 |
| 2013 | -1936 | -2309 | -2620 | -65 |
| 2014 | -1598 | -1944 | -2102 | -52 |
| 2015 | -1898 | -2237 | -2430 | -60 |
| 2016 | -1896 | -2137 | -2370 | -46 |
| 2017 | -2865 | -2998 | -3171 | -97 |

Source: Statistics Norway

Table 8.6 Recalculations in 2020 submission compared to the 2019 submission.

| | Lead | Cadmium | Mercury | Arsenic | Chromium | Copper | PAH-4 | Dioxins | PCB | HCB |
|------|------|---------|---------|---------|----------|--------|-------|---------|-----|-----|
| | Kg | Kg | Kg | Kg | Kg | Kg | Kg | Mg | Kg | Kg |
| 1990 | 436 | 60 | -7 | -158 | 0 | 0 | 0 | 3874 | 0 | 0 |
| 1991 | 393 | 54 | -6 | -142 | 0 | 0 | 0 | 3487 | 0 | 0 |
| 1992 | 385 | 53 | -7 | -139 | 0 | 0 | 0 | 3421 | 0 | 0 |
| 1993 | 440 | 60 | -8 | -160 | 0 | 0 | 0 | 3911 | 0 | 0 |
| 1994 | 469 | 64 | -8 | -170 | 0 | 0 | 0 | 4175 | 0 | 0 |
| 1995 | 458 | 63 | -8 | -166 | 0 | 0 | 0 | 4067 | 0 | 0 |
| 1996 | 487 | 67 | -8 | -177 | 0 | 0 | 0 | 4333 | 0 | 0 |
| 1997 | 521 | 71 | -6 | -176 | 20 | 8 | 0 | 4495 | 0 | 0 |
| 1998 | 483 | 66 | -8 | -175 | 0 | 0 | 0 | 4249 | 0 | 0 |
| 1999 | 493 | 67 | -9 | -180 | 0 | 0 | 0 | 4194 | 0 | 0 |
| 2000 | 508 | 70 | -9 | -185 | 0 | 0 | 0 | 4051 | 0 | 0 |
| 2001 | 530 | 73 | -9 | -193 | 0 | 0 | 0 | 3847 | 0 | 0 |
| 2002 | 604 | 83 | -11 | -219 | 0 | 0 | 0 | 3816 | 0 | 0 |
| 2003 | 572 | 81 | -11 | -217 | 2 | 2 | 0 | 3036 | 0 | 0 |
| 2004 | 559 | 76 | -10 | -203 | 0 | 0 | 0 | 2007 | 0 | 0 |
| 2005 | 562 | 77 | -10 | -204 | 0 | 0 | 0 | 2281 | 0 | 0 |
| 2006 | 559 | 77 | -10 | -202 | 0 | 0 | 0 | 1994 | 0 | 0 |
| 2007 | 546 | 75 | -10 | -199 | 0 | 0 | 0 | 1825 | 0 | 0 |
| 2008 | 547 | 75 | -39 | -198 | 0 | 0 | 0 | 1604 | 0 | 0 |
| 2009 | 560 | 77 | -10 | -204 | 0 | 0 | 0 | 1423 | 0 | 0 |
| 2010 | 693 | 87 | -10 | -225 | 154 | 120 | 0 | 1421 | 0 | 0 |
| 2011 | 597 | 75 | -9 | -199 | 102 | 78 | 0 | 1104 | 0 | 0 |
| 2012 | 629 | 81 | -10 | -208 | 114 | 90 | 0 | 1022 | 0 | 0 |
| 2013 | 531 | 66 | -7 | -163 | 142 | 115 | 0 | 624 | 0 | 0 |
| 2014 | 468 | 58 | -6 | -139 | 145 | 118 | 0 | 573 | 0 | 0 |
| 2015 | 538 | 63 | -5 | -146 | 240 | 194 | 97 | 479 | 0 | 0 |
| 2016 | 613 | 64 | -4 | -133 | 444 | 358 | 17 | 469 | 0 | 0 |
| 2017 | 520 | 59 | -6 | -154 | 191 | 142 | -33 | 111 | 0 | 0 |

Source: Statistics Norway

8.1.3.2 Implications for emission trends

As a result of the different recalculations for 1990-2017 there have been some changes in the trends. The differences are shown in the tables below.

Table 8.7 Trends in emissions 1990-2017. This submission vs. previous submission.

| | SO ₂ | NO _x | NMVOC | CO | NH ₃ |
|-----------------|-----------------|-----------------|-------|-------|-----------------|
| 2020 submission | -69.2 | -20.6 | -49.6 | -47.4 | -0.8 |
| 2019 submission | -68.9 | -15.5 | -50.2 | -50.5 | -0.9 |

Source: Statistics Norway

Table 8.8 Trends in emissions 1990-2017. This submission vs. previous submission.

| | TSP | PM ₁₀ | PM _{2.5} | BC |
|-----------------|-------|------------------|-------------------|-------|
| 2020 submission | -24.3 | -35.1 | -40.4 | -35.7 |
| 2019 submission | -18.4 | -28.4 | -32.5 | -32.2 |

Source: Statistics Norway

Table 8.9 Trends in emissions 1990-2017. This submission vs previous submission.

| | Lead | Cadmium | Mercury | Arsenic | Chromium | Copper | PAH-4 | Dioxins | PCB | HCB |
|-----------------|-------|---------|---------|---------|----------|--------|-------|---------|-------|-------|
| 2020 submission | -96.9 | -70.5 | -83.8 | -66.7 | -67.4 | 14.1 | -70.5 | -84.8 | -88.2 | -98.7 |
| 2019 submission | -97.2 | -73.1 | -83.5 | -63.8 | -69.0 | 13.5 | -70.4 | -84.4 | -88.2 | -98.7 |

Source: Statistics Norway

8.2 Planned improvements

8.2.1 Overview

There are several areas where improvement actions are needed to improve the Norwegian emission inventory system. In this chapter the main issues are listed.

8.2.2 General

- Updating of Uncertainty Analysis. The analysis will include uncertainties on particulate matter and CO
- Development of trend Key Source Analysis
- Update the methodologies used in the Norwegian inventory according to the EMEP/EEA guidebook 2019.

8.2.3 Energy

- Road transport: The HBEFA emission model for road traffic will be updated to version 4.1 when available

8.2.4 Industrial processes and product use

- 2D3 Solvent losses:
 - Provide improved description of the allocation.
 - Development of method for calculateing missing emissions of NMVOC in the periode 1990-2004 for 2D3a and 2D3h.

8.2.5 Agriculture

No improvement has been planned for the agriculture sector.

8.2.6 Waste

No improvement has been planned for the waste sector.

9 Projections

Last update: 31.05.17. Will be updated later this spring.

9.1 Introduction

This chapter describes in some detail the projections of greenhouse gas emissions and long-range transboundary air pollutants in Norway up to 2030. In line with international reporting guidelines under the Framework Convention on Climate Change these projections are based on an extension of measures and policies implemented by the beginning of 2017.²² The base year for the projections is 2013. Table 9.1 contains historic and projected emissions.

9.2 The baseline scenario

NO_x emissions increased in the 1990s mainly due to growing emissions from oil and gas production. In 2015, emissions from oil and gas production constituted 24 per cent of total Norwegian emissions, making it the major emission sector. Going forward, emissions are projected to decrease as production is estimated to decline. In shipping and road traffic, the two other main contributors to NO_x emissions, activity growth has to a larger extent than in oil and gas production, been counteracted by implementation of abatement technologies. This trend is assumed to continue and NO_x emission from road traffic is projected to decrease by almost 70 per cent by 2030. Also for SO₂ and nmVOC, new technologies has contributed substantially to cut emissions. In the projections, oil and gas production is anticipated stay stable up to 2020 and thereafter start to decline.

Table 9.1 shows projected emissions of NO_x, nmVOCs, NH₃, SO₂, PM_{2.5} and BC. The estimates are based on the same assumptions as for the greenhouse gases. Included in the NO_x projections are the effect of the first and second agreement of NO_x reduction between the Ministry of Climate and Environment and the industry for the period 2008-2017. The aim of the first and second agreement is to reduce NO_x emissions by 34 kilotons.

The EFs for PM_{2.5} and BC used in the projection for combustion of wood in household is only partly consistent with EFs used in the nation inventory for 1990-2015. The EFs in the inventory will be revised in the next inventory. The reduced emissions of PM_{2.5} and BC from 2015 to 2020 and 2030 is therefore partly due to that the EFs in the inventory is too high and partly explained by increased phase in of new technology.

²² The projections were presented in the White Paper on Long term Perspectives in Mars 2017.

Table 9.1 Anthropogenic emissions of NO_x, nmVOCs, SO₂, NH₃, PM_{2.5} and BC. Thousand tonnes

| | 1990 | 2000 | 2005 | 2015 | 2020 | 2030 |
|--------------------------------|-------|-------|-------|-------|-------|-------|
| NO _x | 199.9 | 211.6 | 205.5 | 153.2 | 142.1 | 117.5 |
| SO ₂ | 52.3 | 27.2 | 24.1 | 16.4 | 15.7 | 15.4 |
| NM VOC | 301.7 | 390.9 | 229.8 | 157.0 | 148.7 | 136.6 |
| NH ₃ | 23.7 | 25.3 | 26.7 | 26.7 | 25.5 | 26.1 |
| PM _{2.5} ² | 41.4 | 42.3 | 39.0 | 28.1 | 23.5 | 20.9 |
| BC ² | 4.7 | 4.7 | 4.4 | 3.2 | 2.9 | 2.5 |

1 The Norwegian commitment according to the Gothenburg Protocol in brackets.

2 The EF used for combustion of wood in the projections is only partly consistent with historical EFs. The EFs in the national inventory will be revised in next year's inventory.

9.3 Methodology and key assumptions

The emission projections for Norway presented in this report uses various sources and methodologies. For energy-related emissions, the projections are largely based on macroeconomic model simulations supplemented by available micro studies.

The baseline scenario is based on measures adopted and implemented by the beginning of 2017. The projections are based on information up to February 2017.

9.3.1 Macroeconomic assumptions and CO₂ emissions from the mainland economy

In the projections, current policies are assumed to be continued. Accordingly, CO₂-taxes are maintained at 2017-level in real terms. The carbon price for industries included in the emission trading scheme is assumed increase with expectations in the market in the short term and by risk free interest rate in the longer run.

In the projections, labour productivity is projected to increase by 1.5 per cent annually. Total factor productivity is on average assumed to grow annually by almost 1.0 per cent in the mainland economy. As a result, emissions per unit produced in each sector will continue to fall over time as. Moreover, for some emission sources it is assumed stronger than average growth in the development and uptake of cleaner technologies. Consequently, emissions in these sectors will be projected to fall.

The model-based projections of emissions from road transport have been adjusted, using calculations from Statistics Norway's road model. Road transport, measured by total length driven, with personnel cars is anticipated to stabilize per person, but total transport will grow with the projected growth in population. Use of bio fuels for road transportation is assumed to be about 6-7 per cent of fuel sales by 2020 and 2030. The sales of electric cars and plug-in hybrids are estimated to grow and by 2030, 50 per cent of all new passenger cars is assumed to be electric, while plug-in hybrids is anticipated to have 20 per cent of the sale. Overall, the projections imply that emissions of greenhouse gases from road transport will abate to 2030. Emissions of NO_x from road transport are projected to decrease substantially.

Production and emissions from energy intensive manufacturing are strongly correlated with consumption of electricity in this sector. In the long run, as a technical assumption, demand for electricity from energy intensive manufacturing is anticipated to stay relatively stable at the 2016-level. Production is projected to increase, while efficiency gains means that emission stay more or less stable at current levels.

Norway is the sixth largest hydro power producer in the world. Emissions from electricity production are small in Norway, as about 95 per cent (2016) of the supply of electricity originates from hydro power. In the projections, the production of renewables (e.g. hydro and wind power) is exogenously determined. The one gas fired power plant that is still running is announced to be shut down in 2018.

Table 9.2 lists key macroeconomic assumptions underpinning the Norwegian emission projections. In the baseline scenario average annual GDP growth is estimated at 1.5 per cent in 2015-2020 and at 1.7 per cent in 2020-2030. Growth in the mainland economy, i.e. total GDP excluding petroleum activities and ocean transport, is estimated at 2.0 per cent in 2015-2020 and 2.2 per cent 2020-2030.

Table 9.2 Key macroeconomic assumptions

| | 2013 | 2015 | 2020 | 2030 |
|--|------------------|-------|----------------------------|------|
| | Billion 2013 NOK | | Annual average growth rate | |
| Gross domestic product | 3 071 | 3 180 | 1.5 | 1.7 |
| - Petroleum activities and ocean transport | 652 | 683 | -0.9 | -0.8 |
| - Mainland Norway | 2 419 | 2 499 | 2.0 | 2.2 |
| Goods | 408 | 425 | 2.2 | 2.5 |
| Services | 1 075 | 1 094 | 1.7 | 2.2 |
| Consumption | 1 233 | 1 283 | 2.3 | 2.8 |
| Gross fixed capital formation | 717 | 685 | 2.1 | 2.4 |
| - Petroleum activities and ocean transport | 214 | 178 | -5.3 | 1.9 |
| - Mainland Norway | 503 | 508 | 4.2 | 2.5 |
| Population in 1000 | 5 109 | 5 214 | 1,0 | 0.8 |
| Number of persons employed in 1000 | 2 713 | 2 753 | 1,0 | 0.5 |
| | Level | | | |
| Oil price (2013-NOK) per barrel | 639 | 423 | 483 | 483 |
| Gas price (2013-NOK) per m3 | 2.31 | 1.89 | 1.75 | 1.75 |
| EU-ETS price (2013-NOK) | 35 | 67 | 57 | 85 |
| Electricity price (NOK/KWh 2013-NOK) | 0.35 | 0.28 | 0.30 | 0.33 |

Sources: Statistics Norway and Ministry of Finance.

10 Reporting on gridded emissions and LPS

Last update: 05.03.18

10.1 Gridded emissions

Information about the geographical distribution of emissions is useful for modelling and control purposes. The spatial distribution of emissions introduces another dimension (axis) to the general model.

10.1.1 EMEP grid squares

Emissions by EMEP 0.1° x 0.1° grid square are reported to the UNECE and used in models of long-range air pollution. The emissions are allocated to grid squares as follows:

- Emissions from large point sources are allocated directly to the appropriate squares
- Emissions at sea from national sea traffic are allocated to squares on the basis of a AIS-data analysis.
- The remaining emissions are allocated to squares according to the following :
 - When figures for the activity used to calculate emissions are available *directly* at geographical level, these figures are used. Examples are fuel combustion in manufacturing industries and emissions from animals.
 - When the activity at the geographical level is unknown, the national emissions are allocated *indirectly* using surrogate statistical data. For example, fuel combustion in service industries is allocated using employment figures. In a number of cases the activity is known directly at the intermediate level (county), but allocation within counties uses surrogate data.

10.1.2 Scope

Gridded emissions were last reported in 2017 for the years 1990, 1995, 2000, 2005, 2010 and 2015, at the EMEP 0.1° x 0.1° grid. Gridded emissions were reported on G-NFR sources. Gridded emissions of the following components are reported: NO_x, NMVOC, SO_x, NH₃, TSP, PM_{2.5}, PM₁₀, CO, As, Cd, Cr, Cu, Hg, Pb, Dioxins, HCB, PCB, BC, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and Indeno(1,2,3-cd)pyrene.

10.1.3 Recent improvements

In 2017, emissions have been reported for the first time using a 0.1° x 0.1° grid.

10.1.4 Planned improvements

In the next reporting, the Norwegian Environment Agency plans to reduce uncertainty in the methodologies used to allocate emissions on the grid.

10.2 LPS

In 2017, large point sources data have been reported for NO_x, NMVOC, SO_x, NH₃, TSP, PM_{2.5}, PM₁₀, CO, As, Cd, Cr, Cu, Hg, Pb, Dioxins, HCB, PCB, BC, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and Indeno(1,2,3-cd)pyrene, where emissions exceed the reporting limits included in the 2014 reporting guidelines.

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Tier 1 Key Category Analysis- Norway – 2020 submission

Methodology

The submission includes tier 1 key category analysis for the years 1990 and 2018 for the components SO₂, NO_x, NH₃, NMVOC, CO, TSP, PM₁₀, PM_{2.5}, Pb, Hg, Cd, dioxins, PAH, HCB and PCB.

The same procedure has been performed for 1990 and 2018. The emissions are analysed using the NFR19 sources (from the NFR 2019-1 reporting template) for both years. For each component the sources have been sorted according to their share of emissions, and the percentage of emissions of the component has been calculated. Sources are assigned as key until 95 per cent of total emissions are covered.

For convenience, the analysis was performed with a few exceptions from the NFR19:

- Gasoline evaporation (1 A 3 b v) is included in 1 A 3 b i-iv

These exceptions do not change the ranking of the other categories, but they may affect which categories are included at the margin.

The result tables 1-18 are sorted by share of total emissions in 2018 for each component separately. Key categories in 1990 which were not key in 2018 are placed at the bottom of each table.

When a source has become key in 2018, this may either mean that the emissions from this source have increased, or that it has decreased less than other sources. The key category analysis does not give information about the magnitude of emissions from each source, and can thus not be used to evaluate trends in emission levels for any given source.

Results

71 per cent of the 95 sources with emissions in 2018 were key to at least one component. This means that 28 sources have emissions, but are not key category for any component.

Some sources are key category to a wide range of components. This is the case in particular for public electricity and heat production (1 A 1 a), manufacture of solid fuels and other energy industries (1 A 1 c), road traffic (1 A 3 B i-iii), national navigation (1 A 3 d ii), residential plants (1 A 4 b i), ferroalloys production (2 C 2), aluminium production (2 C 3) and field burning of agricultural residue (3 F). The latter is not key to any primary gases, except CO, but key to particulates and a range of heavy metals and POPs.

Iron and steel production (2 C 1) is key category only to POPs and heavy metals.

Looking at the three most dominant sources of emissions for each component in 2018, it becomes clear that there are some sources that are responsible for a large proportion of

emissions. This is the case for emissions from manufacture of solid fuels and other energy industries, passenger cars, residential plants, ferroalloys production and aluminium production. Some distinctive characteristics of the Norwegian society can explain why some sources are dominant key categories for emissions from many components. For instance, long and cold winters lead to high demand for heating of houses, and wood-burning is common. The wood-burning leads to high emissions of CO, particulate matter, cadmium and POPs from residential plants. Due to a history of cheap electricity (hydroelectric power), Norway has a high share of energy-demanding industry. Thus, industries such as ferroalloys and aluminium production dominate the emissions for SO₂, heavy metals and PAH.

Key categories for SO₂

Production of ferroalloys was the dominant source for emissions of SO₂ in both 1990 and 2018, with, respectively, 24 and 35 per cent of the total (*Table A 1*). The importance of public electricity and heat production has grown considerably, from 2 per cent in 1990 to 13 per cent in 2018.

Several sources which were key in 1990 are no longer so in 2018. Particularly, this is the case for road traffic, due to lower sulphur content in petrol and auto diesel.

Table A 1. Key categories for SO₂ emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|----------|--|---------------|---------------|
| 2C2 | Ferroalloys production | 24.2 % | 34.6 % |
| 1A1A | Public electricity and heat production | 2.2 % | 13.0 % |
| 2C3 | Aluminium production | 8.7 % | 8.0 % |
| 2C7C | Other metal production (please specify in the IIR) | 0.9 % | 7.5 % |
| 1A3DII | National navigation (shipping) | 5.0 % | 6.1 % |
| 2B5 | Carbide production | 8.9 % | 4.9 % |
| 2A1 | Cement production | 1.2 % | 4.1 % |
| 1B2AIV | Fugitive emissions oil: Refining / storage | 7.2 % | 4.0 % |
| 1A2F | Stationary combustion in manufacturing industries and construction: Non-metallic minerals | 2.2 % | 3.6 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.9 % | 2.8 % |
| 1A4BI | Residential: Stationary | 2.4 % | 2.3 % |
| 1A1B | Petroleum refining | 0.8 % | 2.3 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 1.7 % | 1.1 % |
| 1A4AI | Commercial/institutional: Stationary | 1.6 % | 1.1 % |
| 2H1 | Pulp and Paper | 3.8 % | 0.3 % |
| 1A2E | Stationary combustion in manufacturing industries and construction: Food processing, beverages and tobacco | 3.3 % | 0.3 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 3.1 % | 0.1 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 2.7 % | 0.6 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 2.7 % | 0.7 % |
| 2B10A | Chemical industry: Other (please specify in the IIR) | 2.6 % | 0.0 % |
| 1A3BI | Road transport: Passenger cars | 2.1 % | 0.1 % |
| 1A4CIII | Agriculture/Forestry/Fishing: National fishing | 1.8 % | 0.8 % |
| 1A2GVII | Mobile Combustion in manufacturing industries and construction: (please specify in the IIR) | 1.1 % | 0.0 % |
| 2B6 | Titanium dioxide production | 1.0 % | 0.0 % |
| 1A2B | Stationary combustion in manufacturing industries and construction: Non-ferrous metals | 0.8 % | 0.0 % |
| 1A4AII | Commercial/institutional: Mobile | 0.7 % | 0.0 % |
| 1A4CII | Agriculture/Forestry/Fishing: Off-road vehicles and other machinery | 0.7 % | 0.0 % |
| 2H3 | Other industrial processes (please specify in the IIR) | 1.2 % | 0.0 % |

Key categories for NO_x

Manufacture of solid fuels and other energy industries was the most important emission source of NO_x in 2018, with 28 per cent of the emissions (*Table A 2*). In 1990, National navigation (shipping) was the dominant source, and the three road transport groups together were responsible for almost one third of the emissions. In 2018, this share was reduced to less than one fifth. The actual emissions were, however, more than halved in the period, partly due to an increased share of cars with catalysts.

Table A 2. Key categories for NO_x emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|-----------|---|---------------|---------------|
| 1A1C | Manufacture of solid fuels and other energy industries | 12.3 % | 27.6 % |
| 1A3DII | National navigation (shipping) | 17.9 % | 23.1 % |
| 1A3BI | Road transport: Passenger cars | 17.7 % | 8.8 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 12.2 % | 5.4 % |
| 2C2 | Ferroalloys production | 4.6 % | 4.4 % |
| 1A4BII | Residential: Household and gardening (mobile) | 1.5 % | 2.9 % |
| 1A4CIII | Agriculture/Forestry/Fishing: National fishing | 6.1 % | 2.7 % |
| 1A3BII | Road transport: Light duty vehicles | 2.5 % | 2.6 % |
| 3DA1 | Inorganic N-fertilizers (includes also urea application) | 2.2 % | 2.6 % |
| 1A4AII | Commercial/institutional: Mobile | 2.7 % | 2.5 % |
| 1A2GVII | Mobile Combustion in manufacturing industries and construction: (please specify in the IIR) | 4.0 % | 2.5 % |
| 3DA2A | Animal manure applied to soils | 1.0 % | 1.4 % |
| 1A1A | Public electricity and heat production | 0.7 % | 1.4 % |
| 1A2F | Stationary combustion in manufacturing industries and construction: Non-metallic minerals | 1.7 % | 1.4 % |
| 1A2A | Stationary combustion in manufacturing industries and construction: Iron and steel | 0.1 % | 0.9 % |
| 1B2AIV | Fugitive emissions oil: Refining / storage | 0.6 % | 0.8 % |
| 1A4CII | Agriculture/Forestry/Fishing: Off-road vehicles and other machinery | 1.5 % | 0.8 % |
| 1A3AII(I) | Domestic aviation LTO (civil) | 0.3 % | 0.7 % |
| 1A4BI | Residential: Stationary | 0.9 % | 0.7 % |
| 1A5B | Other, Mobile (including military, land based and recreational boats) | 0.6 % | 0.7 % |
| 3DA3 | Urine and dung deposited by grazing animals | 0.5 % | 0.6 % |
| 2B2 | Nitric acid production | 1.2 % | 0.6 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 1.1 % | 0.5 % |
| 1A1B | Petroleum refining | 0.9 % | 0.3 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 0.7 % | 0.2 % |
| 1A3C | Railways | 0.7 % | 0.4 % |

Key categories for NH₃

There has been little change in the key categories for NH₃ from 1990 to 2018 (*Table A 3*). Agricultural sources are dominant, particularly animal manure applied to soils, which was responsible for 40 per cent of the emissions in 1990 and 38 per cent in 2018. However, other agriculture and field burning of agricultural residue, which were key categories in 1990, were no longer so in 2018. There has been an opposite development for passenger cars, manure management horses and sewage sludge applied to soils – these were key categories only in 2018.

Table A 3. Key categories for NH₃ emissions, 1990 and 2018. Key categories are given in bold italic

| Source | | 1990 | 2018 |
|--------|--|---------------|---------------|
| 3DA2A | Animal manure applied to soils | 39.7 % | 37.6 % |
| 3DA1 | Inorganic N-fertilizers (includes also urea application) | 13.9 % | 12.0 % |
| 3B1B | Manure management - Non-dairy cattle | 7.6 % | 11.4 % |
| 3B1A | Manure management - Dairy cattle | 13.2 % | 10.9 % |
| 3I | Agriculture other (please specify in the IIR) | 4.3 % | 5.0 % |
| 3B3 | Manure management - Swine | 4.2 % | 4.8 % |
| 3DA3 | Urine and dung deposited by grazing animals | 3.9 % | 3.9 % |
| 3B2 | Manure management - Sheep | 3.3 % | 3.0 % |
| 3B4GI | Manure management - Laying hens | 1.3 % | 1.8 % |
| 1A3BI | Road transport: Passenger cars | 0.5 % | 1.8 % |
| 2B2 | Nitric acid production | 1.4 % | 1.5 % |
| 3B4E | Manure management - Horses | 0.7 % | 1.5 % |
| 3F | Field burning of agricultural residue | 2.9 % | 0.2 % |

Key categories for NMVOC

NMVOC emissions are spread on a wide range of sources. Offshore loading of oil is the dominant emission source (*Table A 4*), but due to increased use of emission reducing technology, this source has become less dominant during the period from 1990 to 2018. Emissions from this source amounted to 35 per cent of the total in 1990, but only 17 per cent in 2018. In 2018, solvents were the second most important source for NMVOC. Due to decreases in emissions from other sources, particularly oil loading, the share is higher than in 1990, although the actual emissions have been reduced. Passenger cars was the second largest emission source for NMVOC in 1990, but this source's share of total emissions was reduced from 16 per cent in 1990 to only 2 per cent in 2018. Emissions from domestic solvent use were not key in 1990, but have become so in 2018.

Table A 4. Key categories for NMVOC emissions, 1990 and 2018. Key categories are given in bold italic

| Source | | 1990 | 2018 |
|---------|--|---------------|---------------|
| 1B2AI | Fugitive emissions oil: Exploration, production, transport | 34.9 % | 17.2 % |
| 2D3I | Other solvent use (please specify in the IIR) | 10.0 % | 16.1 % |
| 2D3A | Domestic solvent use including fungicides | . | 10.7 % |
| 1A4BI | Residential: Stationary | 6.6 % | 10.0 % |
| 1A4BII | Residential: Household and gardening (mobile) | 8.4 % | 9.3 % |
| 1B2C | Venting and flaring (oil, gas, combined oil and gas) | 1.6 % | 5.3 % |
| 2D3D | Coating applications | 3.9 % | 4.1 % |
| 2H2 | Food and beverages industry | 0.5 % | 3.6 % |
| 1B2AIV | Fugitive emissions oil: Refining / storage | 2.8 % | 2.6 % |
| 3B1B | Manure management - Non-dairy cattle | 1.2 % | 2.4 % |
| 1A3BI | Road transport: Passenger cars | 16.4 % | 2.2 % |
| 1B2AV | Distribution of oil products | 3.0 % | 1.8 % |
| 1A3DII | National navigation (shipping) | 0.7 % | 1.7 % |
| 3B1A | Manure management - Dairy cattle | 1.2 % | 1.5 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.3 % | 1.4 % |
| 1A1A | Public electricity and heat production | 0.1 % | 1.3 % |
| 1B2B | Fugitive emissions from natural gas (exploration, production, processing, transmission, storage, distribution and other) | 0.4 % | 1.2 % |
| 1A3BIV | Road transport: Mopeds & motorcycles | 0.6 % | 0.9 % |
| 2C2 | Ferroalloys production | 0.4 % | 0.9 % |
| 3B4GII | Manure management - Broilers | 0.2 % | 0.7 % |
| 2D3E | Degreasing | 0.4 % | 0.5 % |
| 1A3BII | Road transport: Light duty vehicles | 1.5 % | 0.2 % |
| 2D3G | Chemical products | 0.8 % | 0.2 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 0.6 % | 0.2 % |
| 2D3F | Dry cleaning | 0.5 % | 0.1 % |

Key categories for CO

Aluminium production was the most important emission source for CO in 2018. This source's share grew from 13 per cent in 1990 to 35 per cent in 2018. In 1990, passenger cars was the dominant source, with 41 per cent of the total CO emissions; this share was reduced to only 6 per cent in 2018 (*Table A 5*). Emissions from combustion in households, primarily of fire wood, increased its emission share from 13 per cent in 1990 to 23 per cent in 2018, although the actual emissions were reduced. Emission reductions from some sources have caused that several minor sources have become key categories in 2018, although their actual emissions may have been reduced.

Table A 5. Key categories for CO emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|----------|---|---------------|---------------|
| 2C3 | Aluminium production | 12.6 % | 34.5 % |
| 1A4BII | Residential: Household and gardening (mobile) | 12.6 % | 23.3 % |
| 1A4BI | Residential: Stationary | 12.7 % | 21.3 % |
| 1A3BI | Road transport: Passenger cars | 41.3 % | 5.6 % |
| 2B5 | Carbide production | 4.8 % | 2.4 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.6 % | 2.3 % |
| 1A1A | Public electricity and heat production | 0.1 % | 2.2 % |
| 1A3BIV | Road transport: Mopeds & motorcycles | 0.8 % | 1.6 % |
| 1A3DII | National navigation (shipping) | 0.2 % | 0.9 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 0.7 % | 0.7 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 0.5 % | 0.7 % |
| 1A3BII | Road transport: Light duty vehicles | 5.4 % | 0.5 % |
| 3F | Field burning of agricultural residue | 3.2 % | 0.4 % |
| 2C4 | Magnesium production | 2.4 % | . |

Key categories for particulates

The dominant emission source for particulates of all sizes both in 1990 and 2018 is burning of fuel wood in small stoves in households (*Table A 6, Table A 7, Table A 8*). For PM_{2.5}, more than half of the emissions came from this source in 1990 and 2018. The importance of the different other emission sources vary some what between the different PM fractions, but other mineral products is important for TSP and PM₁₀ and ferroalloys production for both PM₁₀ and PM_{2.5}.

Table A 6. Key categories for TSP emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|----------|---|---------------|---------------|
| 1A4BI | Residential: Stationary | 33.8 % | 27.8 % |
| 2A6 | Other mineral products (please specify in the IIR) | 6.1 % | 12.5 % |
| 2D3B | Road paving with asphalt | 4.9 % | 10.1 % |
| 1A3BVI | Road transport: Automobile tyre and brake wear | 4.2 % | 8.4 % |
| 1A3BVII | Road transport: Automobile road abrasion | 10.6 % | 7.7 % |
| 2A5B | Construction and demolition | 2.8 % | 4.9 % |
| 2C2 | Ferroalloys production | 5.5 % | 3.0 % |
| 1A1A | Public electricity and heat production | 0.7 % | 2.9 % |
| 1A3DII | National navigation (shipping) | 1.7 % | 2.7 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.7 % | 2.0 % |
| 2C3 | Aluminium production | 3.5 % | 1.9 % |
| 3B4GI | Manure mangement - Laying hens | 0.8 % | 1.6 % |
| 1A4BII | Residential: Household and gardening (mobile) | 1.2 % | 1.3 % |
| 3B3 | Manure management - Swine | 0.7 % | 1.1 % |
| 3DC | Farm-level agricultural operations including storage, handling and transport of agricultural products | 1.0 % | 1.0 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 1.4 % | 0.9 % |
| 3B4GII | Manure mangement - Broilers | 0.3 % | 0.8 % |
| 3B1B | Manure management - Non-dairy cattle | 0.6 % | 0.7 % |
| 2G | Other product use (please specify in the IIR) | 0.4 % | 0.7 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 0.3 % | 0.6 % |
| 5E | Other waste | 0.3 % | 0.6 % |
| 2B2 | Nitric acid production | 1.8 % | 0.6 % |
| 3B1A | Manure management - Dairy cattle | 0.6 % | 0.6 % |
| 1B2C | Venting and flaring (oil, gas, combined oil and gas) | 1.6 % | 0.6 % |
| 2I | Wood processing | 0.3 % | 0.5 % |
| 3F | Field burning of agricultural residue | 3.3 % | 0.3 % |
| 2B5 | Carbide production | 2.0 % | 0.2 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 1.5 % | 0.3 % |
| 2H3 | Other industrial processes (please specify in the IIR) | 0.8 % | 0.0 % |
| 1A4CII | Agriculture/Forestry/Fishing: Off-road vehicles and other machinery | 0.8 % | 0.3 % |
| 1A3BII | Road transport: Light duty vehicles | 0.7 % | 0.3 % |
| 1A4CIII | Agriculture/Forestry/Fishing: National fishing | 0.6 % | 0.4 % |
| 2H1 | Pulp and Paper | 0.5 % | 0.1 % |
| 1A2GVII | Mobile Combustion in manufacturing industries and construction: (please specify in the IIR) | 0.4 % | 0.1 % |
| 1A3BI | Road transport: Passenger cars | 0.4 % | 0.5 % |

Table A 7. Key categories for PM₁₀ emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|----------|---|---------------|---------------|
| 1A4BI | Residential: Stationary | 43.9 % | 42.4 % |
| 2A6 | Other mineral products (please specify in the IIR) | 4.2 % | 9.7 % |
| 2C2 | Ferroalloys production | 7.5 % | 4.8 % |
| 1A3DII | National navigation (shipping) | 2.3 % | 4.2 % |
| 1A1A | Public electricity and heat production | 0.6 % | 4.2 % |
| 1A3BVII | Road transport: Automobile road abrasion | 4.3 % | 3.7 % |
| 2D3B | Road paving with asphalt | 1.4 % | 3.4 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.9 % | 3.1 % |
| 2C3 | Aluminium production | 3.9 % | 2.5 % |
| 2A5B | Construction and demolition | 1.1 % | 2.3 % |
| 1A4BII | Residential: Household and gardening (mobile) | 1.7 % | 2.0 % |
| 1A3BVI | Road transport: Automobile tyre and brake wear | 0.7 % | 1.7 % |
| 3DC | Farm-level agricultural operations including storage, handling and transport of agricultural products | 1.3 % | 1.5 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 1.8 % | 1.3 % |
| 2G | Other product use (please specify in the IIR) | 0.5 % | 1.0 % |
| 5E | Other waste | 0.4 % | 0.9 % |
| 1B2C | Venting and flaring (oil, gas, combined oil and gas) | 2.0 % | 0.9 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 0.3 % | 0.8 % |
| 1A3BI | Road transport: Passenger cars | 0.5 % | 0.7 % |
| 2B2 | Nitric acid production | 2.0 % | 0.7 % |
| 3B4GII | Manure management - Broilers | 0.2 % | 0.6 % |
| 1A4AI | Commercial/institutional: Stationary | 0.1 % | 0.6 % |
| 1A4CIII | Agriculture/Forestry/Fishing: National fishing | 0.8 % | 0.6 % |
| 3B1B | Manure management - Non-dairy cattle | 0.3 % | 0.5 % |
| 3B4GI | Manure management - Laying hens | 0.2 % | 0.5 % |
| 1A3BII | Road transport: Light duty vehicles | 0.9 % | 0.5 % |
| 1A4CII | Agriculture/Forestry/Fishing: Off-road vehicles and other machinery | 1.1 % | 0.4 % |
| 3F | Field burning of agricultural residue | 4.4 % | 0.4 % |
| 2B5 | Carbide production | 2.7 % | 0.3 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 2.0 % | 0.4 % |
| 2H1 | Pulp and Paper | 0.6 % | 0.1 % |
| 1A2GVII | Mobile Combustion in manufacturing industries and construction: (please specify in the IIR) | 0.6 % | 0.1 % |
| 2H3 | Other industrial processes (please specify in the IIR) | 0.5 % | 0.0 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 0.4 % | 0.1 % |
| 2A1 | Cement production | 0.4 % | 0.1 % |

Table A 8. Key categories for PM_{2.5} emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|----------|---|---------------|---------------|
| 1A4BI | Residential: Stationary | 49.8 % | 52.9 % |
| 2C2 | Ferroalloys production | 9.4 % | 6.5 % |
| 1A3DII | National navigation (shipping) | 2.8 % | 5.4 % |
| 1A1A | Public electricity and heat production | 0.4 % | 5.2 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 1.1 % | 4.1 % |
| 2C3 | Aluminium production | 3.9 % | 2.7 % |
| 1A4BII | Residential: Household and gardening (mobile) | 2.1 % | 2.7 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 2.2 % | 1.7 % |
| 1A3BVI | Road transport: Automobile tyre and brake wear | 0.6 % | 1.6 % |
| 2A6 | Other mineral products (please specify in the IIR) | 0.7 % | 1.4 % |
| 5E | Other waste | 0.5 % | 1.2 % |
| 1B2C | Venting and flaring (oil, gas, combined oil and gas) | 2.0 % | 1.2 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 0.3 % | 1.1 % |
| 1A3BI | Road transport: Passenger cars | 0.6 % | 0.9 % |
| 2G | Other product use (please specify in the IIR) | 0.6 % | 0.9 % |
| 1A3BVII | Road transport: Automobile road abrasion | 0.9 % | 0.8 % |
| 1A4CIII | Agriculture/Forestry/Fishing: National fishing | 1.0 % | 0.7 % |
| 1A4AI | Commercial/institutional: Stationary | 0.1 % | 0.7 % |
| 2B2 | Nitric acid production | 1.8 % | 0.7 % |
| 1A3BII | Road transport: Light duty vehicles | 1.1 % | 0.7 % |
| 2D3B | Road paving with asphalt | 0.2 % | 0.6 % |
| 1A4CII | Agriculture/Forestry/Fishing: Off-road vehicles and other machinery | 1.3 % | 0.6 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 2.4 % | 0.5 % |
| 3F | Field burning of agricultural residue | 5.2 % | 0.5 % |
| 2B5 | Carbide production | 3.3 % | 0.4 % |
| 1A2GVII | Mobile Combustion in manufacturing industries and construction: (please specify in the IIR) | 0.7 % | 0.2 % |
| 2H1 | Pulp and Paper | 0.6 % | 0.1 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 0.5 % | 0.2 % |

Key categories for lead (Pb)

There has been a dramatic change in dominant sources for emissions of lead from 1990 to 2018 (*Table A 9*). Due to high lead content in petrol in 1990, road traffic, particularly passenger cars, was by far the most important source. This source accounted for 77 per cent of total lead emissions. In 2018, petrol no longer contained significant amounts of lead, and other sources had become dominant. The most significant emission source in 2018 was automobile tyre and brake wear, with 28 per cent of the emissions. Domestic aviation was the second most important source in 2018, with 10 per cent of the total emissions. Due to the reduced importance of road traffic, far more sources were key in 2018 than in 1990.

Table A 9. Key categories for Pb emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|-----------|---|---------------|---------------|
| 1A3BVI | Road transport: Automobile tyre and brake wear | 0.6 % | 27.6 % |
| 1A3AII(I) | Domestic aviation LTO (civil) | 0.2 % | 10.4 % |
| 1A4BI | Residential: Stationary | 0.3 % | 8.3 % |
| 2C1 | Iron and steel production | 1.3 % | 7.0 % |
| 2C7C | Other metal production (please specify in the IIR) | 0.1 % | 6.0 % |
| 1A2F | Stationary combustion in manufacturing industries and construction: Non-metallic minerals | 0.1 % | 4.3 % |
| 2C3 | Aluminium production | 0.3 % | 3.7 % |
| 1A4AI | Commercial/institutional: Stationary | 0.0 % | 3.7 % |
| 5E | Other waste | 0.2 % | 3.6 % |
| 2B6 | Titanium dioxide production | 0.1 % | 3.6 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 0.0 % | 3.4 % |
| 2C2 | Ferroalloys production | 0.7 % | 3.1 % |
| 1A3BI | Road transport: Passenger cars | 76.8 % | 2.1 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 0.1 % | 2.1 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 0.6 % | 1.7 % |
| 2B5 | Carbide production | 0.3 % | 1.3 % |
| 1A3DII | National navigation (shipping) | 0.0 % | 1.3 % |
| 1A3AI(I) | International Aviation (LTO) | 0.0 % | 1.2 % |
| 1A1A | Public electricity and heat production | 1.1 % | 1.2 % |
| 1A4BII | Residential: Household and gardening (mobile) | 8.1 % | 0.2 % |
| 1A3BII | Road transport: Light duty vehicles | 5.8 % | 0.9 % |
| 5C1A | Municipal waste incineration | 1.1 % | 0.0 % |
| 1A3BIV | Road transport: Mopeds & motorcycles | 0.8 % | 0.0 % |
| 2A3 | Glass production | 0.8 % | 0.0 % |

Key categories for mercury (Hg)

Mercury emissions stem from a wide variety of sources (*Table A 10*). In 1990, ferroalloys production and other product use (emissions from thermometers, fluorescent tubes and other instruments) dominated, with more than half of the total emissions. In 2018, ferroalloys production had a share of 14 per cent and was the second largest contributor. The largest source in 2018 was national navigation (shipping) with a share of 17 per cent. National navigation has increased from 2 per cent in 1990.

Table A 10. Key categories for Hg emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|-----------|--|---------------|---------------|
| 1A3DII | National navigation (shipping) | 2.3 % | 17.2 % |
| 2C2 | Ferroalloys production | 34.3 % | 14.0 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.6 % | 9.8 % |
| 1A1A | Public electricity and heat production | 6.7 % | 8.5 % |
| 2A1 | Cement production | 1.8 % | 7.9 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 1.8 % | 3.7 % |
| 1A4AI | Commercial/institutional: Stationary | 0.9 % | 3.7 % |
| 1A3BI | Road transport: Passenger cars | 0.9 % | 3.4 % |
| 2G | Other product use (please specify in the IIR) | 19.5 % | 3.3 % |
| 1A2F | Stationary combustion in manufacturing industries and construction: Non-metallic minerals | 0.8 % | 2.8 % |
| 1A4CIII | Agriculture/Forestry/Fishing: National fishing | 0.8 % | 2.7 % |
| 1A4BI | Residential: Stationary | 1.3 % | 2.3 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 0.8 % | 2.0 % |
| 2C1 | Iron and steel production | 6.9 % | 1.9 % |
| 2A6 | Other mineral products (please specify in the IIR) | 0.4 % | 1.7 % |
| 1A3BVI | Road transport: Automobile tyre and brake wear | 0.2 % | 1.6 % |
| 3F | Field burning of agricultural residue | 3.9 % | 1.5 % |
| 1A3AII(I) | Domestic aviation LTO (civil) | 0.1 % | 1.3 % |
| 2B6 | Titanium dioxide production | 0.2 % | 1.3 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 1.3 % | 1.1 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 0.1 % | 0.9 % |
| 2C7C | Other metal production (please specify in the IIR) | . | 0.9 % |
| 1A3AI(I) | International Aviation (LTO) | 0.1 % | 0.9 % |
| 1A5B | Other, Mobile (including military, land based and recreational boats) | 0.2 % | 0.7 % |
| 1A2E | Stationary combustion in manufacturing industries and construction: Food processing, beverages and tobacco | 1.3 % | 0.7 % |
| 2B10A | Chemical industry: Other (please specify in the IIR) | 4.9 % | 0.1 % |
| 5C1A | Municipal waste incineration | 4.1 % | 0.0 % |
| 5C1BV | Cremation | 1.4 % | 0.5 % |

Key categories for cadmium (Cd)

Field burning of agricultural residues and ferroalloys production were the most important emission sources for cadmium in 1990, while combustion in households, particularly of fire wood, dominated in 2018 (*Table A 11*). Metal production was responsible for more than one third of the emissions in 1990, but its share was reduced significantly in 2018. More minor sources were key in 2018 than in 1990.

Table A 11. Key categories for Cd emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|----------|---|---------------|---------------|
| 1A4BI | Residential: Stationary | 9.9 % | 31.2 % |
| 1A1A | Public electricity and heat production | 6.0 % | 9.3 % |
| 2C3 | Aluminium production | 5.5 % | 8.7 % |
| 2C6 | Zinc production | 9.2 % | 6.4 % |
| 1A3BVII | Road transport: Automobile road abrasion | 3.1 % | 5.4 % |
| 3F | Field burning of agricultural residue | 21.4 % | 3.9 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 2.0 % | 3.7 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 7.3 % | 3.6 % |
| 2C2 | Ferroalloys production | 15.3 % | 3.4 % |
| 1A2C | Stationary combustion in manufacturing industries and construction: Chemicals | 0.4 % | 3.4 % |
| 1A3BI | Road transport: Passenger cars | 0.9 % | 3.3 % |
| 1A4AI | Commercial/institutional: Stationary | 0.2 % | 2.7 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.3 % | 2.3 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 0.3 % | 2.0 % |
| 1A2F | Stationary combustion in manufacturing industries and construction: Non-metallic minerals | 1.3 % | 1.9 % |
| 1A3DII | National navigation (shipping) | 0.5 % | 1.5 % |
| 2C1 | Iron and steel production | 2.8 % | 1.0 % |
| 1A3BII | Road transport: Light duty vehicles | 0.1 % | 1.0 % |
| 2G | Other product use (please specify in the IIR) | 0.1 % | 0.7 % |
| 5C1A | Municipal waste incineration | 6.0 % | 0.0 % |
| 2B5 | Carbide production | 5.3 % | 0.2 % |

Key categories for dioxins

In 1990, other industrial processes, i.e. ore mines, and magnesium production were the largest sources of dioxin emissions (*Table A 12*). The enterprises responsible for these emissions have been shut down since 1990, and thus other sources have become dominant. In 2018, residential plants were responsible for 31 per cent of the dioxin emissions in Norway. Most of these emissions came from use of fire wood. Since the major emission sources in 1990 have disappeared, several minor sources have become key in 2018.

Table A 12. Key categories for dioxin emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|----------|---|---------------|---------------|
| 1A4BI | Residential: Stationary | 8.0 % | 31.1 % |
| 5E | Other waste | 1.7 % | 15.7 % |
| 1A3DII | National navigation (shipping) | 2.1 % | 15.3 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.5 % | 7.8 % |
| 2A1 | Cement production | 0.2 % | 6.2 % |
| 1A1A | Public electricity and heat production | 10.3 % | 5.1 % |
| 2C3 | Aluminium production | 0.6 % | 4.7 % |
| 2C2 | Ferroalloys production | 1.2 % | 2.8 % |
| 1A4CIII | Agriculture/Forestry/Fishing: National fishing | 0.8 % | 2.4 % |
| 2C1 | Iron and steel production | 0.9 % | 1.8 % |
| 1A2D | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print | 0.9 % | 1.3 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 0.2 % | 1.0 % |
| 1A3BI | Road transport: Passenger cars | 1.6 % | 0.8 % |
| 2H3 | Other industrial processes (please specify in the IIR) | 41.4 % | . |
| 2C4 | Magnesium production | 24.4 % | . |
| 5C1BIII | Clinical waste incineration | 4.1 % | . |

Key categories for PAH

Aluminium production was the most important emission source of PAH emissions in both 1990 and 2018, although it was less dominant in 2018 (*Table A 13-Table A 16*). For benzo(k)fluoranthene, there is a break in the time series between 2015 and 2016, due to the introduction of a new calculation methodology. This inconsistency in the figures has caused a striking change from 1990 to 2018 for aluminium production in table A 15. A strong reduction in emissions from aluminium production from 1990 to 2018 has brought about that road transport has become key in 2018, although the emission increase in the different road transport groups not has been substantial. Combustion of fire wood in households is also an important emission source for PAH.

Table A 13. Key categories for benzo(a)pyrene emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|---------|--|---------------|---------------|
| 2C3 | Aluminium production | 68.7 % | 49.7 % |
| 1A4BI | Residential: Stationary | 18.5 % | 25.4 % |
| 1A3BI | Road transport: Passenger cars | 2.1 % | 10.5 % |
| 1A3BII | Road transport: Light duty vehicles | 0.4 % | 4.5 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 0.5 % | 2.2 % |
| 5E | Other waste | 0.7 % | 1.9 % |
| 2B5 | Carbide production | 1.5 % | 1.3 % |
| 3F | Field burning of agricultural residue | 3.8 % | 0.7 % |
| 2C7C | Other metal production (please specify in the IIR) | 2.3 % | 0.2 % |

Table A 14. Key categories for benzo(b)fluoranthene emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|---------|---|---------------|---------------|
| 2C3 | Aluminium production | 70.9 % | 55.6 % |
| 1A4BI | Residential: Stationary | 13.4 % | 15.3 % |
| 1A3BVI | Road transport: Automobile tyre and brake wear | 1.8 % | 7.7 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 1.3 % | 5.2 % |
| 1A3BI | Road transport: Passenger cars | 1.1 % | 4.6 % |
| 2B5 | Carbide production | 1.6 % | 2.7 % |
| 1A3BII | Road transport: Light duty vehicles | 0.2 % | 2.0 % |
| 1A2F | Stationary combustion in manufacturing industries and construction: Non-metallic minerals | 0.2 % | 1.2 % |
| 5E | Other waste | 0.5 % | 1.1 % |
| 3F | Field burning of agricultural residue | 4.9 % | 0.8 % |
| 2C7C | Other metal production (please specify in the IIR) | 2.3 % | 0.4 % |

Table A 15. Key categories for benzo(k)fluoranthene emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|---------|--|---------------|---------------|
| 2C3 | Aluminium production | 78.2 % | 48.5 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 2.9 % | 18.0 % |
| 1A4BI | Residential: Stationary | 6.7 % | 11.3 % |
| 1A3BI | Road transport: Passenger cars | 1.5 % | 11.1 % |
| 1A3BII | Road transport: Light duty vehicles | 0.3 % | 4.8 % |
| 5E | Other waste | 0.4 % | 1.4 % |
| 3F | Field burning of agricultural residue | 4.2 % | 1.0 % |
| 2C7C | Other metal production (please specify in the IIR) | 2.6 % | 0.3 % |
| 2B5 | Carbide production | 1.7 % | 0.9 % |

Table A 16. Key categories for indeno(1,2,3_cd)pyrene emissions, 1990 and 2018. Key categories are given in bold italic

| | Source | 1990 | 2018 |
|---------|--|---------------|---------------|
| 2C3 | Aluminium production | 58.2 % | 40.7 % |
| 1A4BI | Residential: Stationary | 22.7 % | 24.3 % |
| 1A3BI | Road transport: Passenger cars | 4.3 % | 13.9 % |
| 1A3BII | Road transport: Light duty vehicles | 0.7 % | 5.6 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 1.2 % | 4.5 % |
| 1A3DII | National navigation (shipping) | 1.3 % | 3.7 % |
| 5E | Other waste | 1.4 % | 2.9 % |
| 3F | Field burning of agricultural residue | 5.6 % | 0.8 % |
| 2C7C | Other metal production (please specify in the IIR) | 1.9 % | 0.2 % |
| 2B5 | Carbide production | 1.3 % | 1.0 % |

Key categories for HCB

In 1990, magnesium production was by far the largest source for HCB emissions, with 99 per cent of the total. This production had ceased to exist in 2018, and road traffic had become the dominant source. The three groups of road transport were together responsible for more than half of the emissions. Aluminium production was the second most important source in 2018. However, HCB emissions in Norway are now negligible.

Table A 17. Key categories for HCB emissions, 1990 and 2018. Key categories are given in bold italic

| Source | | 1990 | 2018 |
|----------|---|---------------|---------------|
| 1A3BI | Road transport: Passenger cars | 0.0 % | 35.1 % |
| 2C3 | Aluminium production | 0.1 % | 15.9 % |
| 1A3BII | Road transport: Light duty vehicles | 0.0 % | 12.6 % |
| 1A1A | Public electricity and heat production | 0.6 % | 8.9 % |
| 1A4BI | Residential: Stationary | 0.1 % | 6.2 % |
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 0.0 % | 5.5 % |
| 1A3DII | National navigation (shipping) | 0.0 % | 4.7 % |
| 2B10A | Chemical industry: Other (please specify in the IIR) | 0.0 % | 3.2 % |
| 2C1 | Iron and steel production | 0.0 % | 1.3 % |
| 1A2GVIII | Stationary Combustion in manufacturing industries and construction: Other (please specify in the IIR) | 0.0 % | 1.2 % |
| 1A1C | Manufacture of solid fuels and other energy industries | 0.0 % | 1.2 % |
| 2C4 | Magnesium production | 98.8 % | . |

Key categories for PCB

Road traffic is by far the most important source for emissions of PCB. There has, however, been a strong shift between the different road traffic groups. In 1990, most of the emissions came from passenger cars, whereas heavy duty vehicles and buses were dominant in 2018.

Table A 18. Key categories for PCB emissions, 1990 and 2018. Key categories are given in bold italic

| Source | | 1990 | 2018 |
|---------|--|---------------|---------------|
| 1A3BIII | Road transport: Heavy duty vehicles and buses | 5.4 % | 86.1 % |
| 5E | Other waste | 1.5 % | 4.3 % |
| 1A3BI | Road transport: Passenger cars | 75.4 % | 4.1 % |
| 1A3BII | Road transport: Light duty vehicles | 5.7 % | 2.0 % |
| 1A4BII | Residential: Household and gardening (mobile) | 7.9 % | 0.0 % |
| 1B2C | Venting and flaring (oil, gas, combined oil and gas) | 2.6 % | 0.0 % |

Emission factors used in the estimations of emissions from combustion

Statistics Norway presents an inventory over important emission factors in:

https://www.ssb.no/_attachment/404602/

In the calculations, the numbers are used with the highest available accuracy. In the tables, though, they are only shown rounded off, which in some cases can lead to the result that the exceptions look the same as the general factors. The tables include the emission factors used for estimating the acidifying pollutants, heavy metals and persistent organic pollutants. Due to practical reasons, also the emission factors used for the Norwegian greenhouse gas inventory are included in the tables.

For road traffic, this general view of the emission factors only includes last year's factors and not all time series.

In the tables for stationary combustion, dotted cells indicate combinations of fuel and source without consumption.

A description of the sector codes used in the tables is given in Appendix D.

The emission factors for BC, PCB and HCB are not included in Statistics Norway's inventory, but some of the factors are presented below in Table B1. These factors vary greatly between sectors, combustion technology and years. In a number of cases, particularly for PCB and HCB, factors are not available and emissions have not been estimated.

Table B1. Emission factors for BC, PCB and HCB

| | BC | PCB | HCB |
|---|---------------------|----------------------|-----------------------|
| | kg/tonne | ug/tonn | ug/tonn |
| Coal | 0.016 ¹ | 160.170 ¹ | 17.422 |
| Coke | 0.016 ¹ | 160.170 ¹ | 17.422 |
| Petrol coke | 0.016 ¹ | 160.170 ¹ | NE |
| Motor gasoline | 0.013 ² | 20 ³ | 0.422 ² |
| Aviation gasoline | 0.005 ⁴ | NE | NE |
| Kerosene (heating) | 0.086 ⁵ | 4.008 | NE |
| Jet kerosene | 0.059 ¹³ | NE | NE |
| Auto diesel | 0.155 ² | 986.339 ² | 456.017 ² |
| Marine gas oil/diesel | 0.600 ⁶ | 355.656 | 80 |
| Light fuel oils | 0.096 ⁵ | 4.008 | 9.482 ⁷ |
| Heavy distillate | 2.210 ⁶ | 603.400 ⁶ | 140 ⁶ |
| Heavy fuel oil | 2.210 ⁶ | 603.400 ⁶ | 140 ⁶ |
| Bio ethanol | 0.014 ² | 20 ² | 0.420 ² |
| Bio diesel | 0.155 ² | 986.34 ² | 456.02 ² |
| Natural gas (1000 Sm ³) | 0.009 ⁸ | NE | NE |
| LPG | 0.009 ⁸ | NE | NE |
| Refinery gas | 0.010 | NE | NE |
| CO gas | 0.010 | NE | NE |
| Fuel gas | 0.010 | NE | NE |
| Landfill gas | 0.010 | NE | NE |
| Biogas | 0.010 ⁸ | 4.518 | NE |
| Fuel wood | 0.653 ⁹ | 15.641 ⁹ | 84 |
| Wood waste | 0.294 | 47.040 | 84 |
| Black liquor | NE | 19.395 | 2.381 |
| Wood pellets | 0.406 ¹⁰ | 7.829 | 84 |
| Municipal waste | 0.0005 | 0.032 ¹¹ | 45.150 ¹² |
| Special waste | 1.008 ¹ | 0.195 ¹¹ | 225.750 ¹² |

¹ Industrial combustion. ² Private cars in 2018. ³ 1997-2018. ⁴ LTO. Cruise=0.001. ⁵ Households.

⁶ Ships. ⁷ Services. ⁸ Stationary combustion. ⁹ New stove. 2018. ¹⁰ Households: 0.037. ¹¹ 2006-2018.

¹² 2005-2018. ¹³ LTO (jet/turboprop). Cruise (jet/turboprop)= 0.08. Helicopter (LTO/cruise)= 0.005/0.001.

Uncertainty analysis

Long-range transboundary air pollutants

Source for the uncertainty estimates for long-range transboundary air pollutants is Rypdal and Zhang (2001).

Table C1. Summary of expert judgements of uncertainties in point sources

| Production type | Number of plants | Pollutant | Emission determination method and uncertainty evaluation | Assessment (average) |
|---|------------------|-----------------|--|----------------------|
| Pulp and paper | 6 | SO ₂ | Continuous emission measurements and estimations from sulphur content of fuel. Diffuse emissions of sulphur compounds when producing sulphite pulp. The latter has a higher uncertainty than both the measured and estimated stack emissions. | ± 4 % |
| Oil refineries | 2 (3) | SO ₂ | Continuous emission measurements and estimations from sulphur content of fuel. | ± 5 % |
| | | NO _x | Based on measurements and calculations. | ± 10 % |
| | | NM VOC | Combination of point measurements and calculations. Emissions are variable with possibilities of systematic errors. Emissions from loading of products have lower uncertainty than the fugitive. Differences between the refineries due to different technology, products and operations. | ± 45 % |
| Petrochemical industries and gas terminal | 4 | NO _x | Annual measurements and/or calculations | ± 7 % |
| | | NM VOC | Several emission points. Difficult to measure properly and high variability. Uncertainty is in any case lower than for the refineries as mostly gas is handled (high demand for security). | ± 25 % |
| Cement | 2 | SO ₂ | Continuous measurements and annual measurements/calculations. High variability as cement plants incinerates special waste. | ± 12 % |
| | | NO _x | Continuous measurements and annual measurements/calculations. High variability as cement plants incinerates special waste. | ± 12 % |
| Ammonia and fertilizer | 2 | NO _x | Continuous/weekly measurements. | ± 7 % |
| | | NH ₃ | Several emission points. Several measurements performed each year. Low variability. | ± 10 % |
| Silicon carbide (SiC) | 3 | SO ₂ | Emissions are estimates based on consumption and sulphur content of coke. The sulphur content is measured independently for every delivery. There is, however, uncertainty connected to the end products and degree of oxidation and definition applied, so reporting can seem inconsistent. | ± 20 % |
| Ferroalloys | 16 | SO ₂ | Emissions are estimates based on consumption and sulphur content of coke and the sulphur in products. The sulphur content is measured independently for every delivery. The sulphur content of products are measured regularly, but shows small variability. | ± 2 % |
| | | NO _x | Estimates using emission factors. Emission factors are based on measurements. Emission factors are, however, only available for some types of ferroalloys and emissions are not estimated for the others. | ± 10-20 %* |
| Aluminium | 8 | SO ₂ | Monthly measurements (covering emissions from stack and ceiling) | ± 7 % |
| | | NO _x | Emissions are estimated based on emission factors (see table 4). | - |
| Waste incineration | 8 | SO ₂ | Annual representative measurements. Variable emissions due to the waste fraction incinerated. | ± 7 % |
| | | NO _x | Annual representative measurements. | ± 10 % |

* Additional uncertainty due to possible incomplete reporting.

Table C2. Summary of standard deviation and probability density of activity data

| SNAP category | Pollutant source | Important for | Standard deviation (2σ). % | Density shape | Source/Comment |
|--------------------|--|--|----------------------------|---------------|----------------------------|
| 01, 02, 03 | Gas combustion | NO _x | ± 4 | Normal | Directorate of oil and gas |
| 01, 02, 03, 07, 08 | Oil combustion (total) | SO ₂ , NO _x | ± 3 | Normal | Spread in data. |
| 0102 | Waste combustion - Energy industries | SO ₂ , NO _x , NMVOC | ± 5 | Normal | Expert judgement |
| 0202 | Coal and coke combustion - Residential | SO ₂ , NO _x , NMVOC | ± 20 | Normal | Expert judgement |
| 090201 | Waste combustion - Other sectors | SO ₂ , NO _x , NMVOC | ± 30 | Lognormal | Expert judgement |
| 01, 02, 03 | Wood combustion - All sectors | SO ₂ , NO _x , NMVOC | ± 30 | Lognormal | Expert judgement |
| 01, 03 | Coal and coke combustion- Industry | SO ₂ , NO _x , NMVOC | ± 5 | Normal | Spread in data |
| 07, 08 | Oil, road/off-road/catalytic/non-catalytic | SO ₂ , NO _x , NMVOC, NH ₃ | ± 20 | Normal | Comparisons of data |
| 0805 | Oil combustion - Aviation | SO ₂ , NO _x , NMVOC | ± 20 | Normal | Expert judgement |
| 0804 | Oil combustion - Shipping | SO ₂ , NO _x , NMVOC | ± 10 | Normal | Comparisons of data |
| 0401 | Refineries (throughput) | NMVOC | ± 3 | Normal | Expert judgement |
| 040301 | Aluminium production | NO _x | ± 3 | Normal | Expert judgement |
| 040302 | Ferroalloy production | NO _x | ± 3 | Normal | Expert judgement |
| 040605 | Bread production | NMVOC | ± 30 | Normal | Expert judgement |
| 040607 | Beer production | NMVOC | ± 10 | Normal | Expert judgement |
| 050202 | Loading of crude oil | NMVOC | ± 3 | Normal | Expert judgement |
| 0505 | Gasoline distribution | NMVOC | ± 3 | Normal | Expert judgement |
| 0601 | Solvent use | NMVOC | | | See emission factor |
| 09 | Waste combustion in small scale | SO ₂ , NO _x , NMVOC | ± 50 | Lognormal | Expert judgement |
| 090201 | Methane incineration (landfills) | NO _x , NMVOC | ± 5 | Normal | Expert judgement |
| 090204 | Flaring of natural gas | NO _x , NMVOC | ± 4 | Normal | As combustion of gas |
| 090204 | "Flaring" of crude oil | SO ₂ , NO _x , NMVOC | ± 10 | Normal | Expert judgement |
| 090203/4 | Other flaring | NO _x , NMVOC | ± 5 | Normal | Expert judgement |
| 090207 | Incineration of hospital waste | NO _x , NMVOC | ± 20 | Normal | Expert judgement |
| 090901 | Cremation | SO ₂ , NO _x , NMVOC | ± 20 | Normal | Expert judgement |
| 10 | Animal population | NH ₃ | ± 5-10 | Normal | Expert judgement |
| 10 | Agricultural soils - Treatment of straw | NH ₃ | | | See emission factor |
| 1001 | Agricultural soils - Fertilizer use | NH ₃ | ± 5 | Normal | Agriculture authorities |
| 1009 | Agricultural soils - Manure use | NH ₃ | ± 20 | Normal | Expert judgement |

Table C3. Summary of standard deviation and probability density of emission factors

| SNAP source category | Pollutant source | Standard deviation (2σ). % | Density shape | Source/Comment |
|----------------------|---|----------------------------|---------------|----------------------------------|
| 01, 02, 03 | SO ₂ - Oil combustion, general | ± 1 | Normal | Expert judgement. Oil companies |
| 01, 02, 03 | SO ₂ - Oil combustion, heavy fuel oil | -50 - +100 | Normal | Expert judgement. Oil companies |
| 01, 03 | SO ₂ - Coal combustion | -50 - +100 | Lognormal | Spread in data |
| 01, 03 | SO ₂ - Wood combustion | -50 - +100 | Lognormal | Spread in data |
| 0804 | SO ₂ - Oil combustion, domestic shipping | ± 25 | Normal | Expert judgement. Oil companies |
| 01, 02 (+03) | NO _x - Combustion in area sources | ± 40-50 | Normal | Spread in data |
| 0105 | NO _x - Combustion off-shore | ± 40 | Lognormal | Expert judgement |
| 040301 | NO _x - Aluminium production | -50 - +100 | Lognormal | Expert judgement |
| 07 | NO _x - Road traffic | ± 25-30 | Normal | Expert judgement, spread in data |
| 0704/0705 | NO _x - Motorcycles | ± 40 | Normal | Expert judgement, spread in data |
| 0801-02, 0806-09 | NO _x - Equipment and railways | ± 40 | Normal | Spread in data |
| 0804 | NO _x - Shipping | ± 15 | Normal | Spread in data |
| 0805 | NO _x - Aircraft | ± 20 | Normal | EEA (2000) |
| 0902 | NO _x - Flaring | ± 40 | Lognormal | Expert judgement |
| 01, 02 (+03) | NM VOC - Combustion in area sources | ± 40-50 | Normal | Spread in data |
| 0105 | NM VOC - Combustion offshore | ± 50 | Lognormal | Expert judgement |
| 040605/07 | NM VOC- Beer and bread production | -50 - +100 | Lognormal | EEA (2000) |
| 050201 | NM VOC- Oil loading onshore | ± 30 | Normal | Rypdal (1999), Expert judgement |
| 050202 | NM VOC- Oil loading offshore | ± 40 | Normal | Rypdal (1999), Expert judgement |
| 0505 | NM VOC - Gasoline distribution | ± 50 | Lognormal | EEA (2000) |
| 0601 | NM VOC - Solvent use | ± 30 | Normal | Rypdal (1995) |
| 0701 | NM VOC - Road traffic (gasoline vehicles) | ± 40-50 | Normal | Expert judgement, spread in data |
| 0703 | NM VOC - Road traffic (diesel vehicles) | ± 20-30 | Normal | Expert judgement, spread in data |
| 0704/0705 | NM VOC - Motorcycles | ± 40 | Normal | Expert judgement, spread in data |
| 0801-02, 0806-09 | NM VOC - Equipment and railways | ± 40 | Normal | Spread in data |
| 0804 | NM VOC - Shipping | ± 50 | Normal | Spread in data |
| 0805 | NM VOC - Aircraft | ± 25 | Normal | EEA (2000) |
| 0902 | NM VOC - Flaring | ± 50 | Lognormal | Expert judgement |
| 07 | NH ₃ - Road traffic | Factor 3 | Lognormal | Expert judgement, spread in data |
| 1001 | NH ₃ -Agriculture, fertilizer | ± 20 | Normal | Expert judgement |
| 1005 | NH ₃ -Agriculture, animal manure | ± 30 | Normal | Expert judgement |
| 10 | NH ₃ -Agriculture, treatment of straw | ± 5 | Normal | Expert judgement |

Table C4. Uncertainty in emission level of pollutants. 1990, 1998 and 2010

| 1990 | μ (mean) ktonnes | Relative standard deviation (σ/μ) | Uncertainty 2 σ (% of mean) | Uncertainty 2 σ (ktonnes) |
|-----------------|-------------------------|---|---------------------------------------|-------------------------------------|
| SO ₂ | 52.7 | 0.02 | 4.0 | 2 |
| NO _x | 219.0 | 0.062 | 12 | 27 |
| NM VOC | 298.4 | 0.09 | 18 | 54 |
| NH ₃ | 22.9 | 0.104 | 21 | 5 |
| 1998 | μ (mean) ktonnes | Relative standard deviation (σ/μ) | Uncertainty 2 σ (% of mean) | Uncertainty 2 σ (ktonnes) |
| SO ₂ | 29.8 | 0.021 | 4.2 | 1 |
| NO _x | 224.0 | 0.062 | 12 | 27 |
| NM VOC | 344.5 | 0.105 | 21 | 72 |
| NH ₃ | 27.0 | 0.091 | 18 | 5 |
| 2010* | μ (mean) ktonnes | Relative standard deviation (σ/μ) | Uncertainty 2 σ (% of mean) | Uncertainty 2 σ (ktonnes) |
| SO ₂ | 22.0 | 0.025 | 5.0 | 1 |
| NO _x | 156.0 | 0.062 | 12 | 19 |
| NM VOC | 194.0 | 0.074 | 15 | 29 |
| NH ₃ | 23.0 | 0.105 | 21 | 5 |

* Projected data with uncertainties as if they were historical.

Table C5. Uncertainties in emission trends 1990-1998 and 1990-2010

| | Absolute change ($\mu_{2010}-\mu_{1990}$) | % change ($(\mu_{2010}-\mu_{1990}) \cdot 100 / \mu_{1990}$) | Relative standard deviation ($\sigma / (\mu_{2010}-\mu_{1990})$) | Uncertainty 2 σ (absolute change) | Uncertainty 2 σ (%-point of change) |
|-----------------|--|--|---|---|---|
| 1990-1998 | | | | | |
| SO ₂ | -23.0 | -43 | -0.04 | 1.7 | 3.2 |
| NO _x | +4.8 | +2 | +3.00 | 28 | 13 |
| NM VOC | +43.8 | +15 | +0.40 | 35 | 12 |
| NH ₃ | +4.1 | +18 | +0.22 | 1.8 | 8.0 |
| 1990-2010 | | | | | |
| SO ₂ | -30.7 | -58 | -0.03 | 1.8 | 3.4 |
| NO _x | -62.8 | -29 | -0.21 | 26.9 | 12 |
| NM VOC | -104.9 | -35 | -0.18 | 38 | 13 |
| NH ₃ | +0.0 | 0 | 61.3 | 3.1 | 13 |

* Projected values with uncertainties as if they were historical.

Economic sectors in the Norwegian emission model

The classification is an aggregated version of the one used in the national accounts. To make the standard sectors more appropriate for emission calculations, a few changes have been made, e.g. "Private households" is defined as a sector. The classification is aggregated from the Norwegian *Standard Industrial Classification*, SIC2007 (Statistics Norway 2009). The sic is identical to the European NACE (rev. 2) classification up to the four-digit level. A national level has been introduced at the five-digit level.

The sector numbers in the model have six or, in a few cases, eight digits. The first two digits refer to the main sectors of the economy: 23 = private sector, 24 = central government, 25 = local government, 33 = private households, and 66 = foreign activity. For clarity, the two first digits are only included for the first sector listed in each main sector in the table below.

The next four digits are approximate SIC codes. The first two of these in most cases correspond to SIC at the two-digit level, but some sector numbers, particularly those used for service industries, are aggregates of several SIC divisions. The detailed relationship is shown in the following table, where the sectors are listed with the corresponding SIC codes.

For emissions from solvents and paraffin wax, figures are available at a somewhat more disaggregated sector level, but since these sectors do not reflect the general detailing level in the emission calculations, they are not included in the table below.

| Sector number | SIC code | Sector name |
|---------------------------------|----------------------------|--|
| Agriculture and forestry | | |
| 230100 | 01.01-5, 01.7 | Agriculture |
| 0160 | 01.6 | Services related to agriculture |
| 0210 | 02 | Forestry and logging |
| Fishing | | |
| 0310-N | 03.1 | Fishing |
| 0320 | 03.2 | Operation of fish farms |
| Energy sectors | | |
| 0500 | 05 | Coal mining |
| 0600.1 | 06 part, 49.5 | Extraction of crude petroleum and natural gas, offshore: Permanent installations |
| 0600.2 | 06 part | Extraction of crude petroleum and natural gas, offshore: Moveable installations |
| 0600.3 | 06 part | Extraction of crude petroleum and natural gas: Plants on shore |
| 1910.2 | 19.1 part | Coking plants |
| 1922 | 19.2 part | Manufacture of refined petroleum products |
| 3510 | 35.12, 35.13, 35.14 | Transmission, distribution and trade of electricity |
| 3511 | 35.11 | Production of electricity |
| 3520 | 35.2 | Manufacture and distribution of gas |
| 3530 | 35.3 | Steam and hot water supply |
| Mining/manufacturing | | |
| 0710 | 07.1, 07.29 | Mining of ores except uranium and thorium |
| 0721 | 07.21 | Mining of uranium and thorium ores |
| 0810 | 08 except 08.92 | Quarrying and mining except ores and extraction of peat |
| 0892 | 08.92 | Extraction and agglomeration of peat |
| 0910 | 09.1, 52.215 | Service activities incidental to oil and gas extraction |
| 0990 | 09.9 | Service activities incidental to mining |
| 1010 | 10.1 | Production, processing and preserving of meat and meat products |
| 1020 | 10.2 | Processing and preserving of fish and fish products |
| 1030 | 10.3 | Processing and preserving of fruit and vegetables |
| 1040 | 10.4 | Manufacture of vegetable and animal oils and fats |
| 1050 | 10.5 | Manufacture of dairy products |
| 1060 | 10.6 | Manufacture of grain mill products, starches and starch products |
| 1080 | 10.7, 10.8 | Manufacture of other food products |
| 1090 | 10.9 | Manufacture of prepared animal feeds |
| 1100 | 11 | Manufacture of beverages |
| 1200 | 12 | Manufacture of tobacco products |
| 1300 | 13 | Manufacture of textiles and textile products |
| 1400 | 14 | Manufacture of wearing apparel |
| 1500 | 15 | Manufacture of leather, leather products and footwear |
| 1610 | 16.1 | Sawmilling and planing of wood, impregnation of wood |
| 1620 | 16.21, 16.22, 16.24, 16.29 | Manufacture of products of wood, cork, straw and plaiting materials, except furniture |
| 1630 | 16.23 | Manufacture of builders' supplies |
| 1711 | 17.11 | Manufacture of pulp |
| 1712 | 17.12 | Manufacture of paper and paperboard |
| 1720 | 17.2 | Manufacture of articles of paper and paperboard |
| 1800 | 18 | Printing and service activities related to printing and reproduction of recorded media |
| 1910.1 | 19.1 part | Manufacture of coke oven products |
| 1921 | 19.2 part | Manufacture of refined petroleum products except oil refineries |
| 2011 | 20.11, 20.12, 20.13 | Manufacture of basic chemicals |
| 2014 | 20.14 | Manufacture of other organic basic chemicals |
| 2015 | 20.15 | Manufacture of fertilizers and nitrogen compounds |

| Sector number | SIC code | Sector name |
|---------------|-------------------------------|--|
| 2016 | 20.16, 20.17 | Manufacture of plastics and synthetic rubber in primary forms |
| 2020 | 20.2 | Manufacture of pesticides and other agrochemical products |
| 2030 | 20.3 | Manufacture of paints and varnishes, printing ink and mastics |
| 2040 | 20.4 | Manufacture of soap and detergents and toilet preparations |
| 2050 | 20.5, 20.6 | Manufacture of other chemical products |
| 2100 | 21 | Manufacture of basic pharmaceutical products and pharmaceutical preparations |
| 2200 | 22 | Manufacture of rubber and plastic products |
| 2310 | 23.1 | Manufacture of glass and glass products |
| 2320 | 23.2, 23.3, 23.4 | Manufacture of refractory products, clay building materials and other porcelain and ceramic products |
| 2350 | 23.5 | Manufacture of cement, lime and plaster |
| 2360 | 23.6, 23.7, 23.9 | Manufacture of products of cement, lime and plaster and other non-metallic mineral products |
| 2411 | 24.101, 24.2, 24.3 | Manufacture of basic iron and steel |
| 2412 | 24.102 | Manufacture of ferroalloys |
| 2440 | 24.4 except 24.42 | Other non-ferrous metal production |
| 2442 | 24.42 | Aluminium production |
| 2451 | 24.51, 24.52 | Casting of iron and steel |
| 2453 | 24.53, 24.54 | Casting of light metals and other non-ferrous metals |
| 2510 | 25.1, 25.2, 25.3 | Manufacture of structural metal products, tanks, reservoirs and containers etc. of metal |
| 2570 | 25.7 | Manufacture of cutlery, tools and general hardware |
| 2590 | 25.4, 25.5, 25.6, 25.9 | Manufacture of other metal products |
| 2610 | 26.1, 26.2 | Manufacture of electronic components and computers |
| 2630 | 26.3 | Manufacture of communication equipment |
| 2640 | 26.4 | Manufacture of consumer electronics |
| 2650 | 26.5, 26.6, 26.7, 26.8 | Manufacture of other electronic and optical products |
| 2750 | 27.5 | Manufacture of domestic appliances |
| 2790 | 27.1, 27.2, 27.3, 27.4, 27.9 | Manufacture of other electrical apparatus and equipment |
| 2810 | 28.1, 28.2 | Manufacture of general-purpose machinery |
| 2830 | 28.3, 28.4, 28.9 | Manufacture of special-purpose machinery |
| 2900 | 29 | Manufacture of motor vehicles and parts and accessories for motor vehicles |
| 3011 | 30.1 except 30.113 and 30.116 | Building of ships and boats |
| 3012 | 30.113, 30.116 | Building of oil platforms |
| 3020 | 30.2 | Manufacture of railway and tramway locomotives and rolling stock |
| 3030 | 30.3 | Manufacture of aircraft and spacecraft |
| 3090 | 30.4, 30.9 | Manufacture of other transport equipment |
| 3100 | 31 | Manufacture of furniture |
| 3210 | 32.1 | Manufacture of jewellery, bijouterie and related articles |
| 3290 | 32.2, 32.3, 32.4, 32.5, 32.9 | Other manufacturing |
| 3310 | 33.1 | Repair of fabricated metal products, machinery and equipment |
| 3320 | 33.2 | Installation of industrial machinery and equipment |

Water supply, sewerage, waste management and remediation activities

| | | |
|------|-------|---|
| 3600 | 36 | Water collection, treatment and supply |
| 3800 | 37-39 | Sewerage, waste collection, treatment and disposal activities; materials recovery |

Construction

| | | |
|------|--------------|--------------|
| 4120 | 41.2, 42, 43 | Construction |
|------|--------------|--------------|

| Sector number | SIC code | Sector name |
|--|--|---|
| Wholesale and retail trade | | |
| 4700 | 45, 46, 47 | Wholesale and retail trade, repair of motor vehicles and personal and household goods |
| Transport etc. | | |
| 4910 | 49.1, 49.2 | Transport via railways |
| 4932 | 49.32 | Taxi operation |
| 4939 | 49.31, 49.39 | Other land passenger transport |
| 4940 | 49.4 | Freight transport by road |
| 5020.N | 50.101, 50.201 | Ocean transport |
| 5030 | 50.102, 50.109, 50.202, 50.203, 50.204, 50.3, 50.4 | Inland and coastal water transport |
| 5100.1N | 51 part | Domestic air transport |
| 5100.2N | 51 part | International air transport |
| 5222 | 52 except 52.215, 79 | Supporting and auxiliary transport activities |
| 5300 | 53, 61 | Post and telecommunications |
| Accommodation and food service activities | | |
| 5500 | 55, 56 | Accommodation, food and beverage service activities |
| Business services | | |
| 5800 | 58 | Publishing activities |
| 6200 | 62, 63, 95 | Information technology services |
| 6600 | 64, 65, 66 | Financial and insurance activities |
| 6810 | 41.1, 68 | Real estate activities |
| 7100 | 69-71, 73-74, 78, 80-82 | Other business activities |
| 7200 | 72 | Research and development |
| 7700 | 77 | Rental and leasing activities |
| 8500 | 85 | Education |
| 8600 | 75, 86-88 | Health and social work |
| 9300 | 59-60, 90-93 | Recreational, cultural and sporting activities |
| 9400 | 94, 99 | Activities of membership organisations |
| 9600 | 96 | Other service activities |
| Central government | | |
| 245222 | 52, 79 | Supporting and auxiliary transport activities |
| 7100 | 69-71, 73-74, 78, 80-82 | Other business activities |
| 7200 | 72 | Research and development |
| 8410 | 84.1, 84.21, 84.23, 84.24, 84.25, 84.3 | Public administration |
| 8422 | 84.22 | Defence |
| 8500 | 85 | Education |
| 8600 | 75, 86-88 | Health and social work |
| 9300 | 59-60, 90-93 | Other service activities |
| Local government | | |
| 253700 | 37 | Sewerage |
| 3800 | 38 | Waste collection, treatment and disposal activities; materials recovery |
| 6000 | 59-60, 90-93 | Other service activities |
| 8410 | 84.1, 84.21, 84.23, 84.24, 84.25, 84.3 | Public administration |
| 8500 | 85 | Education |
| 8600 | 75, 86-88 | Health and social work |
| 9600 | 96 | Other personal service activities |

| Sector number | SIC code | Sector name |
|------------------------------|----------|---|
| Private households | | |
| 330000 | n.a. | Private households |
| Foreign activities in Norway | | |
| 665020 | n.a. | Foreign activities in Norway, ocean transport |
| 665100.2 | n.a. | Foreign activities in Norway, air transport |

Source classifications used in the Norwegian emission inventory

Source classifications used in the official statistics on emissions to air published by Statistics Norway is given at the webpage: <http://www.ssb.no/en/klass/klassifikasjoner/113>

In the reported inventory EMEP/NFR14 source sector categories are being used.

Table E1. EMEP/NFR14 source sector categories

| | |
|-----------|--|
| 1A1a | Public electricity and heat production |
| 1A1b | Petroleum refining |
| 1A1c | Manufacture of solid fuels and other energy industries |
| 1A2a | Stationary combustion in manufacturing industries and construction: Iron and steel |
| 1A2b | Stationary combustion in manufacturing industries and construction: Non-ferrous metals |
| 1A2c | Stationary combustion in manufacturing industries and construction: Chemicals |
| 1A2d | Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print |
| 1A2e | Stationary combustion in manufacturing industries and construction: Food processing, beverages and tobacco |
| 1A2f | Stationary combustion in manufacturing industries and construction: Non-metallic minerals |
| 1A2gvii | Mobile Combustion in manufacturing industries and construction: (please specify in the IIR) |
| 1A2gviii | Stationary combustion in manufacturing industries and construction: Other (please specify in the IIR) |
| 1A3ai(i) | International aviation LTO (civil) |
| 1A3aii(i) | Domestic aviation LTO (civil) |
| 1A3bi | Road transport: Passenger cars |
| 1A3bii | Road transport: Light duty vehicles |
| 1A3biii | Road transport: Heavy duty vehicles and buses |
| 1A3biv | Road transport: Mopeds & motorcycles |
| 1A3bv | Road transport: Gasoline evaporation |
| 1A3bvi | Road transport: Automobile tyre and brake wear |
| 1A3bvii | Road transport: Automobile road abrasion |
| 1A3c | Railways |
| 1A3di(ii) | International inland waterways |
| 1A3dii | National navigation (shipping) |
| 1A3ei | Pipeline transport |

| | |
|---------|--|
| 1A3eii | Other (please specify in the IIR) |
| 1A4ai | Commercial/institutional: Stationary |
| 1A4aai | Commercial/institutional: Mobile |
| 1A4bi | Residential: Stationary |
| 1A4bii | Residential: Household and gardening (mobile) |
| 1A4ci | Agriculture/Forestry/Fishing: Stationary |
| 1A4cii | Agriculture/Forestry/Fishing: Off-road vehicles and other machinery |
| 1A4ciii | Agriculture/Forestry/Fishing: National fishing |
| 1A5a | Other stationary (including military) |
| 1A5b | Other, Mobile (including military, land based and recreational boats) |
| 1B1a | Fugitive emission from solid fuels: Coal mining and handling |
| 1B1b | Fugitive emission from solid fuels: Solid fuel transformation |
| 1B1c | Other fugitive emissions from solid fuels |
| 1B2ai | Fugitive emissions oil: Exploration, production, transport |
| 1B2aiv | Fugitive emissions oil: Refining / storage |
| 1B2av | Distribution of oil products |
| 1B2b | Fugitive emissions from natural gas (exploration, production, processing, transmission, storage, distribution and other) |
| 1B2c | Venting and flaring (oil, gas, combined oil and gas) |
| 1B2d | Other fugitive emissions from energy production |
| 2A1 | Cement production |
| 2A2 | Lime production |
| 2A3 | Glass production |
| 2A5a | Quarrying and mining of minerals other than coal |
| 2A5b | Construction and demolition |
| 2A5c | Storage, handling and transport of mineral products |
| 2A6 | Other mineral products (please specify in the IIR) |
| 2B1 | Ammonia production |
| 2B2 | Nitric acid production |
| 2B3 | Adipic acid production |
| 2B5 | Carbide production |
| 2B6 | Titanium dioxide production |
| 2B7 | Soda ash production |
| 2B10a | Chemical industry: Other (please specify in the IIR) |
| 2B10b | Storage, handling and transport of chemical products (please specify in the IIR) |
| 2C1 | Iron and steel production |
| 2C2 | Ferroalloys production |
| 2C3 | Aluminium production |
| 2C4 | Magnesium production |
| 2C5 | Lead production |
| 2C6 | Zinc production |

| | |
|---------|--|
| 2C7a | Copper production |
| 2C7b | Nickel production |
| 2C7c | Other metal production (please specify in the IIR) |
| 2C7d | Storage, handling and transport of metal products (please specify in the IIR) |
| 2D3a | Domestic solvent use including fungicides |
| 2D3b | Road paving with asphalt |
| 2D3c | Asphalt roofing |
| 2D3d | Coating applications |
| 2D3e | Degreasing |
| 2D3f | Dry cleaning |
| 2D3g | Chemical products |
| 2D3h | Printing |
| 2D3i | Other solvent use (please specify in the IIR) |
| 2G | Other product use (please specify in the IIR) |
| 2H1 | Pulp and paper industry |
| 2H2 | Food and beverages industry |
| 2H3 | Other industrial processes (please specify in the IIR) |
| 2I | Wood processing |
| 2J | Production of POPs |
| 2K | Consumption of POPs and heavy metals (e.g. electrical and scientific equipment) |
| 2L | Other production, consumption, storage, transportation or handling of bulk products (please specify in the IIR) |
| 3B1a | Manure management - Dairy cattle |
| 3B1b | Manure management - Non-dairy cattle |
| 3B2 | Manure management - Sheep |
| 3B3 | Manure management - Swine |
| 3B4a | Manure management - Buffalo |
| 3B4d | Manure management - Goats |
| 3B4e | Manure management - Horses |
| 3B4f | Manure management - Mules and asses |
| 3B4gi | Manure management - Laying hens |
| 3B4gii | Manure management - Broilers |
| 3B4giii | Manure management - Turkeys |
| 3B4giv | Manure management - Other poultry |
| 3B4h | Manure management - Other animals (please specify in IIR) |
| 3Da1 | Inorganic N-fertilizers (includes also urea application) |
| 3Da2a | Animal manure applied to soils |
| 3Da2b | Sewage sludge applied to soils |
| 3Da2c | Other organic fertilizers applied to soils (including compost) |

| | |
|---------|---|
| 3Da3 | Urine and dung deposited by grazing animals |
| 3Da4 | Crop residues applied to soils |
| 3Db | Indirect emissions from managed soils |
| 3Dc | Farm-level agricultural operations including storage, handling and transport of agricultural products |
| 3Dd | Off-farm storage, handling and transport of bulk agricultural products |
| 3De | Cultivated crops |
| 3Df | Use of pesticides |
| 3F | Field burning of agricultural residues |
| 3I | Agriculture other (please specify in the IIR) |
| 5A | Biological treatment of waste - Solid waste disposal on land |
| 5B1 | Biological treatment of waste - Composting |
| 5B2 | Biological treatment of waste - Anaerobic digestion at biogas facilities |
| 5C1a | Municipal waste incineration |
| 5C1bi | Industrial waste incineration |
| 5C1bii | Hazardous waste incineration |
| 5C1biii | Clinical waste incineration |
| 5C1biv | Sewage sludge incineration |
| 5C1bv | Cremation |
| 5C1bvi | Other waste incineration (please specify in the IIR) |
| 5C2 | Open burning of waste |
| 5D1 | Domestic wastewater handling |
| 5D2 | Industrial wastewater handling |
| 5D3 | Other wastewater handling |
| 5E | Other waste (please specify in IIR) |
| 6A | Other (included in national total for entire territory) (please specify in IIR) |

Activity data used in the Norwegian inventory

Table F1. Activity data used for Mineral Products (2A)

| | 2A1-Cement production | 2A2-Lime production | 2A5b-Construction and demolition | 2A6-Other mineral products |
|------|-----------------------|---------------------|---|--------------------------------------|
| | Clinker produced (kt) | Lime produced (kt) | floor space constructed/ demolished (M ³) | sand and crushed stone produced (kt) |
| 1990 | 1244.07 | 62.04 | 644.25 | 39949.02 |
| 1991 | 1147.47 | 58.74 | 496.85 | 42149.04 |
| 1992 | 1260.22 | 59.82 | 414.77 | 40505.84 |
| 1993 | 1605.90 | 64.77 | 465.59 | 43290.53 |
| 1994 | 1621.42 | 70.41 | 489.53 | 17396.16 |
| 1995 | 1682.94 | 86.76 | 538.51 | 17396.16 |
| 1996 | 1655.39 | 102.43 | 570.94 | 17416.38 |
| 1997 | 1775.88 | 100.96 | 581.35 | 24956.89 |
| 1998 | 1675.49 | 107.84 | 670.52 | 19044.94 |
| 1999 | 1663.64 | 79.96 | 674.92 | 20670.05 |
| 2000 | 1649.60 | 84.56 | 655.79 | 29645.71 |
| 2001 | 1592.68 | 77.98 | 683.72 | 28329.08 |
| 2002 | 1634.36 | 85.95 | 678.65 | 28995.90 |
| 2003 | 1681.09 | 115.80 | 629.39 | 35276.36 |
| 2004 | 1323.96 | 114.71 | 643.29 | 35320.40 |
| 2005 | 1454.26 | 102.62 | 754.60 | 41683.38 |
| 2006 | 1507.19 | 113.40 | 692.39 | 44714.95 |
| 2007 | 1636.82 | 122.77 | 827.49 | 52487.07 |
| 2008 | 1533.99 | 189.01 | 832.02 | 37397.99 |
| 2009 | 1528.30 | 188.79 | 722.08 | 43310.83 |
| 2010 | 1433.78 | 315.22 | 648.37 | 42437.13 |
| 2011 | 1415.45 | 294.37 | 679.23 | 46602.28 |
| 2012 | 1399.08 | 289.00 | 820.32 | 74769.46 |
| 2013 | 1399.81 | 293.51 | 805.32 | 52500.00 |
| 2014 | 1374.87 | 295.31 | 826.59 | 50580.18 |
| 2015 | 1284.10 | 286.06 | 835.08 | 51761.73 |
| 2016 | 1306.29 | 293.36 | 897.95 | 58515.06 |
| 2017 | 1461.48 | 282.09 | 904.67 | 70843.49 |
| 2018 | 1429.44 | 269.18 | 886.64 | 63236.95 |

Table F2. Activity data used for Chemical Industry (2B)

| | 2B1-Ammonia production | 2B2-Nitric acid production | 2B5-Carbide production |
|------|------------------------|----------------------------|------------------------|
| | Ammonia produced (kt) | Nitric acid produced (kt) | Carbide produced (kt) |
| 1990 | 393.31 | 1326.32 | 197.41 |
| 1991 | 374.14 | 1346.80 | 172.49 |
| 1992 | 330.88 | 1143.16 | 180.58 |
| 1993 | 385.58 | 1373.54 | 164.22 |
| 1994 | 405.49 | 1422.18 | 157.34 |
| 1995 | 433.62 | 1493.81 | 164.82 |
| 1996 | 415.36 | 1478.92 | 157.24 |
| 1997 | 372.13 | 1449.02 | 152.89 |
| 1998 | 242.07 | 1507.32 | 160.35 |
| 1999 | 148.23 | 1481.67 | 142.91 |
| 2000 | 406.29 | 1503.92 | 140.61 |
| 2001 | 393.13 | 1533.66 | 147.78 |
| 2002 | 407.02 | 1546.28 | 99.09 |
| 2003 | 429.81 | 1593.82 | 36.49 |
| 2004 | 510.26 | 1598.82 | 38.39 |
| 2005 | 366.04 | 1589.06 | 34.96 |
| 2006 | 413.06 | 1626.74 | 26.41 |
| 2007 | 334.70 | 1619.88 | 27.24 |
| 2008 | 394.90 | 1565.21 | 28.88 |
| 2009 | 361.07 | 1151.35 | 18.88 |
| 2010 | 378.51 | 1649.98 | 27.22 |
| 2011 | 352.71 | 1660.69 | 25.33 |
| 2012 | 378.26 | 1707.14 | 15.84 |
| 2013 | 350.63 | 1638.95 | 16.92 |
| 2014 | 333.18 | 1669.51 | 17.87 |
| 2015 | 465.08 | 1729.65 | 17.60 |
| 2016 | 373.77 | 1669.20 | 18.68 |
| 2017 | 244.49 | 1733.65 | 21.20 |
| 2018 | 414.14 | 1980.58 | 21.70 |

Table F3. Activity data used for Metal Production (2C)

| | 2C1-Iron and steel production | 2C2-Ferroalloys production | 2C3-Aluminium production | 2C4-Magnesium production |
|------|-------------------------------|-----------------------------|--------------------------|--------------------------|
| | Steel produced (kt) | Use of reducing agents (kt) | Aluminium produced (kt) | Magnesium produced (kt) |
| 1990 | 289.83 | 834.02 | 869.65 | 48.22 |
| 1991 | 372.64 | 699.51 | 882.70 | 44.62 |
| 1992 | 455.45 | 761.27 | 861.93 | 32.73 |
| 1993 | 506.40 | 826.98 | 882.82 | 27.18 |
| 1994 | 454.49 | 935.77 | 854.35 | 29.42 |
| 1995 | 505.39 | 984.98 | 842.11 | 31.66 |
| 1996 | 500.01 | 996.00 | 848.63 | 37.67 |
| 1997 | 566.90 | 1010.88 | 913.79 | 43.30 |
| 1998 | 643.82 | 1045.63 | 990.38 | 43.35 |
| 1999 | 606.55 | 1042.57 | 1003.55 | 50.61 |
| 2000 | 679.28 | 1073.56 | 1026.09 | 49.49 |
| 2001 | 629.58 | 961.85 | 1033.16 | 48.33 |
| 2002 | 679.61 | 829.83 | 1072.81 | 15.30 |
| 2003 | 685.99 | 844.39 | 1182.38 | 19.00 |
| 2004 | 710.00 | 1012.87 | 1317.69 | 30.06 |
| 2005 | 686.94 | 933.89 | 1389.03 | 23.31 |
| 2006 | 669.92 | 745.97 | 1382.55 | 8.70 |
| 2007 | 694.34 | 831.65 | 1363.50 | NO |
| 2008 | 546.53 | 886.61 | 1369.54 | NO |
| 2009 | 546.06 | 566.78 | 1119.27 | NO |
| 2010 | 516.91 | 858.63 | 1101.71 | NO |
| 2011 | 598.36 | 903.14 | 1119.51 | NO |
| 2012 | 682.13 | 917.44 | 1140.40 | NO |
| 2013 | 585.37 | 994.18 | 1165.10 | NO |
| 2014 | 585.52 | 1103.90 | 1172.56 | NO |
| 2015 | 579.65 | 1066.11 | 1216.51 | NO |
| 2016 | 607.95 | 1110.18 | 1243.16 | NO |
| 2017 | 603.54 | 1120.24 | 1249.72 | NO |
| 2018 | 572.44 | 1101.03 | 1295.97 | NO |

Table F4. Activity data used for Solvents and Product Use (2D)

| | 2D3b-Road paving with asphalt | 2D3d-Coating applications | 2D3e-Degreasing | 2D3f-Dry cleaning | 2D3g-Chemical products |
|------|-------------------------------|---------------------------|--------------------|--------------------|------------------------|
| | Asphalt Production (kt) | Solvents used (kt) | Solvents used (kt) | Solvents used (kt) | Solvents used (kt) |
| 1990 | 246,90 | 19,14 | 2,85 | 2,74 | 1437,96 |
| 1991 | 246,90 | 16,20 | 2,41 | 2,32 | 2088,17 |
| 1992 | 246,90 | 16,10 | 1,95 | 1,88 | 2127,17 |
| 1993 | 246,90 | 16,08 | 0,37 | 0,36 | 3926,93 |
| 1994 | 246,90 | 18,56 | 0,29 | 0,28 | 3453,77 |
| 1995 | 246,90 | 14,06 | 0,21 | 0,21 | 4317,47 |
| 1996 | 245,75 | 14,65 | 0,36 | 0,34 | 3942,32 |
| 1997 | 258,40 | 14,42 | 0,32 | 0,31 | 3319,31 |
| 1998 | 228,10 | 12,06 | 0,39 | 0,37 | 3280,82 |
| 1999 | 238,45 | 13,76 | 0,38 | 0,37 | 3165,36 |
| 2000 | 225,25 | 12,63 | 0,42 | 0,41 | 2853,34 |
| 2001 | 203,00 | 13,46 | 0,49 | 0,41 | 2341,17 |
| 2002 | 195,35 | 14,30 | 0,57 | 0,42 | 1858,77 |
| 2003 | 205,90 | 15,14 | 0,65 | 0,42 | 1373,30 |
| 2004 | 226,70 | 15,98 | 0,72 | 0,43 | 965,82 |
| 2005 | 255,20 | 14,90 | 1,77 | 0,40 | 382,22 |
| 2006 | 255,35 | 12,15 | 1,17 | 1,10 | 228,63 |
| 2007 | 293,40 | 10,95 | 0,96 | 0,25 | 240,72 |
| 2008 | 285,40 | 9,71 | 0,91 | 0,16 | 244,89 |
| 2009 | 328,05 | 11,21 | 0,65 | 0,12 | 171,70 |
| 2010 | 292,50 | 13,85 | 0,69 | 0,12 | 161,74 |
| 2011 | 336,75 | 62,63 | 0,73 | 0,12 | 175,67 |
| 2012 | 315,20 | 10,48 | 1,17 | 0,13 | 187,65 |
| 2013 | 340,45 | 10,66 | 1,00 | 0,14 | 178,72 |
| 2014 | 378,20 | 10,81 | 1,21 | 0,14 | 183,34 |
| 2015 | 344,15 | 11,65 | 1,05 | 0,15 | 188,24 |
| 2016 | 360,05 | 8,50 | 1,23 | 0,15 | 181,20 |
| 2017 | 390,20 | 8,93 | 2,69 | 0,17 | 166,13 |
| 2018 | 376,70 | 8,93 | 2,69 | 0,17 | 166,13 |

Table F5. Activity data used for Other Product use og Production (2G, 2H, 2I)

| | 2-Other product use | | 2H2-Food and beverages industry | | 2I-Wood processing |
|------|---------------------|-------------|---------------------------------|--------------------|--------------------|
| | fireworks (t) | tobacco (t) | Bread production (kt) | Beer (1000 liter) | Wood product (kt) |
| 1990 | 556,80 | 5664,00 | 340,00 | 227639,00 | 181,65 |
| 1991 | 570,23 | 6183,00 | 293,70 | 230412,00 | 181,65 |
| 1992 | 581,13 | 6183,00 | 297,67 | 227318,00 | 181,65 |
| 1993 | 618,31 | 5550,00 | 290,06 | 222876,00 | 185,96 |
| 1994 | 535,25 | 5608,00 | 290,06 | 222837,00 | 202,42 |
| 1995 | 935,61 | 3475,00 | 242,92 | 225523,00 | 200,06 |
| 1996 | 885,01 | 4473,00 | 282,52 | 232919,00 | 198,60 |
| 1997 | 1090,02 | 5463,00 | 272,90 | 239600,00 | 206,50 |
| 1998 | 1491,90 | 5828,00 | 297,84 | 183333,00 | 219,84 |
| 1999 | 2550,02 | 4189,00 | 289,31 | 265111,00 | 243,27 |
| 2000 | 2213,42 | 4533,87 | 266,45 | 215011,00 | 258,01 |
| 2001 | 1860,16 | 5122,87 | 207,84 | 246171,00 | 250,82 |
| 2002 | 2280,16 | 5092,55 | 292,25 | 237745,00 | 239,36 |
| 2003 | 2663,98 | 5160,85 | 302,46 | 217676,00 | 256,15 |
| 2004 | 2472,21 | 5692,44 | 324,84 | 235226,00 | 286,76 |
| 2005 | 2538,70 | 2847,13 | 370,24 | 244176,00 | 298,10 |
| 2006 | 2070,59 | 3835,14 | 353,90 | 241022,00 | 301,25 |
| 2007 | 3006,87 | 3926,11 | 335,11 | 175519,00 | 300,60 |
| 2008 | 2609,03 | 4008,42 | 282,92 | 261186,00 | 253,59 |
| 2009 | 1747,45 | 4059,30 | 309,78 | 267446,60 | 207,43 |
| 2010 | 2989,98 | 3816,31 | 474,30 | 240564,02 | 245,55 |
| 2011 | 2529,22 | 3313,65 | 531,27 | 268733,95 | 250,47 |
| 2012 | 2493,39 | 2905,06 | 361,54 | 230744,54 | 233,87 |
| 2013 | 2497,45 | 2711,34 | 457,11 | 237580,37 | 212,31 |
| 2014 | 2778,30 | 2898,72 | 266,56 | 239641,06 | 231,69 |
| 2015 | 2740,16 | 2870,34 | 268,08 | 236181,42 | 234,42 |
| 2016 | 2200,39 | 2634,98 | 301,17 | 253480,29 | 224,33 |
| 2017 | 2234,95 | 2594,96 | 288,32 | 248188,87 | 229,98 |
| 2018 | 2235,30 | 2323,50 | 301,17 | 138734,82 | 240,04 |

Table F6. Animal population data used in the estimations (3. Agriculture). Animal numbers. 1990, 1995, 2000, 2005, 2010-2018.

| | 1990 | 1995 | 2000 | 2005 | 2010 | 2011 |
|--|------------|------------|------------|------------|------------|------------|
| Mature dairy cattle | 325 896 | 310 346 | 284 880 | 255 663 | 232 294 | 224 721 |
| Beef cow (other mature cattle) | 8 193 | 20 334 | 42 324 | 54 841 | 67 110 | 68 539 |
| Replacement heifer | 143 904 | 138 359 | 129 500 | 118 090 | 111 122 | 110 311 |
| Heifers for slaughter | 29 011 | 27 708 | 38 710 | 32 843 | 29 966 | 28 281 |
| Bulls for slaughter | 185 719 | 179 929 | 198 396 | 175 579 | 160 568 | 158 953 |
| Sheep <1 year (adj. for lifetime) | 622 862 | 683 599 | 643 141 | 685 466 | 659 895 | 653 663 |
| Sheep >1 år | 714 384 | 783 922 | 766 098 | 717 098 | 691 450 | 687 440 |
| Piglets | 204 623 | 235 223 | 244 288 | 269 761 | 282 931 | 274 779 |
| Young pigs for breeding | 23 047 | 24 714 | 31 172 | 38 930 | 40 770 | 42 023 |
| Sows | 62 683 | 63 944 | 57 351 | 60 584 | 56 234 | 53 277 |
| Boars | 2 046 | 1 727 | 1 453 | 1 299 | 1 096 | 1 045 |
| Fattening pigs | 1 059 589 | 1 153 285 | 1 280 884 | 1 404 856 | 1 565 736 | 1 580 311 |
| Deer | 0 | 0 | 2 280 | 4 173 | 7 249 | 7 808 |
| Dairy goats | 64 041 | 58 630 | 50 578 | 44 374 | 35 706 | 34 783 |
| Other goats | 19 759 | 20 082 | 19 131 | 18 163 | 20 793 | 21 186 |
| Horses | 31 430 | 38 013 | 51 156 | 61 784 | 76 752 | 77 101 |
| Laying hens | 2 895 663 | 3 556 841 | 3 228 812 | 3 343 410 | 3 945 607 | 3 877 138 |
| Chickens reared for laying | 3 459 064 | 2 984 493 | 2 184 479 | 3 066 358 | 2 777 268 | 2 426 312 |
| Broilers | 15 864 401 | 23 318 120 | 35 757 612 | 43 612 212 | 61 245 745 | 61 505 467 |
| Turkeys for slaughter | 528 240 | 776 428 | 673 282 | 953 112 | 1 141 867 | 1 187 676 |
| Ducks and geese for slaughter | 18 551 | 27 267 | 81 365 | 69 368 | 153 831 | 188 855 |
| Turkeys, ducks and geese reared for laying | 15 506 | 29 930 | 20 292 | 45 378 | 36 901 | 14 889 |
| Reindeer | 242 443 | 212 333 | 172 407 | 234 608 | 254 384 | 251 071 |
| Mink | 56 411 | 44 199 | 68 526 | 98 247 | 107 980 | 119 741 |
| Foxes | 104 126 | 122 146 | 86 160 | 76 756 | 49 213 | 49 153 |

| | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
|--|------------|------------|------------|------------|------------|------------|------------|
| Mature dairy cattle | 229 767 | 225 163 | 221 032 | 217 576 | 215 015 | 217 318 | 211 523 |
| Beef cow (other mature cattle) | 70 434 | 71 834 | 73 894 | 77 408 | 84 372 | 88 332 | 92 304 |
| Replacement heifer | 108 455 | 109 643 | 109 115 | 110 495 | 112 174 | 115 473 | 114 152 |
| Heifers for slaughter | 24 899 | 27 260 | 37 538 | 34 933 | 34 482 | 24 171 | 30 193 |
| Bulls for slaughter | 156 087 | 160 419 | 153 566 | 148 861 | 149 754 | 169 625 | 181 684 |
| Sheep <1 year (adj. for lifetime) | 634 325 | 642 338 | 676 867 | 706 468 | 757 659 | 746 214 | 696 357 |
| Sheep >1 år | 685 326 | 689 345 | 683 479 | 716 252 | 729 014 | 730 666 | 669 711 |
| Piglets | 275 349 | 276 500 | 269 152 | 270 555 | 274 325 | 267 077 | 248 835 |
| Young pigs for breeding | 42 636 | 42 960 | 42 634 | 43 525 | 41 680 | 41 006 | 38 939 |
| Sows | 53 154 | 53 004 | 50 291 | 48 867 | 47 945 | 47 093 | 44 903 |
| Boars | 1 037 | 1 048 | 953 | 1 058 | 796 | 799 | 1 344 |
| Fattening pigs | 1 535 494 | 1 540 851 | 1 587 993 | 1 537 703 | 1 591 311 | 1 589 084 | 1 642 094 |
| Deer | 8 367 | 7 829 | 7 714 | 7 469 | 7 838 | 7 086 | 7 970 |
| Dairy goats | 33 982 | 31 406 | 31 461 | 33 627 | 34 660 | 34 126 | 34 583 |
| Other goats | 20 061 | 21 013 | 21 750 | 21 891 | 22 198 | 21 112 | 23 413 |
| Horses | 77 086 | 77 059 | 74 635 | 73 303 | 71 350 | 69 552 | 72 472 |
| Laying hens | 4 050 447 | 4 216 858 | 4 320 632 | 4 359 188 | 4 336 730 | 4 365 344 | 4 308 640 |
| Chickens reared for laying | 2 319 570 | 2 938 451 | 2 686 575 | 2 738 693 | 2 614 453 | 2 631 703 | 2 143 725 |
| Broilers | 63 806 388 | 71 899 359 | 73 974 651 | 63 406 519 | 65 898 097 | 63 516 948 | 62 738 774 |
| Turkeys for slaughter | 1 262 231 | 1 174 143 | 1 245 554 | 1 260 617 | 1 179 466 | 1 037 274 | 825 264 |
| Ducks and geese for slaughter | 90 309 | 237 100 | 302 757 | 298 089 | 291 989 | 278 423 | 274 298 |
| Turkeys, ducks and geese reared for laying | 13 557 | 13 257 | 20 662 | 23 811 | 19 530 | 20 601 | 12 336 |
| Reindeer | 258 360 | 248 225 | 232 905 | 211 974 | 211 666 | 213 913 | 213 012 |
| Mink | 141 368 | 182 334 | 174 613 | 161 394 | 143 156 | 107 039 | 136 993 |
| Foxes | 51 079 | 51 916 | 49 143 | 40 734 | 31 828 | 21 124 | 27 554 |

Table F7. Animal population data used in the estimations (3. Agriculture). Animal places. 1990, 1995, 2000, 2005, 2010-2018.

| | 1990 | 1995 | 2000 | 2005 | 2010 | 2011 |
|--|-----------|-----------|-----------|-----------|-----------|-----------|
| Mature dairy cattle | 325 896 | 310 346 | 284 880 | 255 663 | 232 294 | 224 721 |
| Beef cow (other mature cattle) | 8 193 | 20 334 | 42 324 | 54 841 | 67 110 | 68 539 |
| Replacement heifer | 311 279 | 299 284 | 280 121 | 255 862 | 239 839 | 239 007 |
| Heifers for slaughter | 47 020 | 47 103 | 63 512 | 57 619 | 53 410 | 48 778 |
| Bulls for slaughter | 289 945 | 284 237 | 285 349 | 263 170 | 230 872 | 223 536 |
| Sheep <1 year (adj. for lifetime) | 622 862 | 683 599 | 643 141 | 685 466 | 659 895 | 653 663 |
| Sheep >1 år | 714 384 | 783 922 | 766 098 | 717 098 | 691 450 | 687 440 |
| Piglets | 204 623 | 235 223 | 244 288 | 269 761 | 282 931 | 274 779 |
| Young pigs for breeding | 23 047 | 24 714 | 31 172 | 38 930 | 40 770 | 42 023 |
| Sows | 62 683 | 63 944 | 57 351 | 60 584 | 56 234 | 53 277 |
| Boars | 2 046 | 1 727 | 1 453 | 1 299 | 1 096 | 1 045 |
| Fattening pigs | 423 836 | 428 654 | 444 604 | 457 395 | 480 007 | 478 882 |
| Deer | 0 | 0 | 2 280 | 4 173 | 7 249 | 7 808 |
| Dairy goats | 64 041 | 58 630 | 50 578 | 44 374 | 35 706 | 34 783 |
| Other goats | 19 759 | 20 082 | 19 131 | 18 163 | 20 793 | 21 186 |
| Horses | 31 430 | 38 013 | 51 156 | 61 784 | 76 752 | 77 101 |
| Laying hens | 2 895 663 | 3 556 841 | 3 228 812 | 3 343 410 | 3 945 607 | 3 877 138 |
| Chickens reared for laying | 1 729 532 | 1 424 417 | 997 262 | 1 341 532 | 1 166 453 | 1 010 963 |
| Broilers | 3 172 880 | 4 352 716 | 6 257 582 | 7 183 188 | 9 527 116 | 9 462 380 |
| Turkeys for slaughter | 176 080 | 269 504 | 243 775 | 360 637 | 452 438 | 475 070 |
| Ducks and geese for slaughter | 4 638 | 6 434 | 18 177 | 14 714 | 31 062 | 37 771 |
| Turkeys, ducks and geese reared for laying | 15 506 | 29 930 | 20 292 | 45 378 | 36 901 | 14 889 |
| Reindeer | 242 443 | 212 333 | 172 407 | 234 608 | 254 384 | 251 071 |
| Mink | 56 411 | 44 199 | 68 526 | 98 247 | 107 980 | 119 741 |
| Foxes | 104 126 | 122 146 | 86 160 | 76 756 | 49 213 | 49 153 |

| | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
|--|-----------|------------|------------|-----------|------------|-----------|-----------|
| Mature dairy cattle | 229 767 | 225 163 | 221 032 | 217 576 | 215 015 | 217 318 | 211 523 |
| Beef cow (other mature cattle) | 70 434 | 71 834 | 73 894 | 77 408 | 84 372 | 88 332 | 92 304 |
| Replacement heifer | 235 891 | 239 386 | 244 601 | 238 485 | 241 173 | 249 229 | 245 428 |
| Heifers for slaughter | 42 863 | 47 294 | 67 624 | 64 814 | 64 361 | 43 501 | 52 356 |
| Bulls for slaughter | 217 044 | 220 440 | 208 979 | 206 328 | 217 885 | 250 630 | 260 130 |
| Sheep <1 year (adj. for lifetime) | 634 325 | 642 338 | 676 867 | 706 468 | 757 659 | 746 214 | 696 357 |
| Sheep >1 år | 685 326 | 689 345 | 683 479 | 716 252 | 729 014 | 730 666 | 669 711 |
| Piglets | 275 349 | 276 500 | 269 152 | 270 555 | 274 325 | 267 077 | 248 835 |
| Young pigs for breeding | 42 636 | 42 960 | 42 634 | 43 525 | 41 680 | 41 006 | 38 939 |
| Sows | 53 154 | 53 004 | 50 291 | 48 867 | 47 945 | 47 093 | 44 903 |
| Boars | 1 037 | 1 048 | 953 | 1 058 | 796 | 799 | 1 344 |
| Fattening pigs | 465 301 | 466 925 | 481 210 | 465 971 | 482 215 | 481 541 | 497 604 |
| Deer | 8 367 | 7 829 | 7 714 | 7 469 | 7 838 | 7 086 | 7 970 |
| Dairy goats | 33 982 | 31 406 | 31 461 | 33 627 | 34 660 | 34 126 | 34 583 |
| Other goats | 20 061 | 21 013 | 21 750 | 21 891 | 22 198 | 21 112 | 23 413 |
| Horses | 77 086 | 77 059 | 74 635 | 73 303 | 71 350 | 69 552 | 72 472 |
| Laying hens | 4 050 447 | 4 216 858 | 4 320 632 | 4 359 188 | 4 336 730 | 4 365 344 | 4 308 640 |
| Chickens reared for laying | 966 488 | 1 224 355 | 1 119 406 | 1 141 122 | 1 089 355 | 1 096 543 | 893 219 |
| Broilers | 9 816 367 | 11 061 440 | 11 380 716 | 9 754 849 | 10 138 169 | 9 771 838 | 9 652 119 |
| Turkeys for slaughter | 504 892 | 469 657 | 498 222 | 504 247 | 471 786 | 414 910 | 330 106 |
| Ducks and geese for slaughter | 18 062 | 47 420 | 60 551 | 59 618 | 58 398 | 55 685 | 54 860 |
| Turkeys, ducks and geese reared for laying | 13 557 | 13 257 | 20 662 | 23 811 | 19 530 | 20 601 | 12 336 |
| Reindeer | 258 360 | 248 225 | 232 905 | 211 974 | 211 666 | 213 913 | 213 012 |
| Mink | 141 368 | 182 334 | 174 613 | 161 394 | 143 156 | 107 039 | 136 993 |
| Foxes | 51 079 | 51 916 | 49 143 | 40 734 | 31 828 | 21 124 | 27 554 |

Table F8. Total value for annual consumption of synthetic fertilizers in Norway based on sale figures (Norwegian Food Safety Authority Annually).

| | Total amount of fertilizer sold | Total amount for agriculture purposes |
|------|------------------------------------|---|
| | tonn N | tonn N |
| 1990 | 110418 | 110182 |
| 1991 | 110790 | 110398 |
| 1992 | 110875 | 110519 |
| 1993 | 109299 | 109051 |
| 1994 | 108287 | 108051 |
| 1995 | 110851 | 110631 |
| 1996 | 111976 | 111741 |
| 1997 | 112879 | 112627 |
| 1998 | 112327 | 112062 |
| 1999 | 106017 | 105756 |
| 2000 | 107410 | 107254 |
| 2001 | 100592 | 100419 |
| 2002 | 101258 | 101071 |
| 2003 | 104162 | 104075 |
| 2004 | 105096 | 105018 |
| 2005 | 106882 | 106798 |
| 2006 | 104088 | 104050 |
| 2007 | 107588 | 107506 |
| 2008 | 103027 | 102920 |
| 2009 | 93823 | 93709 |
| 2010 | 85131 | 85017 |
| 2011 | 96851 | 96778 |
| 2012 | 95767 | 95682 |
| 2013 | 97010 | 97010 |
| 2014 | 102238 | 102238 |
| 2015 | 104214 | 104214 |
| 2016 | 102460 | 102460 |
| 2017 | 99674 | 99674 |
| 2018 | 102392 | 102392 |

Table F9. Total value for annual consumption of synthetic fertilizers in Norway based on sale figures (Norwegian Food Safety Authority Annually).

| Fertilizer type | Amount of fertilizer (tonnes) | %-N | Loss (g NH ₃ /kg N applied) |
|---------------------------------|-------------------------------|------|--|
| Ammonium nitrate | 4 637 | 34.3 | 15 |
| Ammonium nitrate m/S | 75 964 | 26.9 | 15 |
| Ammonium sulphate nitrate | 1 014 | 30.0 | 0 |
| Potassium sulphate | 117 | 0.0 | 0 |
| Potassium sulphate m/Mg | 741 | 0.0 | 0 |
| Potassium chloride | 95 | 0.0 | 0 |
| Kalkamonsalpeter | 4 604 | 27.0 | 8 |
| Calcium nitrate | 8 345 | 15.5 | 10 |
| Calcium nitrate m/B | 3 798 | 15.4 | 10 |
| NK-fertilizer 22-12 | 5 800 | 22.0 | 15 |
| NP fertilizer 12-23 | 909 | 12.2 | 50 |
| NPK-fertilizer 8-5-19 | 939 | 7.9 | 50 |
| NPK-fertilizer 12-4-18 | 19 693 | 11.8 | 50 |
| NPK-fertilizer 15-7-12 | 1 434 | 15.0 | 50 |
| NPK-fertilizer 18-3-15 | 32 605 | 17.6 | 50 |
| NPK-fertilizer 20-4-11 | 22 911 | 19.6 | 50 |
| NPK-fertilizer 21-3-10 | 1 173 | 21.0 | 50 |
| NPK-fertilizer 21-6-6 | 3 110 | 20.6 | 50 |
| NPK-fertilizer 22-2-12 | 32 520 | 21.6 | 50 |
| NPK-fertilizer 22-3-10 | 106 013 | 21.6 | 50 |
| NPK-fertilizer 24-4-6 | 15 271 | 24.0 | 50 |
| NPK-fertilizer 25-2-6 | 105 432 | 24.6 | 50 |
| NPK-fertilizer 27-3-5 | 7 398 | 26.9 | 50 |
| PK-fertilizer 0-11-21 | 424 | 0.0 | 0 |
| P-fertilizer 0-20-0 | 245 | 0.0 | 0 |
| Urea | 141 | 46.0 | 155 |
| Other fertilizer with N content | 1579 | 16.0 | 1 |
| Other fertilizer | 677.25 | 0.5 | 1 |

Table F10. N amounts of the different fertilizers, tonnes N.

| Tonnes N | Inorganic fertilizer | Manure applied to fields | Sewage sludge | Other organic fertilizers | Crop residues |
|----------|----------------------|--------------------------|---------------|---------------------------|---------------|
| | 3Da1 | 3Da2a | 3Da2b | 3Da3 | 3Da4 |
| 1990 | 110 182 | 49 603 | 738 | 0 | 22 351 |
| 1991 | 110 398 | 50 254 | 938 | 0 | 21 441 |
| 1992 | 110 519 | 50 755 | 1 068 | 0 | 18 975 |
| 1993 | 109 051 | 48 180 | 1 125 | 0 | 22 565 |
| 1994 | 108 051 | 48 408 | 1 069 | 0 | 20 936 |
| 1995 | 110 631 | 48 317 | 1 176 | 0 | 20 328 |
| 1996 | 111 741 | 49 336 | 1 190 | 7 | 21 126 |
| 1997 | 112 627 | 48 373 | 1 191 | 7 | 20 073 |
| 1998 | 112 062 | 48 932 | 1 207 | 7 | 20 073 |
| 1999 | 105 756 | 50 215 | 1 345 | 7 | 18 418 |
| 2000 | 107 254 | 48 562 | 1 334 | 172 | 18 665 |
| 2001 | 100 419 | 48 992 | 1 193 | 175 | 18 044 |
| 2002 | 101 071 | 48 932 | 1 228 | 251 | 15 234 |
| 2003 | 104 075 | 50 815 | 1 393 | 1467 | 14 687 |
| 2004 | 105 018 | 50 892 | 1 401 | 1066 | 15 435 |
| 2005 | 106 798 | 51 628 | 1 484 | 285 | 13 519 |
| 2006 | 104 050 | 51 640 | 1 433 | 293 | 12 667 |
| 2007 | 107 506 | 52 222 | 1 818 | 363 | 12 443 |
| 2008 | 102 920 | 53 019 | 1 922 | 501 | 13 262 |
| 2009 | 93 709 | 54 008 | 1 849 | 1170 | 12 042 |
| 2010 | 85 017 | 54 485 | 1 614 | 1174 | 12 300 |
| 2011 | 96 778 | 53 469 | 1 801 | 713 | 10 967 |
| 2012 | 95 682 | 54 310 | 1 977 | 758 | 12 454 |
| 2013 | 97 010 | 55 594 | 2 330 | 611 | 12 904 |
| 2014 | 102 238 | 55 643 | 2 208 | 611 | 15 352 |
| 2015 | 104 214 | 55 743 | 2 000 | 611 | 16 384 |
| 2016 | 102 460 | 56 443 | 1 854 | 611 | 15 786 |
| 2017 | 99 674 | 56 873 | 1 861 | 611 | 15 369 |
| 2018 | 102 392 | 56 658 | 1 844 | 611 | 9 652 |

Table F11. N amounts of the different fertilizers, tonnes N.

| | Cultivated cropland, hectares | Cultivated cropland and grassland, hectares |
|------|----------------------------------|--|
| | 3Dc | 3De |
| 1990 | 4 388 644 | 8 822 908 |
| 1991 | 4 344 862 | 8 828 277 |
| 1992 | 4 301 081 | 8 833 647 |
| 1993 | 4 257 299 | 8 839 017 |
| 1994 | 4 213 518 | 8 844 387 |
| 1995 | 4 169 737 | 8 849 756 |
| 1996 | 4 125 955 | 8 855 126 |
| 1997 | 4 082 174 | 8 860 496 |
| 1998 | 4 038 392 | 8 865 865 |
| 1999 | 3 994 611 | 8 871 235 |
| 2000 | 3 984 476 | 8 840 505 |
| 2001 | 3 996 958 | 8 861 809 |
| 2002 | 3 914 499 | 8 831 474 |
| 2003 | 3 854 297 | 8 759 425 |
| 2004 | 3 845 468 | 8 736 957 |
| 2005 | 3 804 475 | 8 662 462 |
| 2006 | 3 731 323 | 8 627 623 |
| 2007 | 3 668 837 | 8 574 377 |
| 2008 | 3 609 245 | 8 493 510 |
| 2009 | 3 566 359 | 8 394 810 |
| 2010 | 3 535 985 | 8 301 464 |
| 2011 | 3 476 160 | 8 223 637 |
| 2012 | 3 434 259 | 8 163 288 |
| 2013 | 3 350 490 | 8 107 515 |
| 2014 | 3 325 975 | 8 111 345 |
| 2015 | 3 302 451 | 8 101 627 |
| 2016 | 3 331 652 | 8 081 580 |
| 2017 | 3 318 295 | 8 045 062 |
| 2018 | 3 267 309 | 8 053 735 |

Table F12. Activity data used for Waste incineration (5C)

| | 5C1biii-Clinical waste incineration waste incinerated (kt) | 5C1bv-Cremation Corpes incineration (number) |
|------|---|---|
| 1990 | 0,63 | 13916 |
| 1991 | 0,63 | 13646 |
| 1992 | 0,63 | 13756 |
| 1993 | 0,54 | 14067 |
| 1994 | 0,59 | 14106 |
| 1995 | 0,48 | 13669 |
| 1996 | 0,44 | 13876 |
| 1997 | 0,47 | 13884 |
| 1998 | 0,49 | 13620 |
| 1999 | 0,41 | 14182 |
| 2000 | 0,24 | 14039 |
| 2001 | 0,24 | 13979 |
| 2002 | 0,14 | 14154 |
| 2003 | 0,14 | 13418 |
| 2004 | 0,14 | 13494 |
| 2005 | 0,14 | 13894 |
| 2006 | IE | 14107 |
| 2007 | IE | 14390 |
| 2008 | IE | 14451 |
| 2009 | IE | 14815 |
| 2010 | IE | 14664 |
| 2011 | IE | 15176 |
| 2012 | IE | 15455 |
| 2013 | IE | 15725 |
| 2014 | IE | 15641 |
| 2015 | IE | 16164 |
| 2016 | IE | 16703 |
| 2017 | IE | 16903 |
| 2018 | IE | 17429 |

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